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1 Contrasting effects of bioturbation on metal toxicity of contaminated sediments results

- 2 in misleading interpretation of the AVS-SEM metal-sulfide paradigm
- 3

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6 Supplementary Information

7 Methods

8 General methods

9 All salts used for the preparation of solutions were of AR grade or higher. For physicochemical 10 measurements, salinity was measured using a Mettler Toledo Seven2Go S3 conductivity meter fitted 11 with an InLAB® 73X series conductivity probe; dissolved oxygen (DO) and pH were measured using 12 WTW (Wissenschaftlich-Technische Werstätten) instruments (Multi 3410 with FDO® 925 probe and 13 pH320 with SenTix 41 pH electrode). Dissolved ammonia was measured using a rapid test kit (API Fish

14 Care, LR8600). All probes were pre-calibrated using the manufacturer's instructions.

15

16 Sediment characterisation

Total recoverable metals (TRM) were determined after low-pressure microwave-assisted (MARS 5, 17 CEM) aqua regia digestion (3:1 HNO₃: HCl). Total organic carbon (TOC) analysis was determined as CO₂ 18 19 evolution following a prescribed method.¹. Dried, ground sediment samples were acid-treated to remove inorganic carbonates followed by high temperature combustion in an oxygen atmosphere 20 21 (Leco TruMAC instrument) followed by infrared detection. Frozen sediment mini-cores were extruded within a nitrogen-filled glove box. Sections were taken from the surface (~0.5 cm from the top) and 22 at depth (0.5-1 cm from the bottom) and three replicate core samples for each fraction homogenized 23 before being analyzed for AVS and SEM.² 24

- 25
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 Plant Anal. 1997, 28, 1499-1511.
- (2) Simpson, S. L. A rapid screening method for acid-volatile sulfide in sediments. Environ. Toxicol. Chem. 2001, 20 (12), 2657–
 2661.
- 30
- 31

32 Sulfidization procedure

33 Sulfidised sediment was prepared synthetically by the addition of a neutralized sulfide stock solution

34 in molar excess to the sediment SEM concentration to convert the reactive iron and metals into their

35 respective monosulfides (as shown in Equation S1).

36
$$Fe(OH)_3 + 3HS^- \rightarrow FeS + 3H_2O$$
 Equation S1

37 A neutralized sulfide solution was used because a sulfide solution prepared using just $Na_2S.9H_2O$ 38 resulted in a highly alkaline solution due to the excess of hydroxides produced during the reduction of 39 Fe(OH)₃ (Equation S2).

40
$$8Fe(OH)_3 + 9S^{2-} \rightarrow 8FeS + SO_4^{2-} + 16OH^{-} + 4H_2O$$
 Equation S2

Sulfide stock solutions (Na₂S.9H₂O (Sigma Aldrich) in deoxygenated deionized water) were neutralised 41 with 2 mol L^{-1} of H_2SO_4 (BDH) until a pH of 7.3 was reached. In a bucket purged with N_2 , neutralised 42 43 (1.5 L, 200 mmol L⁻¹) sulfide solution was mixed into 10 kg wet weight sediment and the bucket sealed 44 for 10 days, yielding a total AVS of 16 ± 0.1 mmol kg⁻¹ dry weight. This was repeated using a second neutralised sulfide solution (150 mmol L⁻¹, yielding an AVS of 16 mmol kg⁻¹ dry weight after 6 days); 45 and then a third time using 1.0 L of 200 mmol L⁻¹ sulfide solution and after 5 days, yielding an AVS of 46 47 33 ± 0.7 mmol kg⁻¹ dry weight. The overlying water was removed (3.0 L) and the sediment was then washed with 1.0 L of deoxygenated seawater. To ensure consistency with sediment treatments, the 48 49 reference sediment and non-sulfidized metal-contaminated sediments were washed with 2.5 L of deionised water then covered with 1.0 L of deoxygenated seawater a week prior to use. 50

51 **Figure S1.** Timeline for the toxicity testing phase

Toxicity test phase



Day 0 Harvest chemistry beakers Measure water physicochemistry Exchange waters Add *M. plumulosa* Feed organisms Day 6 Harvest chemistry beakers Measure water physicochemistry Exchange waters of renewal beakers Transfer *M. plumulosa* from initial beakers to renewal beakers

<u>Day 10</u>

Harvest chemistry beakers Measure water physicochemistry Count and collect *M. plumulosa* for bioaccumulation and toxicity analyses Count and collect bioturbators for bioaccumulation analyses



Results

Treatment	Day _	Simult	aneously extra	actable meta	ls (SEM) (1M	SEM	AVS	∑SEM-AVS	∑SEM-AVS/ f _{oc}				
		Fe	Mn	As	Cu	Ni	Pb	V	Zn		۹)	ເmol g⁻¹)	
No AVS _{Low}	10	0.5	76 ± 24	10 ± 2.9	165 ± 45	2.5 ± 1.0	310 ± 103	19 ± 6.4	438 ± 131	18 ± 3.1	0.6 ± 0.3	17 ± 2.7	3.1 ± 0.5
Static AVS _{Low}	0	0.5	51 ± 32	6.8 ± 4.6	108 ± 65	1.2 ± 0.9	199 ± 153	13 ± 10	286 ± 208	7.1 ± 4.9	0.7 ± 0.7	6.4 ± 4.2	1.2 ± 0.8
	6	0.3	39 ± 11	4.9 ± 1.4	88 ± 25	1.5 ± 0.5	138 ± 43	9.0 ± 3.0	219 ± 61	5.4 ± 1.5	0.3 ± 0.1	5.1 ± 1.4	0.9 ± 0.3
	10	5.5	76 ± 24	10 ± 2.9	165 ± 45	2.5 ± 1.0	310 ± 103	19 ± 6.0	483 ± 131	18 ± 2.3	0.8 ± 0.2	17 ± 2.1	3.1 ± 0.4
Bioturbated AVS _{Low}	0	0.4	47 ± 4.0	7.1 ± 1.7	105 ± 6.0	1.4 ± 0.3	190 ± 7.1	13 ± 0.7	286 ± 14	7.0 ± 0.3	0.3 ± 0.3	6.6 ± 0.5	1.2 ± 0.1
	6	0.4	69 ± 26	11 ± 3.6	172 ± 64	3.2 ± 1.5	275 ± 100	19 ± 7.4	433 ± 159	11 ± 3.9	0.7 ± 0.3	10 ± 3.7	1.9 ± 0.7
	10	1	154 ± 15	15 ± 1.3	343 ± 28	5.3 ± 0.9	664 ± 63	37 ± 4.9	865 ± 88	22 ± 2.1	0.5 ± 0.1	21 ± 2.0	3.9 ± 0.4

Table S1. Final physicochemical properties of surficial sediments for the AVS_{Low} mesocosms

For all results, n= 3 (mean ± SD). SEM = 1 M HCl extractable metals. AVS = acid-volatile sulfide. f_{oc} = fraction of organic carbon, where the SEM-AVS fraction is normalised by the fraction of organic carbon.

Treatment	Day _	Simul	taneously ext	ractable meta	als (SEM) (1N	SEM	AVS	∑SEM-AVS	∑SEM-AVS/ f _{oc}				
		Fe	Mn	As	Cu	Ni	Pb	V	Zn		ч)	umol g ⁻¹)	
No AVS _{Low}	10	0.5	59 ± 20	6.4 ± 2.1	83 ± 26	1.8 ± 0.4	228 ± 75	15 ± 5.0	337 ± 110	7.8 ± 3.0	0.2 ± 0.1	7.6 ± 2.9	1.4 ± 0.5
Static AVS _{Low}	0	0.5	27 ± 2.0	3.0 ± 0.8	48 ± 10	0.8 ± 0.2	108 ± 9.0	7.0 ± 0.2	163 ± 13	3.8 ± 0.4	0.4 ± 0.0	3.4 ± 0.4	0.6 ± 0.1
	6	0.4	93 ± 87	5.1 ± 0.4	89 ± 18	1.6 ± 0.5	162 ± 37	11 ± 3.0	253 ± 64	6.1 ± 1.4	0.3 ± 0.1	5.8 ± 1.4	1.1 ± 0.3
	10	8.5	124 ± 15	14 ± 0.8	327 ± 36	4.2 ± 0.6	507 ± 57	28 ± 4.0	674 ± 86	18 ± 5.0	0.9 ± 0.8	17 ± 5.8	3.1 ± 1.1
Bioturbated AVS _{Low}	0	0.3	43 ± 15	5.8 ± 1.4	93 ± 34	1.4 ± 0.6	189 ± 68	12 ± 4.5	277 ± 107	6.6 ± 2.5	0.5 ± 0.2	6.1 ± 2.3	1.1 ± 0.4
	6	0.3	69 ± 11	9.2 ± 3.7	164 ± 44	3.3 ± 1.1	282 ± 42	22 ± 1.2	487 ± 35	11 ± 1.4	0.9 ± 0.6	11 ± 2.1	2.0 ± 0.4
	10	8.4	126 ± 9.0	11 ± 0.3	263 ± 35	4.4 ± 0.3	562 ± 45	30 ± 2.0	722 ± 55	18 ± 1.6	0.4 ± 0.1	18 ± 1.6	3.3 ± 0.3

Table S2. Final physicochemical properties of deeper sediments for the AVS_{Low} mesocosms

For all results, n= 3 (mean ± SD). SEM = 1 M HCl extractable metals. AVS = acid-volatile sulfide. f_{oc} = fraction of organic carbon, where the SEM-AVS fraction is normalised by the fraction of organic carbon.

Treatment	Day _	Simul	taneously ext	ractable meta	als (SEM) (1N	SEM	AVS	∑SEM-AVS	Σ(ΣSEM-AVS)/ f _{oc}				
		Fe	Mn	As	Cu	Ni	Pb	V	Zn		(µmol g⁻¹)	
No AVS _{High}	10	0.4	39 ± 1.8	3.3 ± 0.3	49 ± 3.0	0.8 ± 0.2	105 ± 9.9	11 ± 1.2	185 ± 13	4.1 ± 0.3	2.3 ± 0.4	2.6 ± 0.7	1.4 ± 0.4
Static AVS _{High}	0	0.4	27 ± 5.7	2.7 ± 0.7	35 ± 7.1	0.7 ± 0.1	91 ± 17	9.6 ± 1.5	168 ± 25	3.6 ± 0.6	4.7 ± 3.0	-1.1 ± 3.5	-0.6 ± 1.8
	6	0.8	66 ± 4.3	6.5 ± 1.3	97 ± 15	2.2 ± 0.5	198 ± 36	20 ± 4.1	378 ± 71	8.3 ± 1.5	6.5 ± 5.3	1.8 ± 4.1	0.9 ± 2.2
	10	0.5	38 ± 16	3.9 ± 0.9	58 ± 22	1.3 ± 0.8	127 ± 57	11 ± 4.9	226 ± 104	5.0 ± 2.2	1.0 ± 0.5	3.9 ± 1.7	2.1 ± 0.9
Bioturbated AVS _{High}	0	0.6	37 ± 0.9	6.5 ± 1.3	70 ± 19	1.7 ± 0.9	154 ± 30	16 ± 3.3	275 ± 57	6.1 ± 1.3	6.0 ± 2.1	0.1 ± 1.0	0.05 ± 0.5
	6	0.3	22 ± 4.6	2.8 ± 0.8	47 ± 12	1.0 ± 0.3	96 ± 21	9.7 ± 2.2	117 ± 8.2	3.9 ± 0.9	2.8 ± 1.1	1.2 ± 0.3	0.6 ± 0.2
	10	0.6	37 ± 5.7	3.0 ± 0.1	63 ± 9.2	1.7 ± 0.3	166 ± 24	16 ± 3.0	295 ± 37	6.3 ± 0.8	2.8 ± 1.2	3.5 ± 0.7	1.8 ± 0.4

Table S3. Final physicochemical properties of surficial sediments for the AVS_{High} mesocosms

For all results, n= 3 (mean ± SD). SEM = 1 M HCl extractable metals. AVS = acid-volatile sulfide. f_{oc} = fraction of organic carbon, where the SEM-AVS fraction is normalised by the fraction of organic carbon.

Treatment	Day _	Simul	Simultaneously extractable metals (SEM) (1M HCl –extractable; mg kg ⁻¹ or % for Fe, dry weight)									∑SEM-AVS	∑(∑SEM-AVS), f _{oc}
		Fe	Mn	As	Cu	Ni	Pb	ν	Zn		()	µmol g⁻¹)	
No AVS _{High}	10	0.4	30 ± 1.5	1.1 ± 1.0	26 ± 6.7	0.8 ± 0.2	111 ± 7.5	13 ± 0.4	211 ± 14	4.2 ± 0.3	25 ± 6.4	-20 ± 6.2	-11 ± 3.3
Static AVS _{Hight}	0	0.5	31 ± 7.8	1.6 ± 0.5	31 ± 7.7	1.3 2) 0.2	108 ± 18	12 ± 2.8	225 ± 50	4.5 ± 0.9	7.4 ± 4.4	-2.9 ± 3.9	-1.5 ± 2.0
	6	0.6	44 ± 11	2.9 ± 1.4	42 ± 15	1.7 ± 0.3	156 ± 46	17 ± 4.6	316 ± 98	6.3 ± 2.0	14 ± 4.2	-7.2 ± 2.3	-3.8 ± 1.2
	10	0.6	39 ± 1.1	1.8 ± 0.3	34 ± 7.6	1.7 ± 0.1	149 ± 6.2	16 ± 0.2	302 ± 10	5.9 ± 0.2	19 ± 2.0	-12 ± 2.3	-6.3 ± 1.2
	0	0.5	28 ± 2.8	2.9 ± 1.4	38 ± 3.9	0.9 ± 0.2	112 ± 7.8	12 ± 0.6	211 ± 20	4.4 ± 0.3	7.0 ± 1.9	-2.6 ± 1.7	-1.4 ± 0.9
Bioturbated AVS _{High}	6	0.3	19 ± 1.6	2.4 ± 0.6	40 ± 11	0.7 ± 0.1	81 ± 14	7.4 ± 2.2	137 ± 19	3.1 ± 0.4	2.3 ± 0.5	4.1 ± 1.3	2.2 ± 0.7
	10	0.5	30 ± 3.1	2.8 ± 1.3	43 ± 9.4	1.5 ± 0.04	126 ± 16	12 ± 1.1	233 ± 24	4.9 ± 0.6	3.3 ± 0.6	1.6 ± 0.3	0.8 ± 0.2

Table S4. Final physicochemical properties of deeper sediments for AVS_{High} mesocosms

For all results, n= 3 (mean ± SD). Σ SEM = 1 M HCl extractable metals. AVS = acid-volatile sulfide. foc = fraction of organic carbon, where the Σ SEM-AVS fraction is normalised by the fraction of organic carbon.

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60

		D	issolved meta	al concentration	ıs (0.45 μm-f	iltered, μg L⁻	¹)			Mean physicochemical Parameters				
Seaiment	Fe	Mn	As	Cd	Cr	Cu	Pb	Zn	DO (%)	T (°C)	рН	NH₃ (ppm)	Salinity•	
Reference														
static ^{Tox}	5.8 ± 2.1	2.1 ± 3.4	0.8 ± 0.2	0.1 ± 0.02	0.3 ± 0.1	$2.3 \pm 0.4^{*}$	0.1 ± 0.03	7.8 ± 5.3	97 ± 2.9	20 ± 0.5	7.8 ± 0.3	0.3 ± 0.6	32 ± 0.9	
bioturb. ^{Bio}	3.8 ± 1.4	2.9 ± 2.4	0.7 ± 0.0	0.1 ± 0.1	<0.2	1.1 ± 0.1	0.1 ± 0.1	8.5 ± 3.1	98 ± 1.3	20 ± 0.3	7.9 ± 0.1	1.2 ± 0.7	32 ± 0.6	
bioturb. ^{Tox}	42 ± 105	79 ± 76	2.1 ± 0.5	<0.1	<0.2	$1.8 \pm 1.9^{*}$	0.2 ± 0.2	4.8 ± 4.8	99 ± 0.6	21 ± 0.6	7.9 ± 0.1	0.0 ± 0.1	32 ± 0.6	
AVS _{Low}														
<i>static</i> ^{NO}	2.5 ± 0.5	0.1 ± 0.2	2.3 ± 0.5	0.2 ± 0.1	0.4 ± 0.1	$4.7 \pm 1.1^{*}$	1.1 ± 0.3	15 ± 1.4	97 ± 1.2	20 ± 0.3	7.9 ± 0.1	0.1 ± 0.2	32 ± 0.8	
<i>static</i> ^{Tox}	2.1 ± 0.9	0.1 ± 0.1	2.2 ± 0.4	0.2 ± 0.1	0.3 ± 0.1	$6.9 \pm 2.3^{*}$	1.5 ± 0.5	19 ± 1.8	97 ± 2.4	20 ± 0.6	8.0 ± 0.2	0.0 ± 0.0	32 ± 0.6	
bioturb. ^{Bio}	2.8 ± 1.9	17 ± 15	2.8 ± 0.5	0.2 ± 0.1	<0.2	$5.3 \pm 2.2^{*}$	3.2 ± 0.6	18 ± 4.9	97 ± 1.7	20 ± 0.5	7.9 ± 0.2	0.7 ± 0.6	32 ± 0.6	
bioturb. ^{Tox}	3.3 ± 0.5	85 ± 28	5.1 ± 1.0	0.2 ± 0.1	<0.2	$5.7 \pm 1.6^{*}$	3.7 ± 1.5	12 ± 2.0	99 ± 0.7	21 ± 0.3	8.0 ± 0.1	0.6 ± 0.3	31 ± 0.4	
						AVS _{High}								
<i>static</i> ^{NO}	2.9 ± 0.9	0.2 ± 0.2	1.2 ± 0.5	<0.1	<0.2	$1.5 \pm 0.3^{*}$	0.1 ± 0.1	1.5 ± 0.5	96 ± 3.5	20 ± 0.4	8.2 ± 0.1	0.0 ± 0.1	32 ± 0.5	
static ^{Tox}	1.9 ± 0.4	6.2 ± 19	1.9 ± 0.4	<0.1	<0.2	$3.2 \pm 1.2^{*}$	0.2 ± 0.1	1.8 ± 0.8	94 ± 4.9	21 ± 0.4	8.2 ± 0.1	0.2 ± 0.3	32 ± 0.4	
bioturb. ^{Bio}	7.6 ± 5.6	123 ± 112	5.8 ± 3.3	<0.1	<0.2	1.0 ± 0.9	0.3 ± 0.1	3.5 ± 2.9	94 ± 5.0	20 ± 0.4	8.2 ± 0.1	1.3 ± 0.7	32 ± 0.4	
bioturb. ^{Tox}	12 ± 14	207 ± 76	3.4 ± 1.4	<0.1	<0.2	0.4 ± 0.3	0.4 ± 0.1	1.7 ± 0.7	97 ± 1.0	21 ± 0.4	8.0 ± 0.1	1.0 ± 0.6	32 ± 0.7	
WQGV (μg L	1)	750	NV	5.5	4.4 (Cr(VI))	1.3	4.4	100						

Table S5. Water chemistry and dissolved metal concentrations during the bioturbation (Bio) and Toxicity (Tox) test phases.

All data are mean ± SD (n= 12). The Water Quality Guideline values (WQGV) are reported at 95% species protection level. WQGVs for Cd, Cu, Pb, V and Zn were from ANZECC/ARMCANZ (2000). • WQGV for Mn is the recommended value currently under review Golding et al. (2016). * exceeded the WQGV. •salinity measured on the practical salinity scale (psu).

Sediment			Survival (% o	f control)			Poproduction (% of control)					
&		Renewal			Termination		Reproduction (% of control)					
treatment	Males	Females	Total	Males	Females	Total	Average	Per female	Per surviving female			
Reference												
static	89 ± 5.6	89 ± 5.6	89 ± 2.8	89 ± 5.6	83 ± 9.6	86 ± 5.6	100 ± 8.6	100 ± 8.6	100 ± 19			
bioturb.	94 ± 5.6	94 ± 5.6	94 ± 2.8	89 ± 5.6	94 ± 5.6	92 ± 0.01	94 ± 22	94 ± 22	78 ± 15			
					AVS _{Low}							
static	106 ± 11	94 ± 11	100 ± 3.1	100 ± 13	93 ± 13	97 ± 0.01	4.1 ± 3.1	4.1 ± 3.1	4.6 ± 3.8			
bioturb.	106 ± 6.2	88 ± 6.2	94 ± 5.4	94 ± 11	87 ± 6.7	90 ± 6.5	46 ± 7.5	46 ± 7.5	52 ± 11			
					AVS _{High}							
static	100 ± 6.2	106 ± 6.2	109 ± 3.1	94 ± 11	93 ± 26.7	94 ± 18	57 ± 14	57 ± 14	69 ± 21			
bioturb.	75 ± 0.01	69 ± 6.2	72 ± 3.1	81 ± 6.3	73 ± 6.7	77 ± 5.6	23 ± 4.4	23 ± 4.4	31 ± 8.2			

Table S6. Survival and sub-lethal toxicity data from the 10-day *M. plumulosa* bioassay

All data are expressed as mean \pm standard error, where n=3.

63

	Organism							Toxicity Da	ata				
			r	Number of	Average								
Mesocosm				organisms digested	Survival								
		As	Cr	Cu	Ni	Pb	Zn	N#	(% control)				
Reference													
static	Melita	10 ± 5.1	1.8 ± 1.4	72 ± 16	4.0 ± 0.3	6.1 ± 3.5	85 ± 26	6,5,5	89 ± 5.6				
	Melita	13 ± 3.0	2.5 ± 0.5	113 ± 32	5.1 ± 0.8	11 ± 1.7	163 ± 24	6,5,5	89 ± 5.6				
bioturb.	Victoriopisa	13 ± 5.1	0.7 ± 0.4	213 ± 99	1.4 ± 0.6	5.6 ± 2.7	137 ± 48	4,4,2	83 ± 17				
	Tellina	15 ± 1.8	2.6 ± 0.8	173 ± 51	3.9 ± 0.8	37 ± 6.0	383 ± 35	6,6,6	100 ± 0.0				
	AVS _{Low}												
static	Melita	9.1 ± 3.1	3.2 ± 0.6	159 ± 7.4	16 ± 2.6	8.2 ± 1.5	145 ± 42	6.4.6	100 ± 13				
	Melita	9.1 ± 2.0	4.2 ± 1.2	100 ± 15	34 ± 10	5.2 ± 3.9	155 ± 60	5.6.4	94 ± 11				
bioturb.	Victoriopisa	8.8 ± 1.9	1.4 ± 0.3	172 ± 7.5	1.0 ± 0.2	29 ± 8.8	107 ± 10	4,4,2	100 ± 20				
	Tellina	21 ± 3.3	7.8 ± 2.4	335 ± 126	5.8 ± 1.3	123 ± 30	614 ± 161	6,6,6	100 ± 0.0				
					AVS _{High}								
static	Melita	7.8 ± 3.2	1.8 ± 0.2	112 ± 18	8.1 ± 1.3	3.2 ± 0.6	113 ± 3.7	5.6.4	94 ± 11				
	Melita	9.9 ± 3.6	4.6 ± 2.4	94 ± 21	25 ± 8.1	6.7 ± 2.8	264 ± 30	5.4.4	81 ± 6.3				
bioturb.	Victoriopisa	12 ± 3.8	0.9 ± 0.3	148 ± 24	1.0 ± 0.4	7.4 ± 2.9	117 ± 38	4,4,4	120 ± 0.0				
	Tellina	19 ± 4.0	5.7 ± 2.6	290 ± 51	5.3 ± 1.3	87 ± 31	585 ± 68	6,6,6	100 ± 10				

Table S7. Organism bioaccumulation data following 10-day (*M. plumulosa*) and 27-day (*V. australiensis and T. deltoidalis*) exposures

Melita = amphipod Melita plumulosa (males only), Victoriopisa = amphipod Victoriopisa australiensis, = bivalve Tellina deltoidalis. Bioaccumulated concentrations are based on dry weight of tissue. N= 3, mean ± SE.



Figure S2. Metal accumulation data for *M. plumulosa* (top row) and bioturbators (bottom row) *V.australiensis* and *T. deltoidalis. Reference* = toxicity-control sediment, *AVS_{Low}*= metal- contaminated sediment, *AVS_{High}*= sulfidic metal-contaminated sediment. All data are mean \pm SE (n=3). Different symbols indicate statistical differences between treatments (*p*<0.05).



71

72 **Figure S3.** Relationship between sediment \sum SEM (blue triangles) and \sum SEM-AVS (open black circles) 73 concentrations: A at the end of the bioturbation phase (Day 0 of toxicity testing phase, (n=3, mean ± SD)) and B:

- 74 during the toxicity testing phase (average of Days 0, 6 and 10, (n=9, mean \pm SD)) with reproductive toxicity to *M*.
- 75 plumulosa (n=3 mean ± SE).
- 76





78 Figure S4. Relationship between water column dissolved metals (Cd, Cu, Ni, Pb and Zn) expressed as a toxic

units (TUs) throughout the toxicity test phase and reproductive toxicity in *M. plumulosa* (n=3 mean ± SE). TU = 80 Σ [dM]/WQGV), where dM = Cd, Cu, Ni, Pb, and Zn. *Reference* = toxicity-control sediment, *AVS*_{Low}= metal-

81 contaminated sediment, AVS_{High}= sulfidic metal-contaminated sediment.

- 82
- 83 Explanation for Figure S4:

84 A toxic unit (TU) approach based on water chemistry was used to assist in delineating relationships

85 between dissolved metal (Cr, Cu, Ni, Pb, Zn) exposures and observed toxicity (Figure S4 of the SI). As

86 expected for the AVS_{Low} sediment, lower toxicity in the *bioturbated* AVS_{Low} treatment corresponded

87 with slightly lower TU (TU= 5.4) than in the undisturbed *static AVS*_{Low} (TU= 5.9) sediment. Conversely,

88 the *static* AVS_{High} treatment was less toxic with higher TU (TU= 2.5), whereas the *bioturbated* AVS_{High}

89 treatment was significantly more toxic with lower TU (TU= 0.4) than the *static AVS_{Low}* treatment. These

90 predictions do not match the observed toxicity, and therefore correspond with the observations from

91 the dissolved metal chemistry and suggests that perhaps bulk water samples were not adequate for

- 92 predicting exposure and/or toxic effects.
- 93