

Electronic Supplementary Information (ESI)

for

**Selective H₂O₂ conversion to hydroxyl radicals in electron-rich area of hydroxylated C-
g-C₃N₄/CuCo-Al₂O₃**

Lai Lyu^{ab}, Lili Zhang^a, Guangzhi He^c, Hong He^{bc} and Chun Hu^{,ab}*

^aKey Laboratory of Drinking Water Science and Technology, Research Center for Eco-
Environmental Sciences, Chinese Academy of Sciences, Beijing 100085, China.

^bUniversity of Chinese Academy of Sciences, Beijing 100049, China.

^cState Key Joint Laboratory of Environment Simulation and Pollution Control, Research
Center for Eco-Environmental Sciences, Chinese Academy of Sciences, Beijing 100085,
China

*Tel: +86-10-62849628; fax: +86-10-62923541; e-mail: huchun@rcees.ac.cn

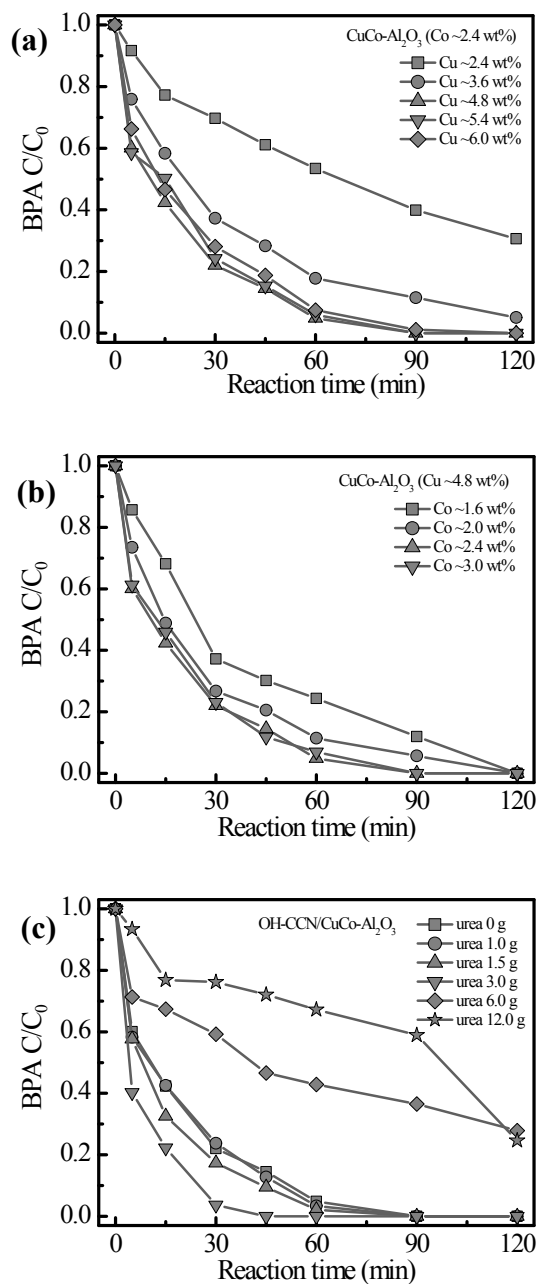


Figure S1. Effect of the amount of the introduced (a) Cu and (b) Co in CuCo-Al₂O₃ and (c) the used urea during the preparation of OH-CCN/CuCo-Al₂O₃ on BPA degradation with H₂O₂ (Initial pH 7, initial BPA concentration 25 mg L⁻¹, initial H₂O₂ concentration 10 mM, catalyst 0.8 g L⁻¹).

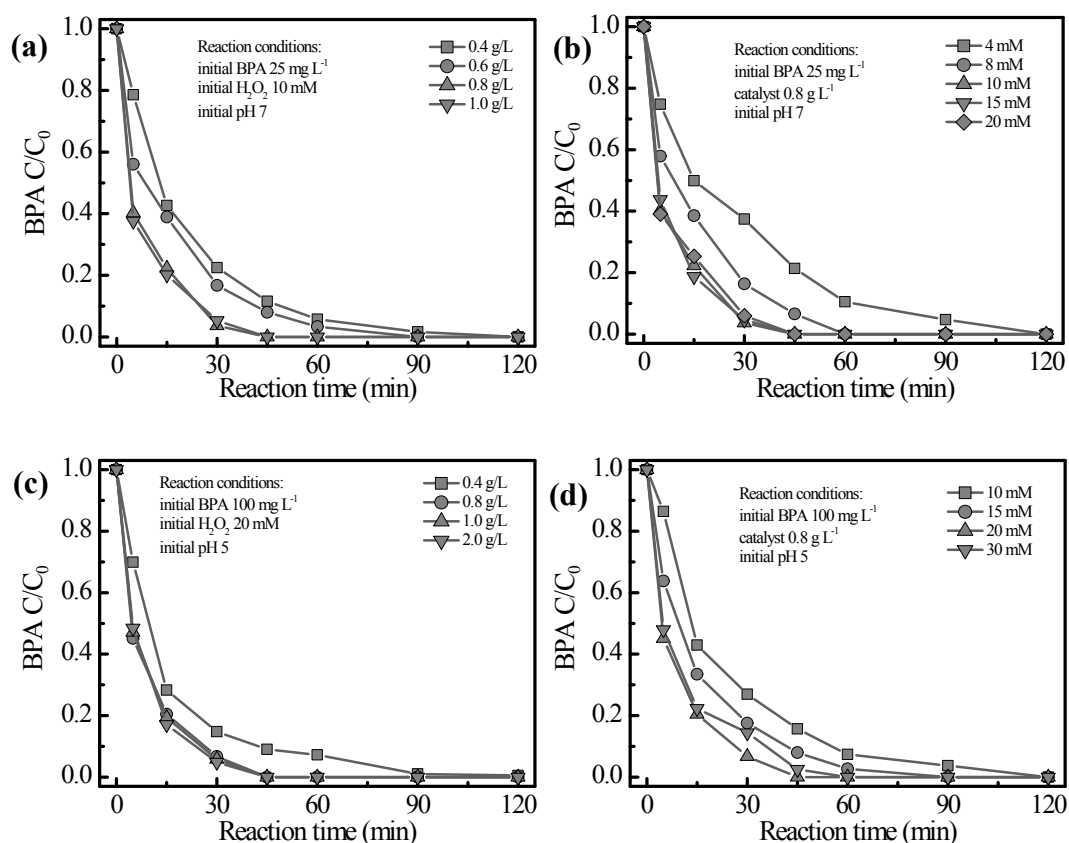


Figure S2. Effect of (a) catalyst concentration and (b) H₂O₂ dosage on 25 mg L⁻¹ BPA degradation in the OH-CCN/CuCo-Al₂O₃ suspension at pH 7; and effect of (c) catalyst concentration and (d) H₂O₂ dosage on 100 mg L⁻¹ BPA degradation in the OH-CCN/CuCo-Al₂O₃ suspension at pH 5.

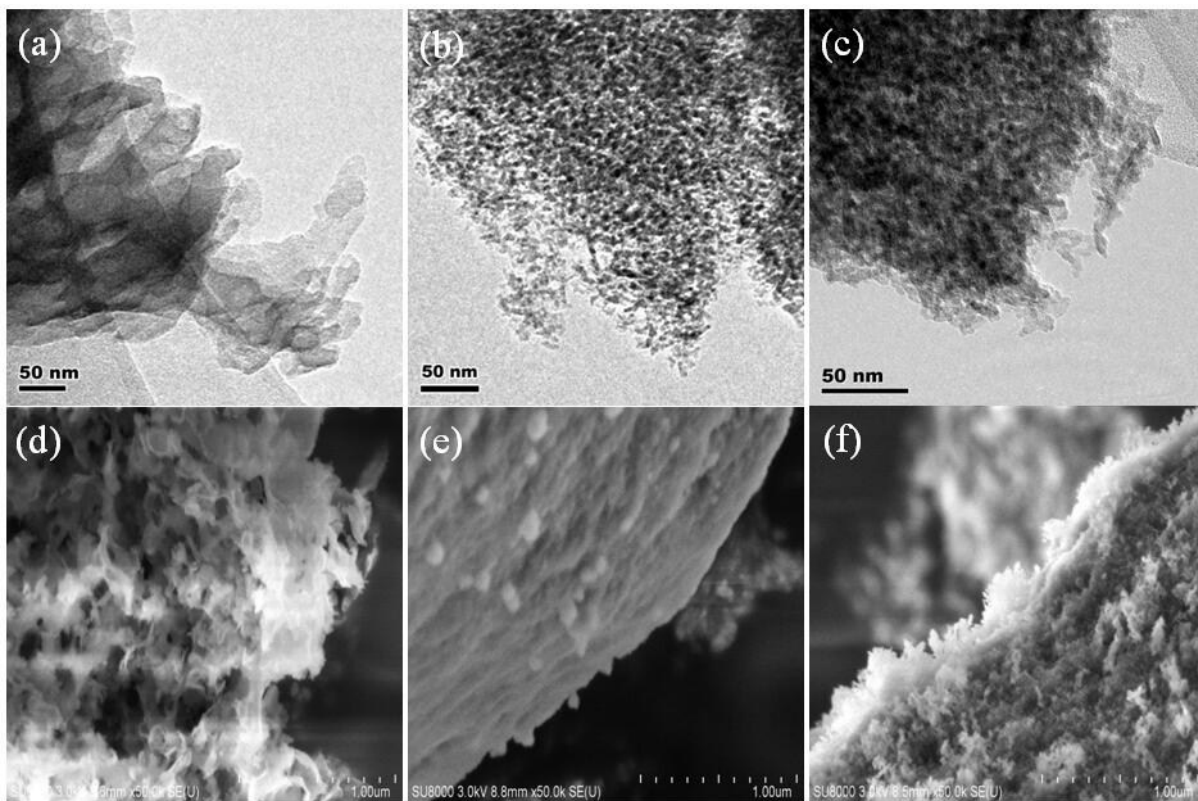


Figure S3. TEM of (a) g-C₃N₄, (b) CuCo-Al₂O₃ and (c) OH-CCN/CuCo-Al₂O₃. SEM of (d) g-C₃N₄, (e) CuCo-Al₂O₃ and (f) OH-CCN/CuCo-Al₂O₃.

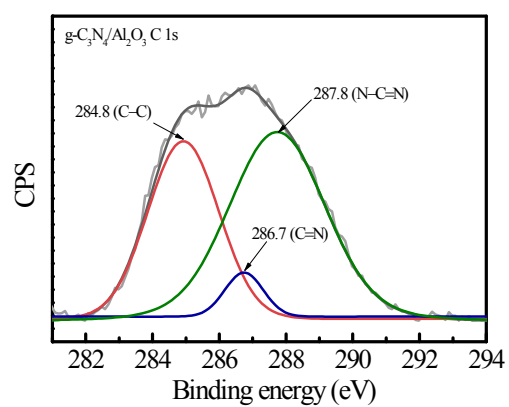


Figure S4. XPS spectra in C 1s for g-C₃N₄/Al₂O₃.

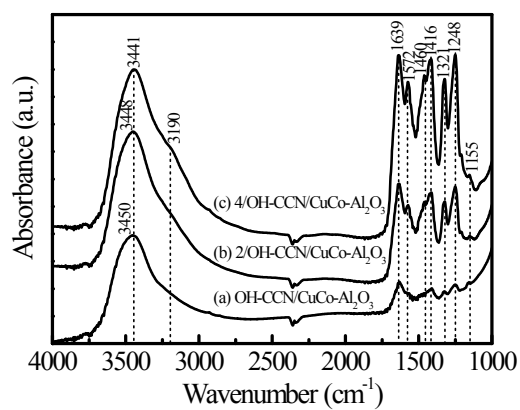


Figure S5. FTIR spectra for OH-CCN/CuCo-Al₂O₃ with different amount of g-C₃N₄: (a) normal OH-CCN/CuCo-Al₂O₃, (b) OH-CCN/CuCo-Al₂O₃ with 2 times of g-C₃N₄ (2/OH-CCN/CuCo-Al₂O₃) and (c) OH-CCN/CuCo-Al₂O₃ with 4 times of g-C₃N₄ (4/OH-CCN/CuCo-Al₂O₃).

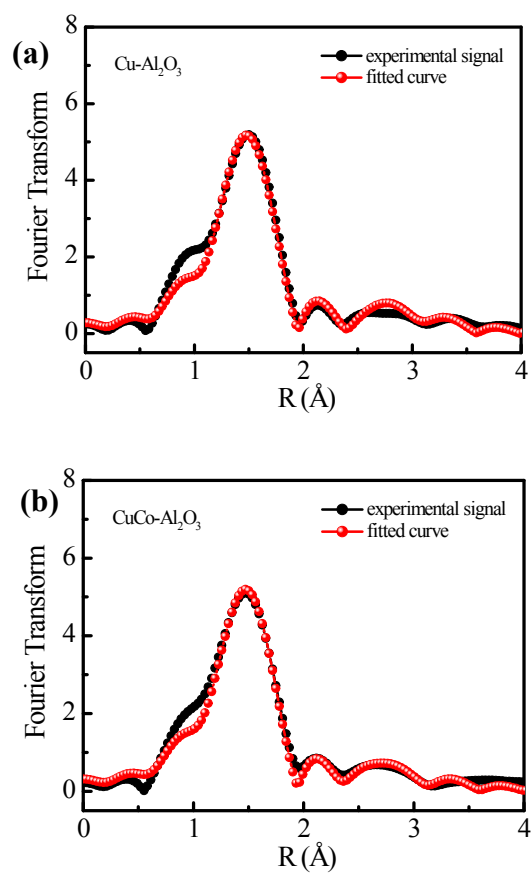


Figure S6. The EXAFS curve fitting for Cu K-edge of (a) $\text{Cu-Al}_2\text{O}_3$ and (b) $\text{CuCo-Al}_2\text{O}_3$. (solid line—experimental signals, dotted line—fitted curves).

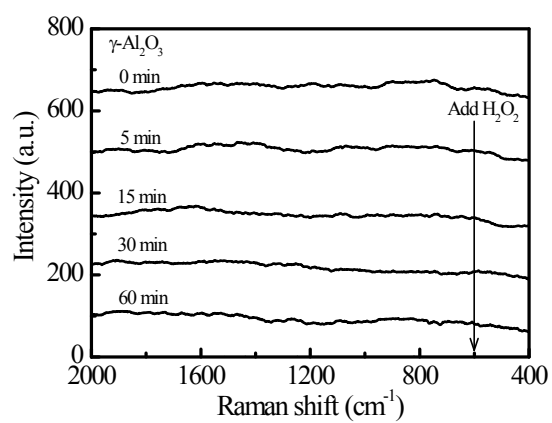


Figure S7. *In situ* Raman spectra of $\gamma\text{-Al}_2\text{O}_3$ aqueous dispersion recorded at different time after adding H_2O_2 .

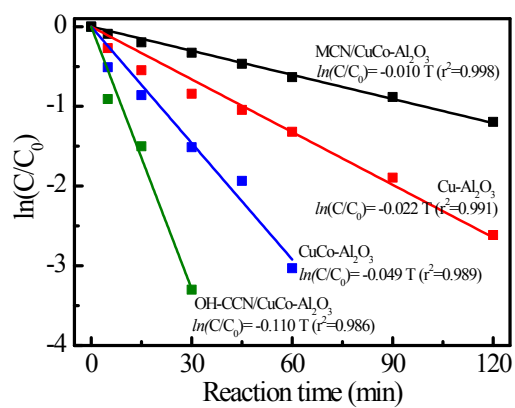


Figure S8. Kinetic curves of BPA degradation in various suspensions (Initial pH 7, initial BPA 0.1 mM, initial H₂O₂ 10 mM, catalyst 0.8 g L⁻¹).

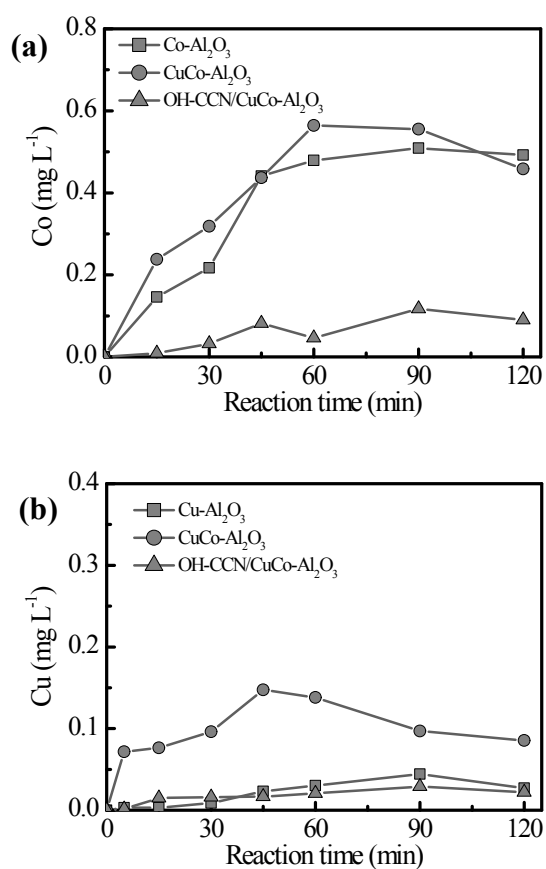


Figure S9. Release of (a) Co and (b) Cu in Co-Al₂O₃, Cu-Al₂O₃, CuCo-Al₂O₃ and OH-CCN/CuCo-Al₂O₃ suspensions during the reaction with H₂O₂. Zero point (0 min) has been calibrated by deducting the metal ion release from the catalyst before adding H₂O₂.

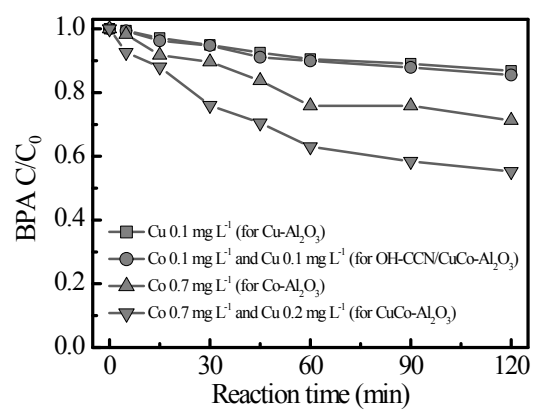


Figure S10. Homogeneous Fenton degradation of BPA in various suspensions (Initial BPA 0.1 mM, initial H₂O₂ 10 mM).

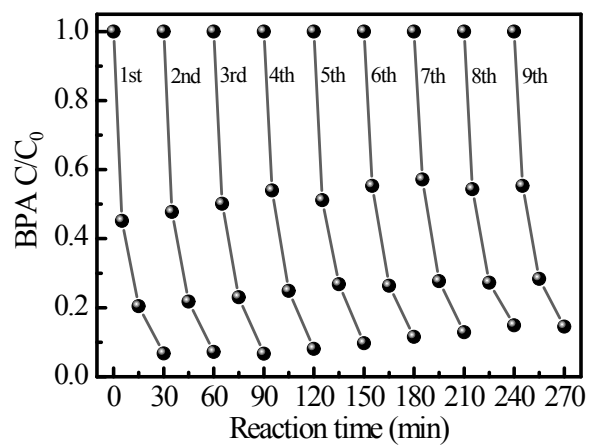


Figure S11. Cycling runs for the degradation of BPA in the OH-CCN/CuCo-Al₂O₃ suspension (Initial pH 7, initial BPA 0.1 mM, initial H₂O₂ 10 mM, catalyst 0.8 g L⁻¹).

Table S1. BET surface area, pore diameter, pore volume and metal content information of the various samples.

Samples	BET surface area (m ² g ⁻¹)	Pore diameter (nm)	Pore volume (cm ³ g ⁻¹)	Cu content (wt%)	Co content (wt%)
γ -Al ₂ O ₃	270.7	5.5	0.48	0	0
Cu-Al ₂ O ₃	222.0	6.3	0.45	4.79	0
CuCo-Al ₂ O ₃	186.0	8.5	0.48	4.82	2.45
OH-CCN/CuCo-Al ₂ O ₃	188.8	8.2	0.47	4.81	2.43

Table S2. EPR-measured g and A values for Cu(II) species in various samples.

Samples	$g_{//}$	g_{\perp}	$A_{//}$ (G)
Cu-Al ₂ O ₃	2.3862	2.1333	150
CuCo-Al ₂ O ₃	2.3825	2.1163	150
OH-CCN/CuCo-Al ₂ O ₃	2.3751	2.0995	150
2/OH-CCN/CuCo-Al ₂ O ₃	2.3665	2.0990	150
4/OH-CCN/CuCo-Al ₂ O ₃	2.3630	2.0985	150

Table S3. Actual H₂O₂ consumption ([ΔH₂O₂]_A) and stoichiometric H₂O₂ consumption ([ΔH₂O₂]_S) for mineralizing the pollutants during the reaction.

(a) [ΔH₂O₂]_A, [ΔH₂O₂]_S and η[H₂O₂] for mineralizing equivalent BPA in different catalyst suspensions.

entry	catalyst ^a	time (min)	TOC removal (mg L ⁻¹)	[ΔH ₂ O ₂] _A (mM)	[ΔH ₂ O ₂] _S (mM)	η[H ₂ O ₂] (%)
1	Cu-Al ₂ O ₃	120	72.1%	3.51	2.86	81.5
2	CuCo-Al ₂ O ₃	120	75.1%	3.77	2.97	78.8
3	OH-CCN/CuCo-Al ₂ O ₃	60	73.5%	3.23	2.91	90.1

^aReaction conditions: initial [BPA] = 25 mg L⁻¹, initial [H₂O₂] = 0.01 mol L⁻¹, [catalyst] = 0.8 g L⁻¹, initial pH=7, 308 K.

(b) $[\Delta\text{H}_2\text{O}_2]_{\text{A}}$ and $[\Delta\text{H}_2\text{O}_2]_{\text{S}}$ for mineralizing various pollutants in the OH-CCN/CuCo-Al₂O₃ system.

entry	contaminants ^a	initial conc. (mg L ⁻¹)	time (min)	$[\Delta\text{H}_2\text{O}_2]_{\text{A}}$ (mM)	$[\Delta\text{H}_2\text{O}_2]_{\text{S}}$ (mM)
1	benzene	100	30	9.07	8.57
2	phenol	100	30	12.61	11.72
3	benzoic acid	100	30	3.93	3.63
4	phloroglucinol	100	30	6.04	5.59
5	bisphenol A ^b	25	30	2.91	2.66
6	bisphenol A	100	30	12.12	11.32
7	2-chlorophenol	50	30	2.99	2.76
8	4-isopropylphenol	100	30	8.58	7.89
9	ibuprofen	30	30	2.93	2.67
10	phenytoin	30	30	2.65	2.42
11	diphenhydramine	30	30	3.05	2.76
12	2,4-dichlorophenoxyacetic acid	100	30	3.74	3.45
13	methyl orange	50	30	4.69	4.24
14	rhodamine B	50	30	4.52	4.2
15	methylene blue	80	30	5.44	4.72
16	atrazine	15	30	1.3	1.04

^aReaction conditions: initial $[\text{H}_2\text{O}_2] = 0.02 \text{ mol L}^{-1}$, $[\text{catalyst}] = 0.8 \text{ g L}^{-1}$, initial pH=5, 308 K.

^binitial $[\text{H}_2\text{O}_2] = 0.01 \text{ mol L}^{-1}$, $[\text{catalyst}] = 0.8 \text{ g L}^{-1}$, initial pH=7, 308 K.

Table S4. Comparison of the degradation of BPA by heterogeneous Fenton methods.

Entry	Catalyst	BPA Conc. (mM)	H ₂ O ₂ dose (mM)	Conditions	Reaction time (min)	Removal efficiency (%)	Utilization efficiency of H ₂ O ₂ (%)	Ref.
1	Fe ₃ O ₄ NPs	0.1	20	pH 5	120	19	5.2	1
2	Cu ₂ O MPs	0.1	20	pH 5	120	60	19.6	1
3	CuFeO ₂ MPs	0.1	20	pH 5	120	99.2	57.8	1
4	Fe ₃ O ₄ MNPs	~0.1	160	pH 3 ultrasonic processing	480	100	—	2
5	BiFeO ₃	0.1	10	pH 5 303 K	120	20.4	—	3
6	OH-CCN/ CuCo- Al ₂ O ₃	~0.1	10	ambient conditions (~pH 7)	30	96.3	90.5	This work
7	Au/SRAC	~0.5	~16	pH 3 313 K	720	89	—	4
8	Au/Fe ₂ O ₃	~0.5	~16	pH 3 313 K	720	10.1	—	4
9	Au- Fe ₂ O ₃ /Al ₂ O ₃	~0.5	~16	pH 3 313 K	720	6.6	—	4
10	OH-CCN/ CuCo- Al ₂ O ₃	~0.5	20	pH 5 308 K	30	93.3	93.4	This work

Supplementary Methods

Characterization of Catalyst

N₂ adsorption-desorption: The Brunner–Emmet–Teller (BET) specific surface area, pore diameter and pore volume were obtained on a gas sorption analyzer (ASAP 2020 HD88) at the temperature of liquid nitrogen. The samples were previously outgassed for 30 min at 150 °C and 4 Pa and then heated for 120 min at 350 °C.

XRD: Powder X-ray diffraction (XRD) patterns were recorded on a Scintag-XDS-2000 diffractometer with Cu K α radiation ($\lambda = 1.540598 \text{ \AA}$) operating at 40 kV and 40 mA. The wide diffraction angle analysis was conducted in a range of 5 to 90° with a scanning speed of 5°/min.

EXAFS: Extended X-ray absorption fine structure (EXAFS) spectra were recorded at the beam lines BL14W1 of the Shanghai Synchrotron Radiation Facility (SSRF), China. Cu foil, Cu₂O and CuO were used as references. The samples were sealed between two layers of adhesive PVC tape. The Cu K-edge (8.979 keV) EXAFS spectra of the samples were collected under ambient conditions in transmission mode. The parameters for EXAFS measurements, data collection modes and error calculations were all controlled according to guidelines set by the International XAFS Society Standards and Criteria Committee. The EXAFS data were analyzed by the Athena program. EXAFS oscillations $\chi(k)$ were extracted using spline smoothing and weighted by k^3 to compensate for the diminishing amplitude in the high k range. The filtered k^3 weighted $\chi(k)$ were Fourier transformed (FT) to R space in the k range of 2 to 11 \AA^{-1} . A nonlinear least-squares algorithm was applied to the EXAFS fitting with phase correlation in the R space between 1 and 4 \AA for the Cu K-edge. The structural parameters of the samples were obtained via curve fitting procedures using the FEFF8.4 code.⁵

XPS: X-ray photoelectron spectroscopy (XPS) data were recorded on an AXIS-Ultra instrument using monochromatic Al K α radiation (225 W, 15 mA, 15 kV) and low-energy

electron flooding for charge compensation. To compensate for surface charge effects, the binding energies were calibrated using the C1s hydrocarbon peak at 284.80 eV.

FTIR: Fourier-transform infrared spectroscopy (FTIR) was recorded on a Nicolet 8700 FTIR spectrophotometer (Thermo Fisher Scientific Inc., USA) in the range of 4000-400 cm^{-1} . The dry samples were supported on KBr pellets at a fixed proportion (1 wt%). All spectra were measured with a resolution of 4 cm^{-1} and with an accumulation of 32 scans.

Procedures, Analysis and Calculation

Detection of $\bullet\text{OH}$ and $\text{O}_2^{\bullet-}$ EPR signals: BMPO-trapped EPR signals were detected in different air-saturated methanol/aqueous dispersions of the corresponding samples. Typically, in the absence of H_2O_2 , 0.05 g of the prepared powder sample was added to 1 mL of water (for detecting $\bullet\text{OH}$) or water/methanol (10%/90%, V/V, for detecting $\text{O}_2^{\bullet-}$). 20 μL of BMPO (250 mM) was added and held for 5 min. Then, the solution was sucked into the capillary. The EPR spectra were recorded on a Bruker A300-10/12 EPR spectrometer at room temperature. In the presence of H_2O_2 , 0.01 g of the prepared powder sample was added to 1 mL of water (for detecting $\bullet\text{OH}$) or methanol (for detecting $\text{O}_2^{\bullet-}$). Then, 100 μL of the above suspension, 10 μL of BMPO (250 mM) and 10 μL of H_2O_2 (30%, w/w) were mixed and held for 5 min. The solution was sucked into the capillary to carry out EPR detection.

Measurement of *in situ* Raman spectra: *In situ* Raman spectra of the samples were collected on an inVia-Reflex confocal microscopic Raman spectrometer (Reinishaw, UK) with 40 mW 532 nm laser light irradiation. In a typical procedure, 0.1 g of the prepared powder sample were mixed with 3 mL of water, and the mixture was placed into the reaction cell, which was scanned from 400 to 2000 cm^{-1} at a resolution of 1 cm^{-1} for 120 s on the Raman spectrometer. Then, 50 μL of H_2O_2 (30%, w/w) was added followed by immediate stirring. At given time intervals, the scanning procedure described above was repeated.

Computational methods of density functional theory (DFT): The optimization of geometry and wave function was performed using B3LYP functional⁶ and 6-31G(d) basis set with the Gaussian 09 package.⁷ The valence-electron density was analyzed with the Multiwfn package.⁸ The geometry was visualized using VMD software.⁹ The carbon-doped g-C₃N₄ was modeled using a C₉N₁₂H₆ planar cluster with one N atom replaced by one C atom. The carbon-doped g-C₃N₄ was complexed with the Cu atom doped in γ -Al₂O₃ by the surface oxygen functional group. The dangling bonds of the O atoms coordinated to Cu in γ -Al₂O₃ were terminated with H atoms to obtain neutral cluster. Due to the size and edge effects, the properties estimated with the finite-size model may vary from those of the real system to some extent. However, it can be expected that the results obtained with the current model would be qualitatively reliable in predicting the local chemical properties.

Quantification of •OH by the TPA probe method: The generation of •OH in the OH-CCN/CuCo-Al₂O₃ aqueous dispersion was quantitatively measured using the terephthalic acid (TPA) probe method, in which the trapping of •OH by TPA could generate the strongly fluorescent 2-hydroxyl-terephthalic acid (2-HTPA). The TPA solution was prepared with 4 mM of TPA and 12 mM of NaOH mixture. In a typical procedure, 50 mL of the TPA solution and 0.02 g of the catalyst powder were placed in a beaker. The pH value was adjusted to 6-6.5 using the aqueous hydrochloric acid solution. The suspensions were magnetically stirred for 5 min under ambient air conditions. Then, H₂O₂ with different concentrations (1 mM, 2 mM, 5 mM, 10 mM and 20 mM) was added to the suspensions under magnetic stirring throughout the experiment. At given time intervals, 3 mL aliquots were collected and filtered through a Millipore filter (pore size 0.45 μ m) to detect the H₂O₂ consumption and the 2-HTPA (•OH) production based on the fluorescence intensity. The fluorescence intensity of the produced 2-HTPA was detected by a CARY ECLIPSE fluorescence spectrophotometer. The excitation and emission wavelengths of the detector were set at 310 and 424 nm, respectively.

Simultaneously, the concentration of H_2O_2 was determined using a DPD method, as previously reported in literature.¹⁰

Measurement and calculation of TOF: The turnover frequency (TOF) of OH-CCN/CuCo- Al_2O_3 was obtained from the conversion number of H_2O_2 into $\bullet\text{OH}$ per second on one active site. It is well known that the reaction involves free radical chain reactions in solution. As the reaction time increased, the accuracy of the measured value decreased. Thus, we selected a high concentration of the TPA probe (665 mg L^{-1}) and only detected the TOF during the first 20 seconds. Nevertheless, the final measured and calculated values are still not the most accurate results, but the actual values are definitely greater than the calculated values. Therefore, it is reasonable to use this method to calculate the TOF to reflect the efficiency of $\bullet\text{OH}$ formation on the catalyst. In the first 20 s, the concentration of the generated $\bullet\text{OH}$ was measured as $7.44 \times 10^{-5} \text{ mol L}^{-1}$, when the concentration of the catalyst [$C(\text{catal.})$] was 0.4 g L^{-1} and the initial concentration of H_2O_2 [$C(\text{H}_2\text{O}_2)$] was 10 mM . The formation rate of $\bullet\text{OH}$ [$V(\bullet\text{OH})$] was thus obtained from the above $\bullet\text{OH}$ quantification experiment within the first 20 s ($V(\bullet\text{OH}) = 3.72 \times 10^{-6} \text{ mol L}^{-1} \text{ s}^{-1}$). The main active sites are the Cu centers, and their amount depends on the surface Cu concentration, which was determined by the Langmuir adsorption isotherms of phenol on the CuCo- Al_2O_3 matrix (before growing g- C_3N_4 on the surface) due to the strong complexing of Cu and phenol. In a typical procedure, 0.02 g of CuCo- Al_2O_3 was added into 50 mL of the phenol solution with phenol concentrations of 2, 5, 10, 20, 30, 50, 75, 100, and 150 mg L^{-1} . The suspensions were magnetically stirred throughout the experiment. At given time intervals, 3 mL aliquots were collected and filtered through a Millipore filter (pore size $0.45 \text{ }\mu\text{m}$) for analysis until the maximum adsorption amount of phenol was attained for each set of experiments. The concentrations of phenol were determined using a 1200 series HPLC (Agilent, U.S.A.) equipped with a ZORBAX Eclipse XDB- C_{18} column and a UV detector. The mobile phase consisted of a mixture of

methanol/water at a flow rate of 1 mL min⁻¹. The 9 maximum adsorption amount values were drawn into an adsorption curve of phenol and fitted via the Langmuir adsorption isotherm model, and then the equilibrium adsorption amount was obtained as $q_e = 7.14 \times 10^{-6}$ mol g⁻¹. Thus, the number [$N(\text{surf.})$] of active sites on the surface of the catalyst was approximately 4.3×10^{18} g⁻¹. The TOF of OH-CCN/CuCo-Al₂O₃ can be calculated by the following equation (eq. S1):

$$\text{TOF} = V(\bullet\text{OH}) \cdot N_A / [C(\text{catal.}) \cdot N(\text{surf.})] = 1.3 \text{ s}^{-1} \quad (\text{S1})$$

Calculation of the utilization efficiency of H₂O₂. The complete mineralization of one mole of benzene, phenol, benzoic acid, phloroglucinol, bisphenol A, 2-chlorophenol, 4-isopropylphenol, ibuprofen, phenytoin, diphenhydramine, 2,4-dichlorophenoxyacetic acid, methyl orange, rhodamine B, methylene blue and atrazine will theoretically consume 15, 14, 15, 12, 36, 13, 23, 33, 39, 46, 15, 43, 73, 51 and 35 moles of H₂O₂, respectively (eqs. S2~S16).

1. benzene:



2. phenol:



3. benzoic acid:



4. phloroglucinol:



5. bisphenol A:



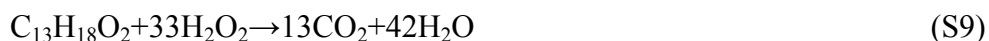
6. 2-chlorophenol:



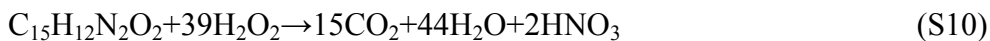
7. 4-isopropylphenol:



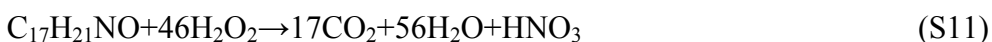
8. ibuprofen:



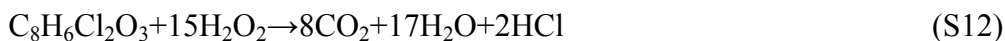
9. phenytoin:



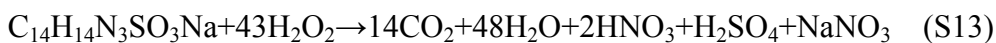
10. diphenhydramine:



11. 2,4-dichlorophenoxyacetic acid:



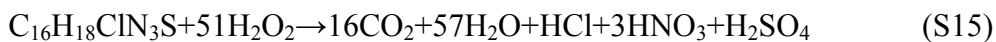
12. methyl orange:



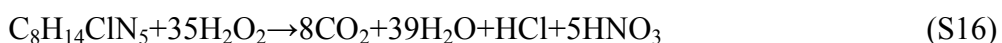
13. rhodamine B:



14. methylene blue:



15. atrazine:



The utilization efficiency of H_2O_2 (η) is defined as the ratio of the stoichiometric H_2O_2 consumption ($[\Delta\text{H}_2\text{O}_2]_s$) for the mineralization of pollutants to the actual H_2O_2 consumption ($[\Delta\text{H}_2\text{O}_2]_A$) in the Fenton-like reaction¹¹ and is expressed in eq. S17:

$$\eta = [\Delta\text{H}_2\text{O}_2]_s / [\Delta\text{H}_2\text{O}_2]_A \quad (\text{S17})$$

By measuring the TOC change in the pollutant solutions, the amounts of the mineralized contaminants were obtained, and the value of $[\Delta\text{H}_2\text{O}_2]_s$ was calculated. The actual H_2O_2 consumption ($[\Delta\text{H}_2\text{O}_2]_A$) at different reaction times was measured using the DPD method. The detailed data for $[\Delta\text{H}_2\text{O}_2]_A$ and $[\Delta\text{H}_2\text{O}_2]_s$ are presented in **Table S3**.

References

- 1 X. Zhang, Y. Ding, H. Tang, X. Han, L. Zhu and N. Wang, *Chem. Eng. J.*, 2014, **236**, 251-262.
- 2 R. Huang, Z. Fang, X. Yan and W. Cheng, *Chem. Eng. J.*, 2012, **197**, 242-249.
- 3 N. Wang, L. Zhu, M. Lei, Y. She, M. Cao and H. Tang, *ACS Catal.*, 2011, **1**, 1193-1202.
- 4 X. Yang, P.-F. Tian, C. Zhang, Y.-Q. Deng, J. Xu, J. Gong and Y.-F. Han, *Appl. Catal. B: Environ.*, 2013, **134-135**, 145-152.
- 5 A. L. Ankudinov, B. Ravel, J. J. Rehr and S. D. Conradson, *Phys. Rev. B*, 1998, **58**, 7565-7576.
- 6 A. D. Becke, *J. Chem. Phys.*, 1993, **98**, 5648-5652
- 7 M. J. Frisch et al, *Gaussian09, Revision D.01*, Gaussian, Inc., Wallingford CT, 2009.
- 8 T. Lu and F. Chen, *J. Comput. Chem.*, 2012, **33**, 580-592.
- 9 W. Humphrey, A. Dalke and K. Schulten, *J. Mol. Graphics Modell.*, 1996, **14**, 33-38.
- 10 H. Bader, V. Sturzenegger and J. Hoigne, *Water Res.*, 1988, **22**, 1109-1115.
- 11 L. Lyu, L. Zhang, Q. Wang, Y. Nie and C. Hu, *Environ. Sci. Technol.*, 2015, **49**, 8639-8647.