

1 **Mutual effects of silver nanoparticles and antimony (III)/(V) co-exposed to**
2 **Glycine max (L.) Merr. in hydroponic system: uptake, translocation,**
3 **physiochemical responses and potential mechanisms**

4 Weicheng Cao^{a,b}, Jilai Gong^{a,b,*}, Guangming Zeng^{a,b,*}, Biao Song^{a,b}, Peng Zhang^{a,b},

5 Juan Li^{a,b}, Siyuan Fang^{a,b}, Lei Qin^{a,b}, Jun Ye^c and Zhe Cai^c

6 *^a College of Environmental Science and Engineering, Hunan University, Changsha*
7 *410082, P.R. China,*

8 *^b Key Laboratory of Environmental Biology and Pollution Control, Hunan University,*
9 *Ministry of Education, Changsha 410082, P.R. China*

10 *^c Hunan Qing Zhi Yuan Environmental Protection Technology Co., Ltd, Changsha*
11 *410082, P.R. China*

12

13 Pages: 29

14 Figures: 6

15 Tables: 2

16

17

18 * Corresponding author

19 E-mail address: jilaigong@gmail.com (J.L. Gong) and zgming@hnu.edu.cn (G. Zeng).

20

21 **Table S1.** Different exposure protocols of soybean to Ag NPs and Sb species in the
 22 hydroponic system

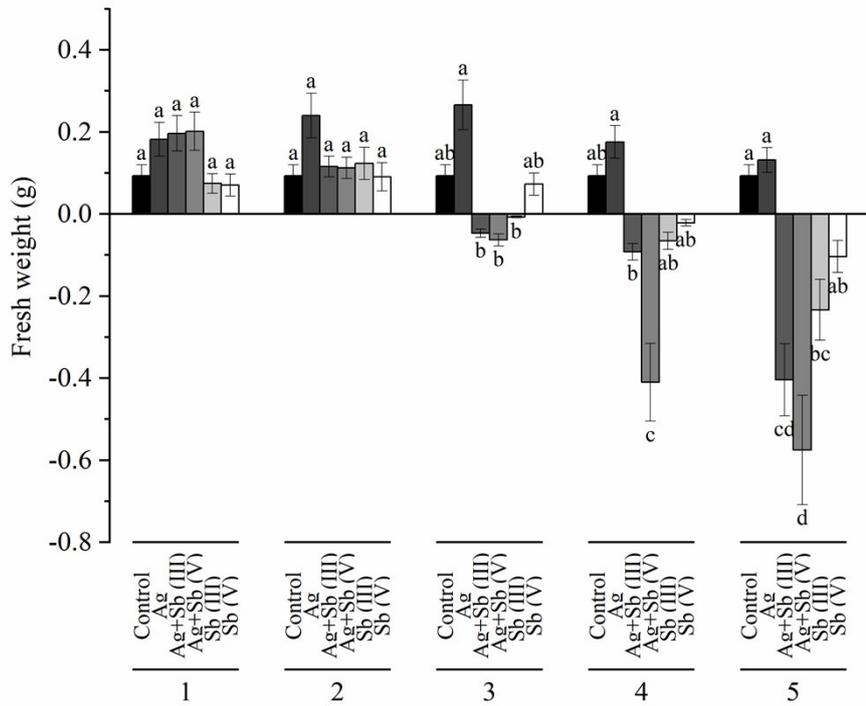
Concentration (mg L ⁻¹)	Ag NPs 1	Ag NPs 2	Ag NPs 3	Ag NPs 4	Ag NPs 5
Sb (III) / Sb (V) 1	5+0.02(N 1)	5+0.10	5+0.20	5+0.65	5+1.0
Sb (III) / Sb (V) 2	10+0.02	10+0.10(N2)	10+0.20	10+0.65	10+1.0
Sb (III) / Sb (V) 3	25+0.02	25+0.10	25+0.20(N3)	25+0.65	25+1.0
Sb (III) / Sb (V) 4	50+0.02	50+0.10	50+0.20	50+0.65(N4)	50+1.0
Sb (III) / Sb (V) 5	100+0.02	100+0.10	100+0.20	100+0.65	100+1.0(N5)

23 * The concentration of Sb (III)/ Sb (V) ranged from 5–100 mg L⁻¹ and the concentration of Ag NPs
 24 ranged from 0.02–1 mg L⁻¹. The numbers after Ag NPs/Sb (1–5) present the concentration gradient.
 25 The N1–N5 present the concentration gradients of combined exposure of Ag NPs with Sb (III)/Sb
 26 (V).
 27

29 **Table S2.** Abscission of leaves and cotyledons.

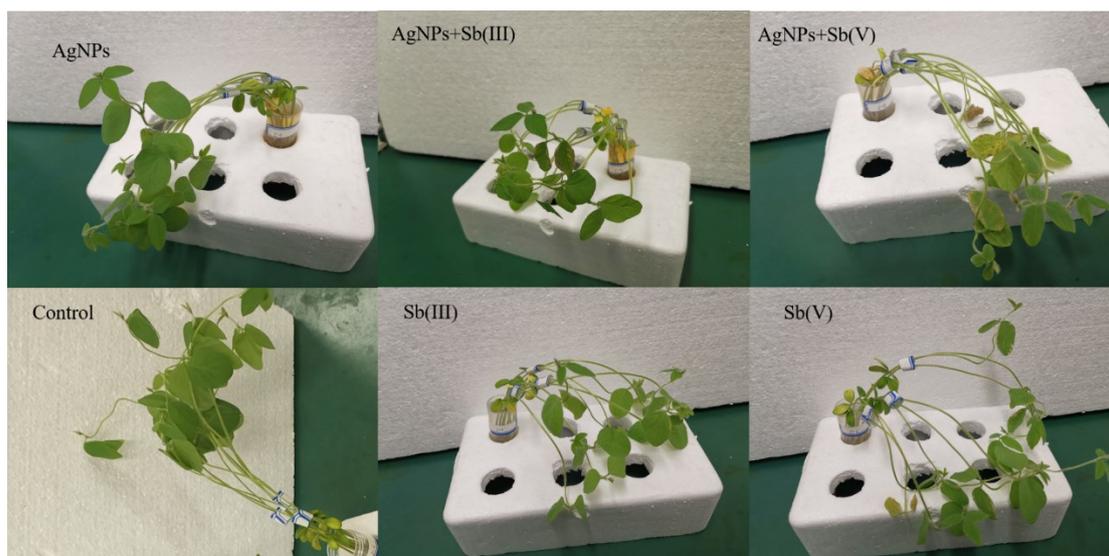
Treatments/ N5	AgNPs		AgNPs+Sb (III)		AgNPs+Sb (V)		Sb (III)		Sb (V)	
	Fall off	trait	Fall off	trait	Fall off	trait	Fall off	trait	Fall off	trait
Day 1	1 CL	NC	/	NC	1 leaf+1 CL	NC	/	NC	/	NC
Fresh weight (g)	0.0382		/		0.0657		/			
Day 2	/	NC	/	SY	3 cotyledons	SY	/	NC	/	NC
Fresh weight (g)	/		/		0.1085		/			
Day 3	/	NC	5 CLs	3 TSY	1 leaf+4 CLs	5 SLW	3 CL	7 SLW	/	3 SY
Fresh weight (g)	/		0.173		0.2436		0.2207			
Day 4	/	2 SLW	/	3 TSW	2 leaves+2 CL	1 TW+ 5 SLW	/	1 TW+ 2 SLW	/	3 SY+ 1 SLW
Fresh weight (g)	/		/		0.1951		/			
Day 5	1 CL	NC	3 CLs	1 TW+ 8 SLW	1 leaf	1 TW+ 6 LW	/	1 TW+ 2 LW+ 3 SLW	/	1 LW
Fresh weight (g)	0.0853		0.0664		0.0264		/			
Day 6	2 CL s	4 SLW	/	1 TW + 8 DLW	/	1 TW + 5 DLW	1 CL	1 TW+ 4 LW + 4 SLW	1 leaf	1 LW+ 3 SLW
Fresh weight (g)	0.0419		/		/		0.023		0.0036	
Day 7	3 CL s	4 LW	7 CLs	3 TW	/	3 TW + 6 DLW	3 CL s	1 TW+ 4 LW + 4 SLW	/	4 LW
Fresh weight (g)	0.1169		0.3075		/		0.1379			
Total Fresh weight (g)	0.2823		0.5469		0.6393		0.3816		0.0036	

30 * CL presents the abbreviation for “cotyledon”, NC is the abbreviation for “no obvious change”, (T)SY is the
31 abbreviation for “(totally) slightly yellowing”, T(S)W is the abbreviation for “total (slightly) withering of whole
32 plant” and (S/D) LW is the abbreviation for “(slightly/dramatically) leaf withering”.



33

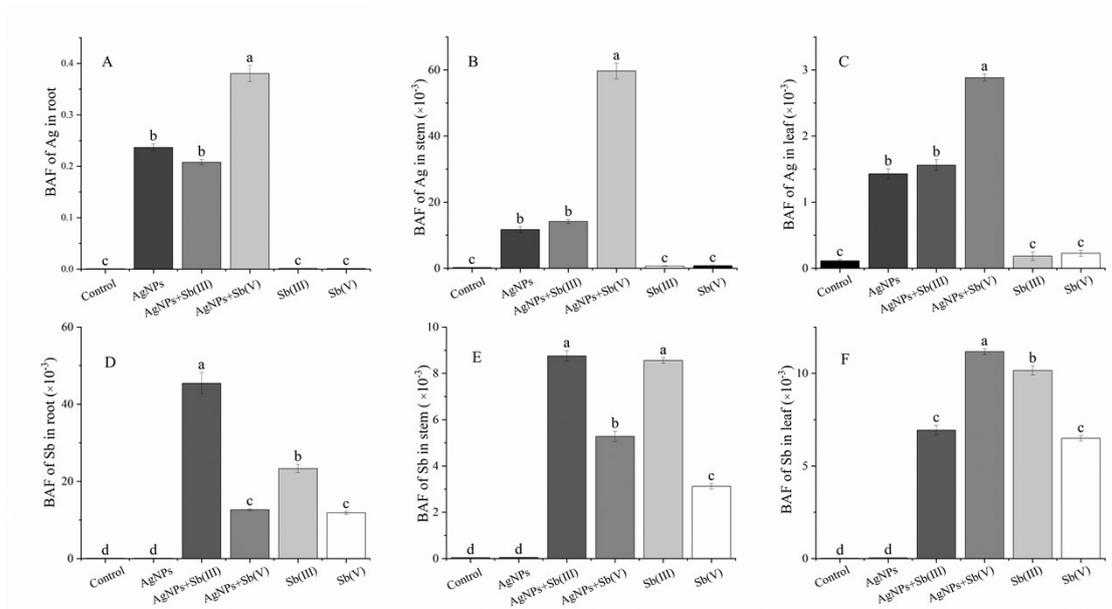
34 **Figure S1.** Effects of different treatments on fresh weight. The fresh weight of each
 35 treatment was significantly affected upon exposure to Ag NPs or/with Sb (III)/(V) with
 36 under different concentrations (see Table S1). Error bars indicate the mean standard
 37 deviation ($n = 3$) and different letters represent the significantly difference by *post hoc*
 38 Tukey's test multiple comparison ($p < 0.05$).



40

41 **Figure S2.** Photographs of soybean plantlets upon exposure to different treatments at
42 termination. Yellowing, brown spotting emerged on leaves and different luxuriant
43 degree of leaves could be observed.

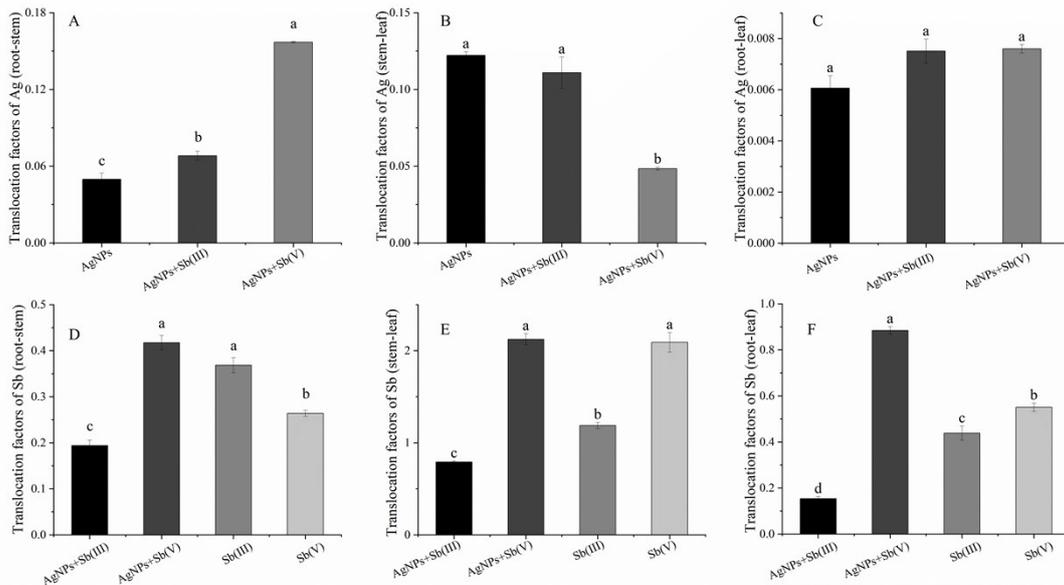
44



45

46 **Figure S3.** Bioaccumulation factors (BAF) of Ag (A, B, C) and Sb (D, E, F) in roots,
 47 stems and leave upon exposure to different treatments with Ag NPs or/with Sb (III) by
 48 N5 concentration. Error bars indicate the mean standard deviation (n = 3) and different
 49 letters represent the significantly difference by *post hoc* Tukey's test multiple
 50 comparison ($p < 0.05$).

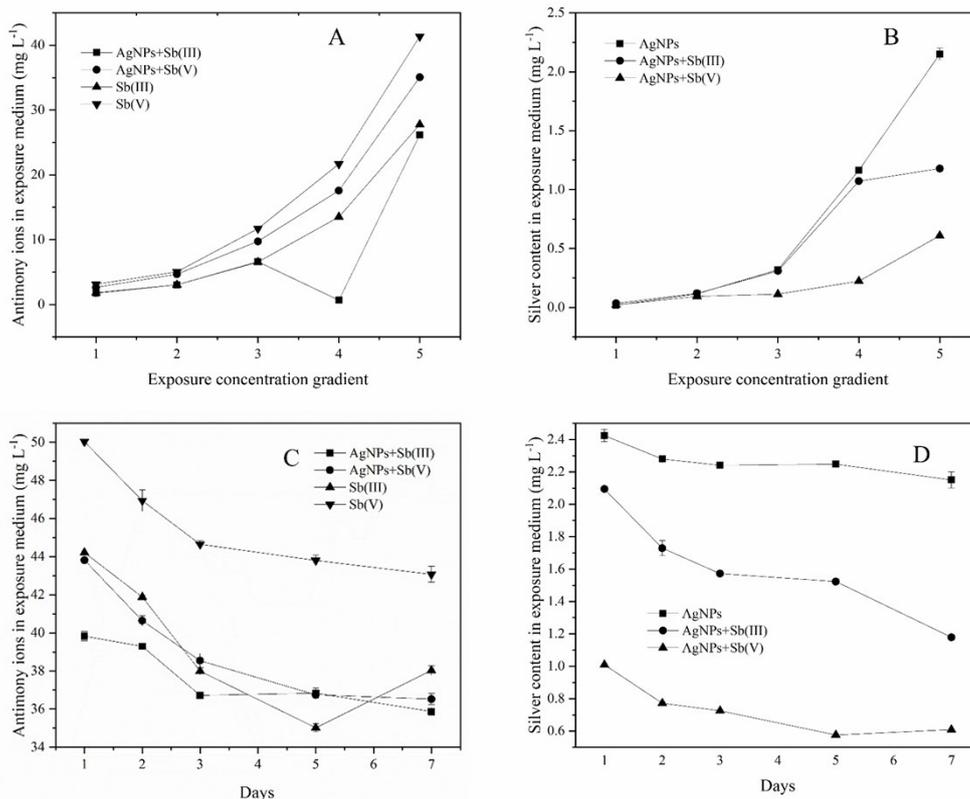
51



52

53 **Figure S4.** Translocation factors (TF) of Ag (A, B, C) and Sb (D, E, F) from roots to
 54 stems (TF_{r-s}), from stems to leaves (TF_{s-l}) and from roots to leaves (TF_{r-l}) upon exposure
 55 to different treatments with Ag NPs or/with Sb (III) by N5 concentration. Error bars
 56 indicate the mean standard deviation ($n = 3$) and different letters represent the
 57 significantly difference by *post hoc* Tukey's test multiple comparison ($p < 0.05$).

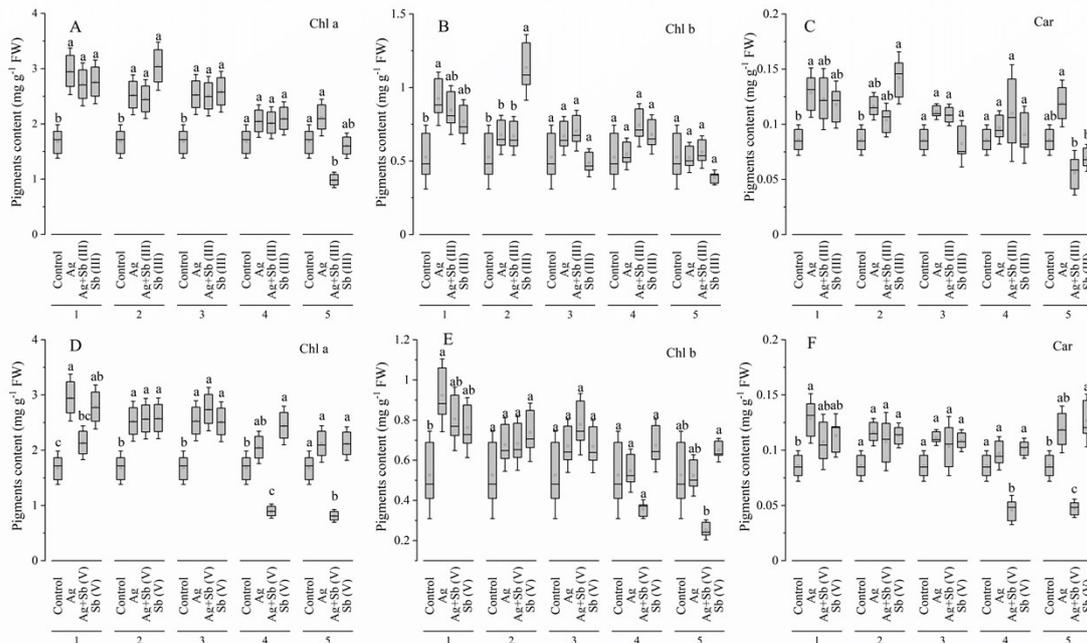
58



59

60 **Figure S5.** Residual Sb and Ag contents in culture medium when experiments
 61 terminated (A, B) and dynamic variation of Sb and Ag concentrations (see Table S1) in
 62 culture medium during the exposure period (C, D). All data were the means of three
 63 replicates ($n = 3$). Error bars indicate the mean standard deviation ($p < 0.05$).

64



65

66 **Figure S6.** Severely alteration of chlorophyll a, chlorophyll b, carotenoid contents
 67 exposed to different treatments with Ag NPs or/with Sb (III) by N1–N5 concentration
 68 (A, B, C), and Ag NPs or/with Sb (V) by N1–N5 concentrations (D, E, F). All the
 69 treatments displayed the significant promotion of pigment contents at low
 70 concentrations and gradually decreased with increasing concentration. Dots represent
 71 individual values, and boxplots display statistics (i.e., median, average, 25th and 75th
 72 percentile, and largest or smallest value extending from upper/lower quartile to 1.5-fold
 73 of the interquartile range). Error bars indicate the mean standard deviation ($n = 3$) and
 74 different letters represent the significantly difference by *post hoc* Tukey's test multiple
 75 comparison ($p < 0.05$).

76

77 **Supplemental experimental section**

78 Materials and methods

79 *Preparation of Ag NPs*

80 According to previous literature, PVP-coated Ag NPs were synthesized as
81 following.¹ Briefly, 15mL ethylene glycol (EG) containing 0.5g of PVP was heated up
82 to approximately 180 °C with vigorous stirring and kept for 30 min. Thereafter, when
83 the temperature was slowly adjusted to 120°C, 5 mL of EG containing 0.10 g AgNO₃
84 was dropwise added within 2 min and kept at 120°C for another 3 min. The Ag NPs
85 were synthesized successfully with appearance of the brown colloidal dispersion
86 generation. Subsequently, the solution was cooled to ambient temperature and excess
87 acetone was added to generate a brown precipitate. In order to remove the free Ag ions,
88 the products were rinsed and centrifuged several times (20000rpm, 30min) for
89 purification. Finally, the as-synthesized nanoparticles were resuspended in deionized
90 water as stock solution. All suspensions were stored at 4 °C in dark. For measuring the
91 concentration of stock solution, one aliquot of the suspension was withdrawn and
92 diluted and measure the maximum absorbance value in the range $\lambda = 200\text{--}800$ nm. The
93 concentration of the stock solution was determined by Beer–Lambert’s law.²⁻⁴

$$94 \quad C = \frac{A}{\epsilon L} \quad (1)$$

95 where C is the molar concentration of nanoparticle suspensions to be measured. A is
96 the absorbance of peak value of each measured sample. ϵ is the molar of silver
97 nanoparticle ($\text{cm}^{-1} \text{M}^{-1}$), which was determined according to previous literature. L is
98 the optical pathway length, which was counted as the length of the cuvette (1 cm) in

99 this study. Molar concentration of nanoparticles was converted from mM to mg L⁻¹ for
100 the unification of units in this study.

101 Dynamic light scattering (DLS) for Z-average hydrodynamic diameter was
102 conducted by a Nano Zetasizer analyzer (Malvern Zetasizer Nano-ZS, Westborough,
103 MA). The absorbance of as-synthesized Ag NPs suspension was determined by surface
104 plasmon resonance spectra recorded over the range of 200–800 nm using ultraviolet-
105 visible spectrophotometer (UV-2550, Kyoto, Japan). As reported above⁴, the
106 concentration of stock solution of Ag NPs was 30 mg L⁻¹.

107 *Plant Growth Conditions and Exposure Experiments*

108 *Glycine max* (Linn.) Merr. (soybean) seeds were obtained from commercial supplier.
109 The sterilized soybean seeds were firstly sowed in a seed tray which contained ultrapure
110 water underlying the grid layer and moist blotting paper overlaying seeds. For
111 germination, containers were transferred to a growth chamber (25-28 °C) under
112 fluorescence intensity approximately 200 μmol m⁻² s⁻¹ photons flux, with a light/dark
113 cycle of 16 : 8 h for 3 days. After germination, the young seedlings were timely
114 transplanted into 50 mL centrifugal tubes and filled with 25% Hoagland nutrient
115 solution (pH 5.8). Then the containers were transferred into growth chamber with the
116 same culture conditions for another 2 weeks. During the growth of young seedlings, the
117 25% Hoagland nutrient solution decreased by the simultaneous effects of transpiration,
118 so the corresponding experimental nutrient solution should be compensated necessarily
119 to ensure the roots dipped in the hydroponics system. The daily transpiration rates were
120 observed and recorded in terms of the water loss in an photoperiod.⁵

121 After that, the plantlets were transferred from the tubes containing Hoagland solution
122 to the new tubes with ultrapure water for 2 days to remove the surface Hoagland
123 solution from root. Then the plantlets were amended with different treatments. In
124 details, the plants were exposed to different treatments containing various individual or
125 co-exposed conditions with corresponding concentration gradient except control
126 groups: (1) Ag NPs, (2) Ag NPs+Sb (III), (3) Ag NPs+Sb (V), (4) Sb (III) and (5) Sb
127 (V). The control and exposure treatments with respective combining forms were chosen
128 and noted as described in Table S1. The exposure concentrations of Ag ($\sim 0\text{--}1\text{ mg L}^{-1}$)
129 and Sb ($\sim 5\text{--}100\text{ mg L}^{-1}$) were selected according to previous reports that could cause
130 slight toxicity alone to seedlings while enough to be detected in tissues.⁶⁻⁸ These
131 protocols would be helpful to identify the independent and collaborative uptake and
132 accumulation of the two metal elements in plants, meanwhile the impacts of Sb
133 speciation on the dissolution of Ag NPs would be evaluated. Notably, the medium
134 solution was replaced by ultrapure water during the exposure period due to the potential
135 aggregation of Ag NPs induced by the high ionic strength of the Hoagland solution.⁹

136 The exposure treatments were sustained in the same cultivated condition for 7 days.
137 Fresh weight and transpiration loss were measured and recorded every photoperiod.
138 However, because of the individual specificity of each seeding, the transpiration loss of
139 each treatment was obviously different that the compensation using treatment solution
140 would lead to relatively greater deviation of cumulated exposure dosage.⁵ Therefore,
141 the transpiration loss was replenished with ultrapure water instead of treatment
142 solutions and the dosage of water was determined according to the weight loss

143 monitored for each replicate over a 24 h interval. In the course of the exposures
144 performed, each treatment condition was triplicate prepared, and every replicate
145 contained at least six plantlets.

146 Determination of metal residue, uptake and distribution

147 The contents of Sb and Ag in medium

148 The total contents of different metallic element in solution and plant tissues were
149 determined by strong acid digestion, following Environmental Protection Agency
150 method no. 3050b (U.S) as previously reported.¹⁰ The concentration of Ag⁺ and Sb
151 dynamic variation in aqueous solution during exposure was measured by withdrawing
152 aliquots from each exposure medium at each photoperiod interval. Briefly, 0.1 mL-
153 aliquot of exposure medium of each treatment was withdrawn after every entire
154 photoperiod and collected in 10 mL centrifuge tubes. Then the subsamples were
155 digested with 1 mL of 70 % (v/v) HNO₃ at 90 °C for 2 hours and then diluted by 1%
156 nitric acid for quantification.

157 Sb and Ag contents in plant tissue

158 To evaluate the distribution of Ag and Sb in different plant tissues (including the
159 attached fraction on root surface), 50 mg (dry weight) of tissues were predigested with
160 2 mL of 70 % (v/v) nitric acid for overnight at room temperature. Then the blend was
161 further digested at 90 °C in a hot block for 4 h. After cooling to ambient temperature, 1
162 mL of 30 % hydrogen peroxide (H₂O₂) was added and incubated at 95 °C for another 2
163 h. Finally, the digestate was filtered using 0.22 µm sterile syringe filters (Millipore,
164 Billerica, MA, USA), then 0.5 mL of filtrate was withdrawn and diluted to 7 mL with

165 1 % (v/v) HNO₃. The total ion contents in different tissues were quantified by ICP-MS
166 spectrometry.

167 Enzymatic digestion for extracting metal elements from roots was conducted on fresh
168 tissues to distinguish the dissolved and particulate fractions of Ag component following
169 the recently development protocol.¹¹⁻¹³ Freshly washed tissues of roots (0.2 g) were cut
170 into small pieces, and immersed into a centrifuge tube containing 3.6 mL of 20 Mm 2-
171 (N-morpholino) ethanesulfonic acid (MES) buffer (pH 5). A handheld homogenizer
172 was applied to homogenize the mixture thoroughly with subsequent addition of 0.4 mL
173 of 30 mg/mL Macerozyme R-10 enzyme (prepared in MES above). The mixture was
174 shaken in a vibrating incubator (New Brunswick Scientific, Edison, NJ) at 37°C for 24
175 h.^{5, 11} Afterwards, 2 mL aliquot of the digestate was centrifuged using the 10 kDa
176 centrifugal ultrafiltration (Amicon Ultra-4) at 4000 rpm for 10 min to separate the
177 dissolved contents from particulate fractions. The Ag ions in the filtrate was quantified
178 by ICP-MS and considered as the dissolved fractions of Ag in root tissues. While the
179 particulate fractions were calculated by the difference between total contents and ionic
180 contents of Ag.¹⁴

181 Bioaccumulation factor (BAF), defined as the ratio of metal ion concentration in
182 plant tissues (root, stem, leaf) in response to the exposure concentration, was calculated
183 to compare and analyze the differences of Ag and Sb uptake in plants.¹⁵ While the
184 translocation factor (TF) was calculated as the ratio of concentration of metal ions in
185 different plant tissues to assess the ions distribution and translocation in plants. The two
186 factors could be calculated by the equations as follows,^{16, 17}

187
$$BAF = \frac{C_{root/stem/leaf}}{C_{ES}} \quad (2)$$

188
$$TF_{r-s} = \frac{C_{stem}}{C_{root}} \quad (3)$$

189
$$TF_{s-l} = \frac{C_{leaf}}{C_{stem}} \quad (4)$$

190
$$TF_{r-l} = \frac{C_{leaf}}{C_{root}} \quad (5)$$

191 where the $C_{root/stem/leaf}$ is the Ag or Sb ionic concentration in roots, stem and leaf
 192 respectively. C_{ES} represents the Ag or Sb ionic concentration in exposure system.
 193 TF_{r-s} , TF_{s-l} , and TF_{r-l} represents TF from roots to stems, from stems to leaves and
 194 from roots to leaves respectively.

195 Traits variation and Pigments analysis

196 The pigments contents (i.e., chlorophyll (*Chl a*, *Chl b*) and carotenoid (*Car*)),
 197 important indicators of photosynthesis efficiency and adverse physiological inhibition,
 198 were evaluated with slight modification according to previous protocol.^{18, 19} In brief,
 199 fresh leaf tissue (50 mg) was divided into strips (ap. 0.2 cm × 1 cm) and added to
 200 centrifuge tubes containing 8 mL of 95 % ethanol/acetone (1:1, v/v) mixture solution.
 201 After incubated in dark for 8 h at room temperature, the absorbance (*A*) were measured
 202 by UV-vis spectrophotometer at the wavelengths of 470, 663 and 645 nm for
 203 calculating carotenoid and chlorophyll values. The calculation equations for
 204 chlorophyll and carotenoid were expressed as follows:²⁰⁻²²

205
$$\text{Chlorophyll } a \text{ (Chl } a) \text{ (mg/g)} = \frac{(12.21 \times A_{663} - 2.81 \times A_{646})V}{1000FW} \quad (6)$$

$$206 \quad \text{Chlorophyll } b \text{ (Chl } b) \text{ (mg/g)} = \frac{(20.13 \times A_{646} - 5.03 \times A_{663})V}{1000FW} \quad (7)$$

$$207 \quad \text{Total Chlorophyll (total Chl) (mg/g)} = \frac{(7.18 \times A_{663} + 17.32 \times A_{646})V}{1000FW} \quad (8)$$

$$208 \quad \text{Car(mg/g)} = \frac{1000 \times A_{470} - 3.27 \times \text{Chl } a - 104 \times \text{Chl } b}{229 \times FW \times 1000} \quad (9)$$

209 where A_{470} , A_{645} and A_{663} is the optical density at wavelength of 647, 663 and 645 nm
 210 respectively; V is the volume of extracting solution (mL); FW is the fresh weight of
 211 plant tissues (g).

212 Measurement of anthocyanin

213 Anthocyanin level in leaf was detected which was both related to the antioxidation
 214 and light-relating process in plant.²³ In brief, 20 mg of fresh leaves were cut into pieces
 215 (ap. 0.2 cm × 1 cm), and placed in 10 ml centrifuge tubes following amended with 1.2
 216 ml of 2% HCl-MeOH solution. The tubes were sealed and incubated under
 217 ultrasonication for 5 hours. After incubation, 0.8 ml of deionized water and 2 ml of
 218 chloroform were added and fully mixed. Then the anthocyanin was separated from
 219 chlorophylls by centrifuging at 16000 rpm for 5 min. The supernatant was collected
 220 and the total anthocyanin contents were evaluated by the absorbance at 530 and 657
 221 nm. The relative content of anthocyanin was calculated as the absorbance difference (
 222 $A_{530} - A_{657}$) per gram fresh weight.^{23, 24}

223 ROS accumulation in plant tissues

224 The production of ROS in plant tissues was determined by fluorescence staining
 225 using an cell permeable indicator, 2',7'-dichlorodihydrofluorescein diacetate
 226 (H_2DCFDA), as reported in previous literatures.^{25, 26} Fresh tissues of different plant

227 parts from each treatment were weighted and washed three times with deionized water.
228 Then, the plant tissues were cut and transferred into 96 well-black plate containing 250
229 μL of $5\ \mu\text{M}$ H_2DCFDA in deionized water for 30 min in the dark at $25\ ^\circ\text{C}$. The non-
230 fluorescent compound could be taken up by cells and hydrolyzed by cellular
231 endogenous esterases to the form of 2',7'-dichlorodihydrofluorescein (H_2DCF). The
232 reduced form of H_2DCF could be better trapped in cells and transformed into a highly
233 fluorescent probe, 2',7'-dichlorofluorescein (DCF), in the presence of ROS (especially
234 hydroperoxides).^{26,27} The accumulation of ROS in different plant tissues was measured
235 and recorded using a fluorescence microplate reader (Thermo Scientific Multiskan GO,
236 USA) at 530 nm fluorescence emission with the excitation wavelength of 485 nm.^{18, 25}
237 The data of fluorescence intensity were normalized by fresh weight.

238 Lipid Peroxidation measurement

239 Lipid peroxidation usually occurred when plants was suffered from oxidative stress,
240 along with the production of malondialdehyde (MDA) which was considered as a direct
241 indicator of cellular damage and applied to evaluate the membrane integrity lever. The
242 method was slightly modified according to the protocol described by Devasagayam et.
243 al. (1987) and Jambunathan et. al. (2010).^{28, 29} In details, plant tissues were weighted
244 (0.5 g) accurately and homogenized with 2 mL of 10 % (w/v) of trichloroacetic acid
245 (TCA). Another 8 mL of TCA was added to make further homogenized. After
246 centrifuged at 4000 rpm for 15 min, the supernatant was collected and stored in 4°C .
247 Amout of 2 mL supernatant was placed in centrifuge tube and reacted with 2 mL of 0.6
248 % (w/v) of thiobarbituric acid (TBA) at $100\ ^\circ\text{C}$ for 15 min. Then the reddish brown

249 solution was fast cooled down by ice-bath before measuring the absorbance at 450, 532,
 250 and 600 nm. In addition, the reaction between MDA and TBA could be interfered by
 251 soluble saccharides, therefore the MDA content was qualified by means of a
 252 bicomponent analysis according to the Beer-Lambert's equation:

$$253 \quad A_{450} = C_1 \times 85.4 \quad (10)$$

$$254 \quad A_{532} - A_{600} = C_1 \times 7.4 + C_2 \times 155000 \quad (11)$$

255 The content of MDA was normalized with fresh weight:

$$256 \quad C_1(\text{mol g}^{-1}) = \frac{11.71 \times 10^{-3} \times A_{450}}{FW} \times V \quad (12)$$

$$257 \quad C_2(\text{mol g}^{-1}) = \frac{6.45 \times (A_{532} - A_{600}) - 0.56 \times A_{450}}{FW} \times 10^{-6} \times V \quad (13)$$

258 where A_{450} , A_{532} , A_{600} is the absorbance at 450, 532, and 600 nm respectively; V is
 259 the volume of reaction solution (times by diluted ratio); C_1 , C_2 is the content of
 260 soluble saccharides and MDA (mol g^{-1}) reacting with TBA, respectively; FW
 261 presents the fresh weight of plant tissues (g).

262 In addition, the soluble sugar content in exposure medium, which could be an
 263 imprecise indicator of the damaged plasma-membrane permeability, was analyzed as
 264 well for simply evaluating the toxic effects of different treatments. The soluble sugar
 265 content was determined in terms of the anthracenone chromogenic reaction protocol.³⁰
 266 Briefly, 0.1 ml of exposure solution was withdrawn and diluted to 0.5 ml using
 267 deionized water. The chromogenic reaction was performed by adding 2 mL of
 268 anthracenone and boiling for 15 min. Then the absorbance was detected after cooling
 269 to room temperature by a UV-vis spectrophotometer at the wavelengths of 620 nm.

270 Measurement of the activity of antioxidant enzymes

271 For observing the enzyme activities, roots, stems and leaves were separately
272 homogenized and centrifuged. The activity of superoxide dismutase (SOD) was
273 assayed following the protocol of SOD commercial kit (Nanjing Jiancheng, Nanjing,
274 China) by UV-Vis spectrophotometer³¹. One activity unit of SOD was expressed as the
275 50% reaction-inhibition of nitrobluetetrazolium (NBT) photoreduction in a 1 ml aliquot
276 of reaction solution according to the absorbance at 550 nm. The values were normalized
277 by fresh weight (FW) (U g^{-1} FW).

278 The peroxidase (POD) assay was performed according to the guaiacol oxidation
279 protocol under the presence of H_2O_2 .³² The absorbance of the oxidative product (4-o-
280 methoxyphenol) was measured at 470 nm to calculate the POD activity. In detail, 0.5 g
281 of plant tissues were accurately weighted and transferred into a mortar containing 0.2
282 g quartz sand and 0.5 ml PBS buffer solution. After sufficiently grinding, the
283 homogenate was transferred into a centrifuge tube. The mortar was rinsed with 5 mL
284 of deionized water and the rinse water was transfer into the same tube. After fixed to 8
285 mL, the tubes were centrifuged at 4°C with 8000 rpm for 15 min. The supernatant was
286 stored at 4°C. Amount of 0.1 ml supernatant was mixed with 1 ml of guaiacol solution
287 (0.1%, v/v) and 1 ml of acetate buffer (pH = 5) in centrifuge tube containing 0.9 ml
288 deionized water. The mixture was fully homogenized and deposited in water-bath at 37°C
289 for 5 min following addition of 1 ml of H_2O_2 (0.08%, v/v). The absorbance was
290 measured at 470 nm 2 min later. The blank control for each sample was prepared as
291 described above while the H_2O_2 solution was replaced by deionized water. One unit of

292 POD activity was represented as the 0.01 units of absorbance change per minute per
293 milligram of fresh weight ($\text{U min}^{-1} \text{g}^{-1} \text{FW}$).

294 Owing to the ability of catalase (CAT) assay decomposing hydrogen peroxide
295 molecules into water and oxygen rapidly, the activity of CAT was defined as the 0.01
296 units of absorbance change of H_2O_2 per minute per milligram of fresh weight (U min^{-1}
297 $\text{g}^{-1} \text{FW}$) at 240 nm.³³ In details, 1 ml aliquot of phosphate buffer ($\text{pH} = 7.8$), 0.1 ml
298 aliquot of enzyme solution and 0.9 ml of deionized water were fully mixed in a 10 ml
299 centrifuge tube and treated in water-bath at 30°C for 5 min. Thereafter, the absorbance
300 was immediately measured following addition of 1 ml aliquot of H_2O_2 solution (0.08%)
301 into the preheated tube at 240 nm. Blank control for each sample was prepared by
302 replacing the H_2O_2 solution with deionized water.

303 **Supplemental discussion**

304 ***BAFs and TFs of Ag and Sb in plant tissues***

305 The BAF values presents the Ag or Sb bioaccumulation efficiency from exposure
306 solutions to plant tissues. Results indicated that, in indivial and combined exposure
307 systems, Ag and Sb were both primarily accumulated in roots (see Figure S3).
308 Additionally, the Ag accumulated in stems were higher than that in leaves, nevertheless,
309 the accumulations of Sb in stem and in leaves were in the same order of magnitude.

310 The translocation factors ($\text{TF}_{\text{r-s}}$, $\text{TF}_{\text{s-l}}$ and $\text{TF}_{\text{r-l}}$) were conducted to evaluate the transfer
311 ability of the elements in different tissue (Figure S4). On one hand, Sb (III) in individual
312 exposure system was more readily taken up and translocated from roots to stems. On
313 the other hand, Sb (V) in co-exposure system significantly promoted the translocation

314 of Ag from roots to stems while played significant inhibition of Ag from stems to
315 leaves. As comparison, Sb (III) in co-exposure system exhibited no obvious effects on
316 the translocation of Ag no matter from roots to stem or from stem to leaves. In addition,
317 both of Sb (III) or Sb (V) in co-exposure system led to no change of the translocation
318 of Ag from roots to leaves. According to the TFs of Sb, in individual exposure system,
319 Sb (V) resulted in higher efficiency translocation of Sb to leaves than Sb (III). The
320 results were in accordance with the reports by Zhou et al..³⁴ However, the translocation
321 factor values could be affected by different plant species, because different uptake
322 capacities and different dominate storage areas of Sb in different plants could lead to
323 the variety distribution of Sb in plants.^{35, 36}

324 For co-exposure system, Ag NPs led to significant greater transfer of Sb from roots
325 to stems (by 58.1%) and then to leaves (by 60.7 %) when compared to that in individual
326 Sb (V) exposure system. Whereas, Ag NPs significantly restrained the Sb translocation
327 to stems and leaves by 47.3% and 65 % respectively. No obvious effects on the
328 translocation of Sb from stems to leaves were found in Ag NPs and Sb (V) co-exposure
329 system. Unexpected, the co-exposure of Ag NPs with Sb (III) significantly reduced the
330 translocation rate of Sb both from roots to stems and from stems to leaves (Figure S3
331 D~E). These results were identified as the effects of Ag NPs altering the Sb
332 bioavailability from hydroponic system to plants and both of Sb (V)/Sb (III) could
333 promote Ag transfer to plants though Sb (III) displayed very limited effects.

334 ***Sb and Ag concentrations in culture medium***

335 Figure S4 (A and B) shows the Sb and Ag concentrations in medium after

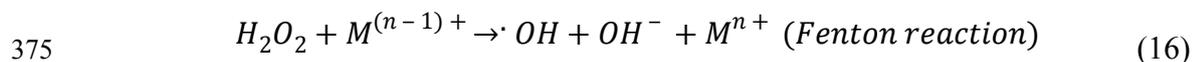
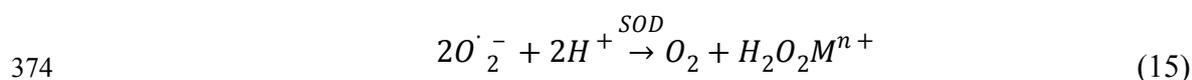
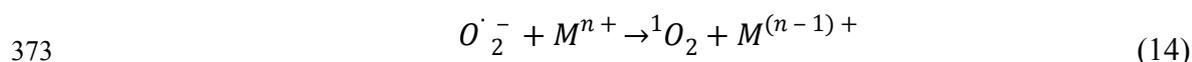
336 experiments finished under the highest concentration exposure while the dynamic
337 changes of Sb and Ag concentrations during exposure period were displayed in Figure
338 S4 (C and D). Results showed that the concentrations of total Sb contents in different
339 treatment groups displayed an order, Sb (V) > Ag NPs+ Sb (V) > Sb (III) > Ag NPs+
340 Sb (III) after exposure (see Figure S5 A). Therefore, co-exposure Sb with Ag NPs could
341 be transferred into plant more efficiently than exposure to Sb individually. These results
342 were contributed to the effect of Ag NPs altering the Sb bioavailability and
343 accumulation from hydroponic system to plants under different treatments condition.
344 On the other hand, the presence of Sb under Ag NPs exposure also influenced the Ag
345 translocation to plant (Figure S5 C). Both Sb (V) and Sb (III) in co-exposure system
346 could promote the Ag contents transfer to plants when compared with that in Ag
347 individual exposure with concentration gradient increasing (shown in Figure S5 B).
348 Whereas, Sb (V) could result in significantly concentrations reduction of total Ag
349 residual up to -71.7 % (0.61mg L^{-1}) relative to the independent exposure of Ag NPs
350 (2.15 mg L^{-1}), while Sb (III) led to a relatively lower effect (-45.2 %, ap. 1.18mg L^{-1}).
351 As shown in Figure S5 (C and D), both the Sb and Ag contents in medium were
352 gradually decreased with the experiment time extending. The total Sb content variation
353 after experiments were reduced up to 56.9% for Sb (V) individual exposure, 61.9 % for
354 Sb (III) individual exposure, 63.5 % for Ag NPs and Sb (V) co-exposure, and 64.1 %
355 for Ag NPs and Sb (III) co-exposure respectively, indicating that Ag NPs could promote
356 the translocation of Sb species to plants, especially for Sb (V). Similar in Figure S5 D,
357 Ag contents exhibited a specific dynamic variation. In which, single Ag NPs treatment

358 resulted in about 42.7 % translocation at the end of experiments, as a contrast, Ag
 359 contents in medium decreased about up to 83.8 % when treatment with Ag NPs with
 360 Sb (V) co-exposure and up to 68.6 % translocation when treatment with Ag NPs with
 361 Sb (III). These results were consistent with the conclusion in biomass increment
 362 analysis, thus demonstrating the synergistic effects of Ag NPs and Sb leading to higher
 363 accumulation both Ag and Sb contents in plants and greater negative biomass increment
 364 relative to single treatments.

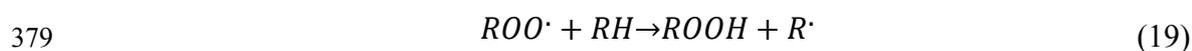
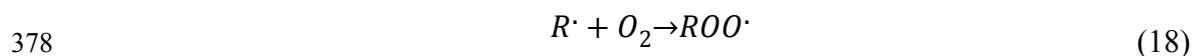
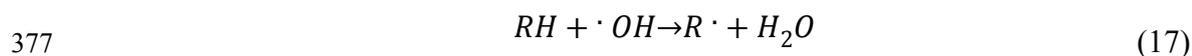
365 ***The potential generation for ROS***

366 In general, once excessive H₂O₂ are generated either from external abiotic stress or
 367 alternatively from the defense process of SOD to superoxide anion radicals. Besides,
 368 the ·OH very likely emerged during the detoxicating process of SOD to $O_2^{\cdot-}$ in which
 369 ·OH could be generated by the Haber-Weiss reaction between $O_2^{\cdot-}$ and H₂O₂ or Fenton
 370 reactions catalyzed by trace metal ions (Fe²⁺, Cu²⁺ or Zn²⁺),^{15, 37, 38} leading to higher
 371 MDA contents in plants and mediating serious peroxidation.^{53, 101, 102}

372 Fenton-like Haber-Weiss reaction during SOD catalysis:³⁷⁻⁴⁰



376 Initiation of lipid peroxidation,⁴¹



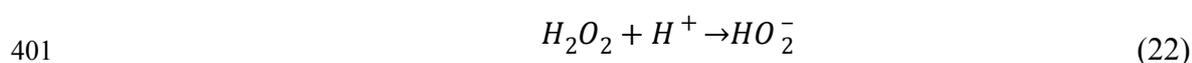
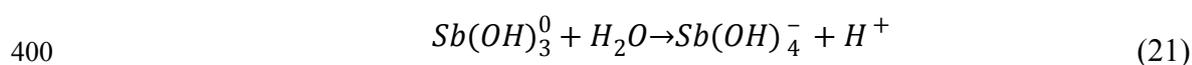


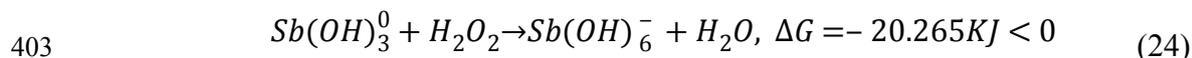
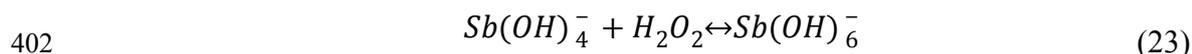
381 where, M^{n+} are the transition metals acting as electron acceptors such as Fe^{3+}/Sb^{5+} , RH
 382 represents the unsaturated fatty acid while $R\cdot$ are the lipid free radicals.

383 ***Potential valence transformation of Ag NPs and Sb species***

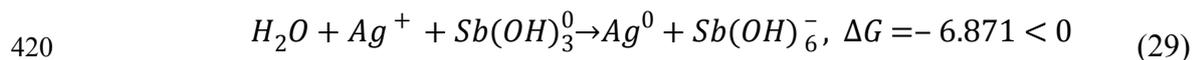
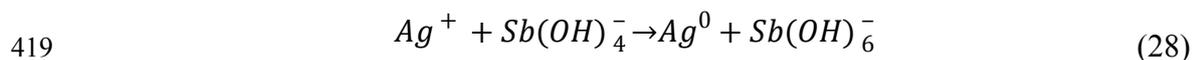
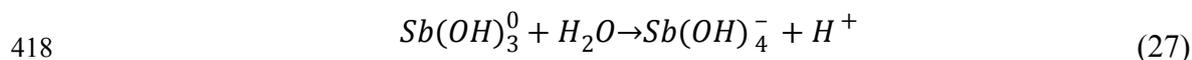
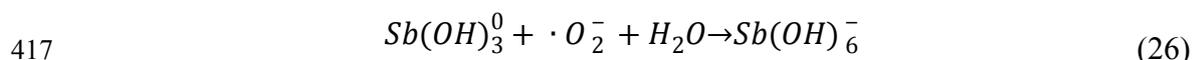
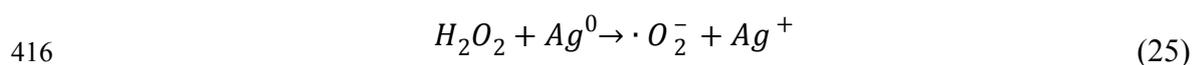
384 As described in this study, the phytotoxicity induced by single exposure of Ag NPs
 385 and Sb (III)/(V) could be obviously enhanced by co-exposure of Ag NPs and Sb
 386 (III)/(V). As reported, the phytotoxicity induced by Ag NPs were regarded as the
 387 synergetic effects between Ag^+ and Ag NPs⁴². It was revealed that Ag^+ dissolved from
 388 Ag NPs in environment occurred in cooperative oxidation involving protons and
 389 dissolved oxygen according to the global reaction stoichiometry,⁴³ and the Ag^+ was
 390 tightly correlated with bio-toxicity responses in organisms.^{6, 44}

391 The bioavailability, uptake, and toxicity of Sb were quite determined by the oxidation
 392 state and chemical speciation in aquatic systems.⁴⁵ Thus it was critical to understand
 393 the speciation stability and transformation in exposure situations. Sb generally occurred
 394 in two predominant oxidation states (i.e., Sb (III) and Sb (V)) , which was typical in
 395 natural waters.^{46, 47} As reported, Sb (III) was unstable which could be oxidized by O_2
 396 or H_2O_2 upon exposure or even storage.⁴⁸ However, Sb oxidation kinetics had been
 397 developed and validated that oxygen was not the driving force for Sb oxidation in
 398 natural waters, but it was dependent on other oxidants (H_2O_2),^{45, 48} which could be
 399 described as follow equations.^{2, 94}





404 However, when co-exposed with Sb (III), the reducibility of Ag NPs was deduced
 405 for inhibition of the Sb oxidation process. According to the Standard Electrode
 406 Potentials, the E^0 (Ag⁺/Ag), E^0 (O₂/H₂O₂) and E^0 (Sb (OH)₆⁻/Sb (OH)₃⁰) were 0.7996,
 407 0.659 and 0.764 V in circumneutral pH respectively, which exhibited the potential
 408 redox interactions between Ag and Sb species though both of them were susceptible to
 409 be oxidized by oxygen.^{49, 50} Furthermore, the zero-valent Ag on the surface of Ag NPs
 410 could promote the reduction of Sb (V) to Sb (III) and retard the oxidizing of Sb (III) to
 411 Sb (V) by oxygen, which resulted in higher accumulation of Sb (III) in tissues and
 412 greater damages in the co-exposure of Ag NPs with Sb (V) or Sb (III). In addition, the
 413 interactions between Ag NPs and Sb species were deduced to modulate the dissolution
 414 of Ag⁺ from Ag NPs which correlated with Ag⁺ accumulation in vivo and biochemical
 415 responses. The interactions between Ag NPs and Sb could be expressed as follows:



421 Accordingly, neither Sb (III) nor Sb (V) could promote the release of Ag ions. This
 422 might be supported by the decreased ionic Ag contents in roots. The uptake of Sb had
 423 been validated as a valence-dependent process and Sb predominately accumulated in

424 the roots of plants.⁵¹ Ren et al. found that the concentrations of Sb (III) in exposure
425 medium and its accumulation in plant tissues in proportion to the total Sb were quite
426 higher than that exposed to Sb (V).⁵² Whereas, the translocation efficiency of Sb (III)
427 were lower than that of Sb (V) though it might be dependent on the species of plant.⁵³
428 Additionally, the phytotoxicity and selective uptake had also been reported to be
429 dependent on plant species in which Sb (V) could exhibited higher toxicity than Sb (III)
430 treatments.⁵⁴

431 **Reference**

- 432 1. Li, L.; Sun, J.; Li, X.; Zhang, Y.; Wang, Z.; Wang, C.; Dai, J.; Wang, Q.,
433 Controllable synthesis of monodispersed silver nanoparticles as standards for
434 quantitative assessment of their cytotoxicity. *Biomaterials* **2012**, *33* (6), 1714-21.
- 435 2. Shang, J.; Gao, X., Nanoparticle counting: towards accurate determination of the
436 molar concentration. *Chemical Society reviews* **2014**, *43* (21), 7267-7278.
- 437 3. Navarro, J. R.; Werts, M. H., Resonant light scattering spectroscopy of gold, silver
438 and gold-silver alloy nanoparticles and optical detection in microfluidic channels.
439 *Analyst* **2013**, *138* (2), 583-592.
- 440 4. Paramelle, D.; Sadovoy, A.; Gorelik, S.; Free, P.; Hobley, J.; Fernig, D. G., A rapid
441 method to estimate the concentration of citrate capped silver nanoparticles from UV-
442 visible light spectra. *Analyst* **2014**, *139* (19), 4855-4861.
- 443 5. Rossi, L.; Sharifan, H.; Zhang, W.; Schwab, A. P.; Ma, X., Mutual effects and in
444 planta accumulation of co-existing cerium oxide nanoparticles and cadmium in
445 hydroponically grown soybean (*Glycine max* (L.) Merr.). *Environmental Science:*
446 *Nano* **2018**, *5* (1), 150-157.
- 447 6. Wang, J.; Koo, Y.; Alexander, A.; Yang, Y.; Westerhof, S.; Zhang, Q.; Schnoor,
448 J. L.; Colvin, V. L.; Braam, J.; Alvarez, P. J., Phytostimulation of poplars and
449 *Arabidopsis* exposed to silver nanoparticles and Ag⁺ at sublethal concentrations.
450 *Environmental science & technology* **2013**, *47* (10), 5442-5449.
- 451 7. WHO, Antimony in Drinking Water. Background Document for Development of
452 WHO Guidelines for Drinking Water Quality. WHO/SDE/WSH/03, 04/74. World
453 Health Organization, Geneva. **2003**.
- 454 8. Müller, K.; Daus, B.; Mattusch, J.; Vetterlein, D.; Merbach, I.; Wennrich, R.,
455 Impact of arsenic on uptake and bio-accumulation of antimony by arsenic
456 hyperaccumulator *Pteris vittata*. *Environmental pollution* **2013**, *174*, 128-133.
- 457 9. Li, X.; Lenhart, J. J.; Walker, H. W., Aggregation kinetics and dissolution of
458 coated silver nanoparticles. *Langmuir* **2012**, *28* (2), 1095-1104.
- 459 10. USEPA, Acid Digestion of Sediments, Sludges, and Soils, in Test Methods for

- 460 Evaluating Solid Waste Physical. *SW-846, Environmental Protection Agency, Office of*
461 *Solid Waste*, **1996**, p. 12.
- 462 11. Stowers, C.; King, M.; Rossi, L.; Zhang, W.; Arya, A.; Ma, X., Initial Sterilization
463 of Soil Affected Interactions of Cerium Oxide Nanoparticles and Soybean Seedlings
464 (*Glycine max* (L.) Merr.) in a Greenhouse Study. *ACS Sustainable Chemistry &*
465 *Engineering* **2018**, *6* (8), 10307-10314.
- 466 12. Zhang, W.; Dan, Y.; Shi, H.; Ma, X., Elucidating the mechanisms for plant uptake
467 and in-planta speciation of cerium in radish (*Raphanus sativus* L.) treated with cerium
468 oxide nanoparticles. *Journal of Environmental Chemical Engineering* **2017**, *5* (1), 572-
469 577.
- 470 13. Dan, Y.; Zhang, W.; Xue, R.; Ma, X.; Stephan, C.; Shi, H., Characterization of
471 gold nanoparticle uptake by tomato plants using enzymatic extraction followed by
472 single-particle inductively coupled plasma-mass spectrometry analysis. *Environmental*
473 *science & technology* **2015**, *49* (5), 3007-3014.
- 474 14. Li, C. C.; Dang, F.; Li, M.; Zhu, M.; Zhong, H.; Hintelmann, H.; Zhou, D. M.,
475 Effects of exposure pathways on the accumulation and phytotoxicity of silver
476 nanoparticles in soybean and rice. *Nanotoxicology* **2017**, *11* (5), 699-709.
- 477 15. Ma, C.; White, J. C.; Dhankher, O. P.; Xing, B., Metal-based nanotoxicity and
478 detoxification pathways in higher plants. *Environmental science & technology* **2015**,
479 *49* (12), 7109-7122.
- 480 16. Wang, X.; Li, F.; Yuan, C.; Li, B.; Liu, T.; Liu, C.; Du, Y.; Liu, C., The
481 translocation of antimony in soil-rice system with comparisons to arsenic: Alleviation
482 of their accumulation in rice by simultaneous use of Fe(II) and NO₃⁻. *The Science of*
483 *the total environment* **2019**, *650* (Pt 1), 633-641.
- 484 17. Wan, X.; Yang, J.; Lei, M., Speciation and uptake of antimony and arsenic by
485 two populations of *Pteris vittata* L. and *Holcus lanatus* L. from co-contaminated soil.
486 *Environmental science and pollution research international* **2018**, *25* (32), 32447-
487 32457.
- 488 18. Jiang, H. S.; Yin, L. Y.; Ren, N. N.; Zhao, S. T.; Li, Z.; Zhi, Y.; Shao, H.; Li, W.;
489 Gontero, B., Silver nanoparticles induced reactive oxygen species via photosynthetic
490 energy transport imbalance in an aquatic plant. *Nanotoxicology* **2017**, *11* (2), 157-167.
- 491 19. Moran, R., Formulae for determination of chlorophyllous pigments extracted with
492 n,n-dimethylformamide. *Plant physiology* **1982**, *69* (6), 1376-1381.
- 493 20. Ma, C.; Chhikara, S.; Minocha, R.; Long, S.; Musante, C.; White, J. C.; Xing,
494 B.; Dhankher, O. P., Reduced Silver Nanoparticle Phytotoxicity in *Crambe abyssinica*
495 with Enhanced Glutathione Production by Overexpressing Bacterial gamma-
496 Glutamylcysteine Synthase. *Environmental science & technology* **2015**, *49* (16),
497 10117-10126.
- 498 21. Wellburn, A. R.; Lichtenthaler, H., Formulae and Program to Determine Total
499 Carotenoids and Chlorophylls A and B of Leaf Extracts in Different Solvents. In
500 *Advances in Photosynthesis Research: Proceedings of the VIth International Congress*
501 *on Photosynthesis, Brussels, Belgium, August 1-6, 1983 Volume 2*, Sybesma, C., Ed.
502 Springer Netherlands: Dordrecht, 1984; pp 9-12.
- 503 22. Rossi, L.; Zhang, W.; Schwab, A. P.; Ma, X., Uptake, Accumulation, and in

- 504 Planta Distribution of Coexisting Cerium Oxide Nanoparticles and Cadmium in
505 Glycine max (L.) Merr. *Environmental science & technology* **2017**, *51* (21), 12815-
506 12824.
- 507 23. Yoo, J.; Shin, D. H.; Cho, M.-H.; Kim, T.-L.; Bhoo, S. H.; Hahn, T.-R., An ankyrin
508 repeat protein is involved in anthocyanin biosynthesis in Arabidopsis. *Physiol Plant*
509 **2011**, *142* (4), 314-325.
- 510 24. Fankhauser, C.; Casal, J. J., Phenotypic characterization of a photomorphogenic
511 mutant. *The Plant journal : for cell and molecular biology* **2004**, *39* (5), 747-760.
- 512 25. Oukarroum, A.; Barhoumi, L.; Pirastru, L.; Dewez, D., Silver nanoparticle
513 toxicity effect on growth and cellular viability of the aquatic plant Lemna gibba.
514 *Environ Toxicol Chem* **2013**, *32* (4), 902-907.
- 515 26. Greenberg, B. M.; Dixon, D. G.; Lampi, M. A.; Tripuranthakam, S.; Babu,
516 T. S.; Akhtar, T. A., Similar Stress Responses are Elicited by Copper and Ultraviolet
517 Radiation in the Aquatic Plant Lemna gibba: Implication of Reactive Oxygen Species
518 as Common Signals. *Plant and Cell Physiology* **2003**, *44* (12), 1320-1329.
- 519 27. Wiederschain, G. Y., The Molecular Probes handbook. A guide to fluorescent
520 probes and labeling technologies. *Biochemistry (Moscow)* **2011**, *76* (11), 1276-1276.
- 521 28. Devasagayam, T. P. A.; Tarachand, U., Decreased lipid peroxidation in the rat
522 kidney during gestation. *Biochemical and biophysical research communications* **1987**,
523 *145* (1), 134-138.
- 524 29. Jambunathan, N., Determination and Detection of Reactive Oxygen Species
525 (ROS), Lipid Peroxidation, and Electrolyte Leakage in Plants. In *Plant Stress*
526 *Tolerance: Methods and Protocols*, Sunkar, R., Ed. Humana Press: Totowa, NJ, 2010;
527 pp 291-297.
- 528 30. Yemm, E. W.; Willis, A. J., The estimation of carbohydrates in plant extracts by
529 anthrone. *The Biochemical journal* **1954**, *57* (3), 508-514.
- 530 31. Rico, C. M.; Hong, J.; Morales, M. I.; Zhao, L.; Barrios, A. C.; Zhang, J.-Y.;
531 Peralta-Videa, J. R.; Gardea-Torresdey, J. L., Effect of Cerium Oxide Nanoparticles on
532 Rice: A Study Involving the Antioxidant Defense System and In Vivo Fluorescence
533 Imaging. *Environmental science & technology* **2013**, *47* (11), 5635-5642.
- 534 32. Maehly, A. C., *Plant Peroxidases: Methods in Enzymology*. Methods in
535 Enzymology: 2019.
- 536 33. Li, X.; Ke, M.; Zhang, M.; Peijnenburg, W.; Fan, X.; Xu, J.; Zhang, Z.; Lu, T.;
537 Fu, Z.; Qian, H., The interactive effects of diclofop-methyl and silver nanoparticles on
538 Arabidopsis thaliana: Growth, photosynthesis and antioxidant system. *Environmental*
539 *pollution* **2018**, *232*, 212-219.
- 540 34. Zhou, X. J.; Sun, C. Y.; Zhu, P. F.; Liu, F., Effects of Antimony Stress on
541 Photosynthesis and Growth of Acorus calamus. *Frontiers in Plant Science* **2018**, *9*, 579.
- 542 35. Feng, R.; Wei, C.; Tu, S.; Wu, F.; Yang, L., Antimony accumulation and
543 antioxidative responses in four fern plants. *Plant and Soil* **2008**, *317* (1-2), 93-101.
- 544 36. Tisarum, R.; Lessl, J. T.; Dong, X.; de Oliveira, L. M.; Rathinasabapathi, B.; Ma,
545 L. Q., Antimony uptake, efflux and speciation in arsenic hyperaccumulator Pteris
546 vittata. *Environmental pollution* **2014**, *186*, 110-114.
- 547 37. Freinbichler, W.; Colivicchi, M. A.; Stefanini, C.; Bianchi, L.; Ballini, C.;

548 Misini, B.; Weinberger, P.; Linert, W.; Varešlija, D.; Tipton, K. F.; Della Corte,
549 L., Highly reactive oxygen species: detection, formation, and possible functions.
550 *Cellular Molecular Life Sciences* **2011**, *68* (12), 2067-2079.

551 38. Yim, M. B.; Boon Chock, P.; Stadtman, E. R., Cu,Zn superoxide dismutase
552 catalyzes hydroxyl radical production from hydrogen peroxide. *Free Radical Biology*
553 *and Medicine* **1990**, *9*, 43.

554 39. Upham, B. L.; Jahnke, L. S., Photooxidative reactions in chloroplast thylakoids.
555 Evidence for a Fenton-type reaction promoted by superoxide or ascorbate.
556 *Photosynthesis Research* **1986**, *8* (3), 235-247.

557 40. Gill, S. S.; Tuteja, N., Reactive oxygen species and antioxidant machinery in
558 abiotic stress tolerance in crop plants. *Plant physiology and biochemistry : PPB* **2010**,
559 *48* (12), 909-930.

560 41. SHIMOMURA; Osamu, Y., Superoxide-triggered Chemiluminescence of the
561 Extract of Luminous Mushroom (*Panellus stipticus*) after Treatment with Methylamine.
562 *Journal of Experimental Botany* **1991**, *42* (237), 555-560.

563 42. Miao, A. J.; Luo, Z.; Chen, C. S.; Chin, W. C.; Santschi, P. H.; Quigg, A.,
564 Intracellular uptake: a possible mechanism for silver engineered nanoparticle toxicity
565 to a freshwater alga *Ochromonas danica*. *PloS one* **2010**, *5* (12), e15196.

566 43. Liu, J.; Hurt, R. H., Ion release kinetics and particle persistence in aqueous nano-
567 silver colloids. *Environmental science & technology* **2010**, *44* (6), 2169-2175.

568 44. Yin, L.; Cheng, Y.; Espinasse, B.; Colman, B. P.; Auffan, M.; Wiesner,
569 M.; Rose, J.; Liu, J.; Bernhardt, E. S., More than the ions: the effects of silver
570 nanoparticles on *Lolium multiflorum*. *Environmental science & technology* **2011**, *45*
571 (6), 2360-2367.

572 45. Obiakor, M. O.; Tighe, M.; Pereg, L.; Wilson, S. C., Bioaccumulation,
573 trophodynamics and ecotoxicity of antimony in environmental freshwater food webs.
574 *Critical Reviews in Environmental Science and Technology* **2018**, *47* (22), 2208-2258.

575 46. Filella, M.; Belzile, N.; Chen, Y.-W., Antimony in the environment: a review
576 focused on natural waters: I. Occurrence. *Earth-Science Reviews* **2002**, *57* (1), 125-176.

577 47. Pierart, A.; Shahid, M.; Séjalon-Delmas, N.; Dumat, C., Antimony
578 bioavailability: Knowledge and research perspectives for sustainable agricultures.
579 *Journal of Hazardous Materials* **2015**, *289*, 219-234.

580 48. Herath, I.; Vithanage, M.; Bundschuh, J., Antimony as a global dilemma:
581 Geochemistry, mobility, fate and transport. *Environmental pollution* **2017**, *223*, 545-
582 559.

583 49. Filella, M.; May, P. M., Computer simulation of the low-molecular-weight
584 inorganic species distribution of antimony(III) and antimony(V) in natural waters.
585 *Geochimica et Cosmochimica Acta* **2003**, *67* (21), 4013-4031.

586 50. Kulp, T. R.; Miller, L. G.; Braiotta, F.; Webb, S. M.; Kocar, B. D.; Blum,
587 J. S.; Oremland, R. S., Microbiological reduction of Sb(V) in anoxic freshwater
588 sediments. *Environmental science & technology* **2014**, *48* (1), 218-226.

589 51. Vaculik, M.; Mrazova, A.; Lux, A., Antimony (SbIII) reduces growth, declines
590 photosynthesis, and modifies leaf tissue anatomy in sunflower (*Helianthus annuus* L.).
591 *Environmental science and pollution research international* **2015**, *22* (23), 18699-706.

- 592 52. Ren, J. H.; Ma, L. Q.; Sun, H. J.; Cai, F.; Luo, J., Antimony uptake, translocation
593 and speciation in rice plants exposed to antimonite and antimonate. *The Science of the*
594 *total environment* **2014**, *475*, 83-89.
- 595 53. Tisarum, R.; Lessl, J. T.; Dong, X.; de Oliveira, L. M.; Rathinasabapathi, B.; Ma,
596 L. Q., Antimony uptake, efflux and speciation in arsenic hyperaccumulator *Pteris*
597 *vittata*. *Environmental pollution* **2014**, *186*, 110-114.
- 598 54. Shtangeeva, I.; Steinnes, E.; Lierhagen, S., Uptake of different forms of antimony
599 by wheat and rye seedlings. *Environmental Science Pollution Research* **2012**, *19* (2),
600 502-509.

601