

Supplementary Information

Potential of the colloidal removal from petrochemical secondary effluent by coagulation-flocculation coupled with persulfate process

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Factors of PS oxidation treatment

The main reasons for the decrease of $c(\text{Fe}^{2+})$ excess system efficiency are as follows: (1) Excess Fe^{2+} can catalyze the generation of more $\text{SO}_4^{\cdot-}$ (Eq.1), and the sulfate radicals will self-scavenge (terminate the chain reaction, Eq.2);¹ (2) The rapid formation of $\cdot\text{OH}$ and $\text{SO}_4^{\cdot-}$ tends to react with Fe^{2+} (Eq.3 and Eq.4), which reduces the effective utilization of $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$;² (3) It has been reported that excess Fe^{2+} was oxidized to Fe^{3+} and then transformed to various Ferric oxyhydroxides through hydrolysis (Eq.5-7).³ Rao et al. found that Fe^{3+} , FeOH^{2+} , Fe(OH)_2^+ , and $\text{Fe}_2(\text{OH})_2^{4+}$ activated PS with low efficiency.⁴

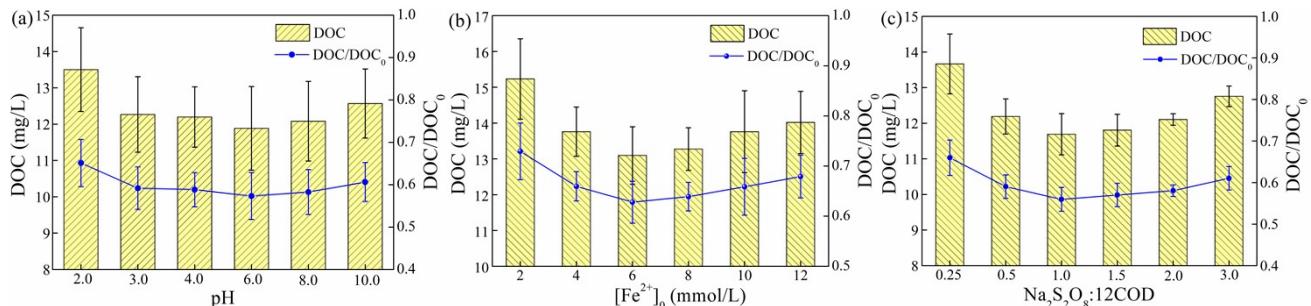
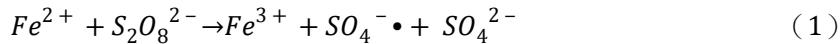
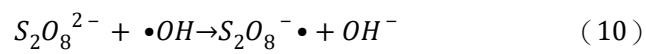
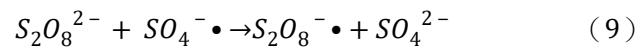
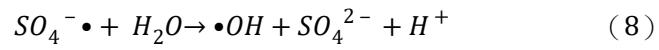


Fig. S1 Factors of DOC removal in PS process. (a) different initial pH value. Conditions: $[\text{Fe}^{2+}]_0 = 6.0 \text{ mmol/L}$, $[\text{Na}_2\text{S}_2\text{O}_8]_0 = 6.0 \text{ mmol/L}$; (b) different $[\text{Fe}^{2+}]_0$ at $[\text{Na}_2\text{S}_2\text{O}_8]_0 = 6.0 \text{ mmol/L}$; (c) different ratio of initial $\text{Na}_2\text{S}_2\text{O}_8:12\text{COD}$ at $[\text{Fe}^{2+}]_0 = 6.0 \text{ mmol/L}$.

Excess $\text{Na}_2\text{S}_2\text{O}_8$ reduces the degradation efficiency for the following three reasons: (1) Excess $\text{S}_2\text{O}_8^{2-}$ leads to elimination of $\text{SO}_4^{\cdot-}$ (Eq.1 and Eq.2);⁵ (2) Excess $\text{S}_2\text{O}_8^{2-}$ will produce more $\text{SO}_4^{\cdot-}$, and its hydrolysis will be further converted to $\cdot\text{OH}$ (Eq.2 and Eq.8);⁶ (3) Excess $\text{S}_2\text{O}_8^{2-}$ inhibits $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ (Eq.9 and Eq.10).⁷ Compared with $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$, the $\text{S}_2\text{O}_8^{\cdot-}$ has a lower REDOX potential.⁸ Therefore, the optimal $c(\text{Na}_2\text{S}_2\text{O}_8):12\text{COD}$ ratio was 1.0, that is, the dosage of PS was 1020 mg/L.



Effect of coagulation pretreatment on initial pH

As shown in Fig. S2, the optimal pH value of coagulation pretreatment was 8.0, and the removal reached 52.0%, which was 9.0% higher than that of single PS oxidation process (when the optimal pH value was 6.0). With the change of pH, the DOC removal did not change much, indicating that the pH has almost no effect on the DOC removal of the wastewater in the CF-PS process.

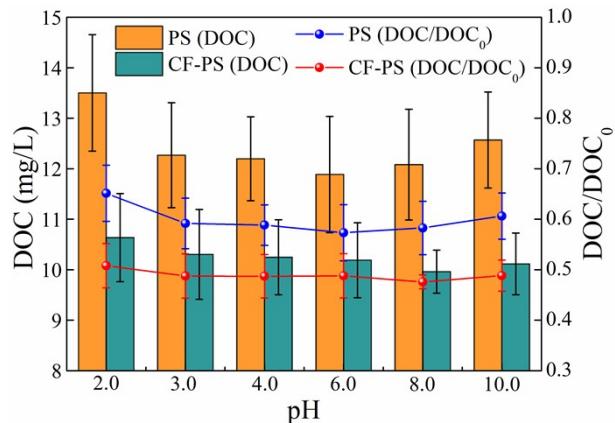


Fig. S2 The effect of initial pH on DOC removal in the PS and CF-PS processes. Conditions: $[PAC]_0 = 20.0 \text{ mg/L}$; $[Fe^{2+}]_0 = 6.0 \text{ mmol/L}$; $[Na_2S_2O_8]_0 = 6.0 \text{ mmol/L}$

Variation of colloidal content with different MWs

The formula of X removal is:

$$X\% = (A_t - A_0)/A_t * 100\% \quad (11)$$

where the A_0 (mg/L) and A_t (mg/L) represents the value of DOC/UV₂₅₄ at the initial and time t (min), respectively.

Table.S1 DOC degradation contribution of each MW fraction with CF, PS, and CF-PS processes

MWs	DOC degradation (%)			
	CF	PS	CF-PS	
High-MWs	>100 K	-64.2	-31.3	46.3
	50~100 K	28.6	-47.6	33.3
medium-MWs	30~50 K	85.1	59.8	97.6
	10~30 K	96.2	57.3	79.4
	3~10 K	84.3	85.4	19.1
low-MWs	<3 K	-4.3	40.3	40.6

Table.S2 UV₂₅₄ degradation contribution of each MW fraction with CF, PS, and CF-PS processes

MWs	UV ₂₅₄ degradation (%)			
	CF	PS	CF-PS	
High-MWs	> 100 K	-700.0	0.0	0.0
	50~100 K	0.0	100.0	0.0
Medium-MWs	30~50 K	77.0	88.9	100.0
	10~30 K	100.0	60.0	-100.0
	3~10 K	75.0	75.0	0.0
Low-MWs	< 3 K	-8.7	33.0	36.2

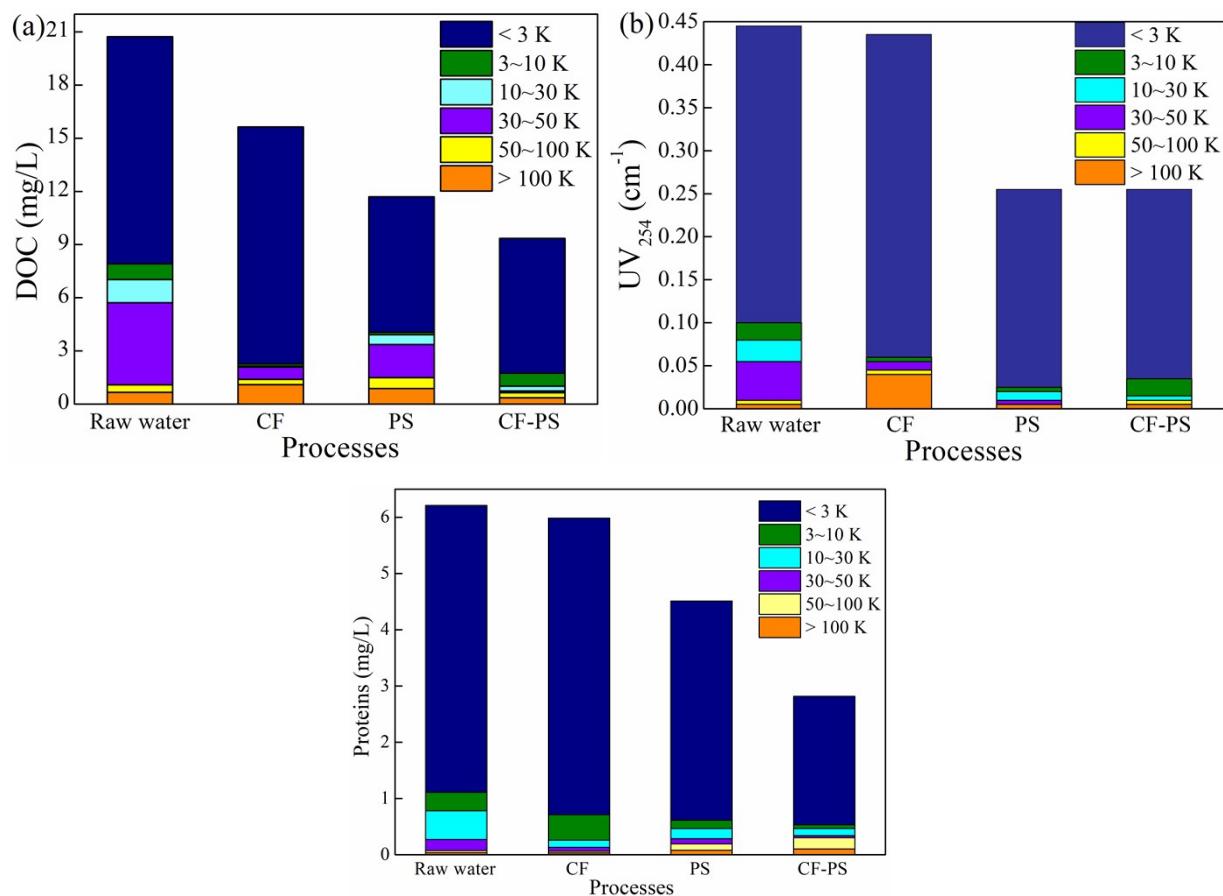


Fig. S3. MW distribution of organic matter under different processes. (a) DOC; (b) UV₂₅₄; (c) Proteins.

Fluorescent organics with different MWs

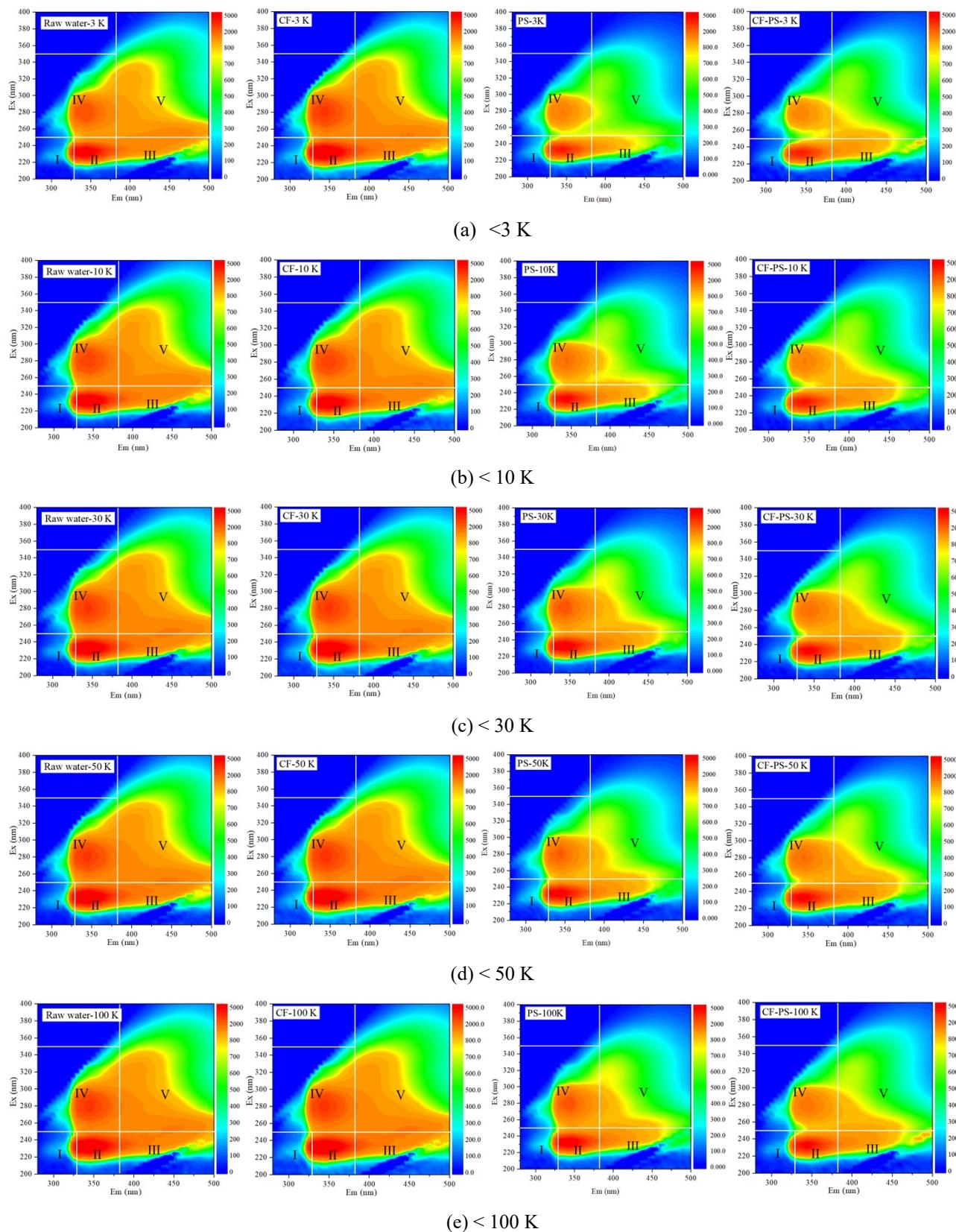


Fig. S4 3D-EEM spectra of organic components with different MWs

Table.S3 Regional fluorescence integral standard volume with different MWs

Processes	MWs	Regional fluorescence integral standard volume, $\Phi_{i,n}/\times 10^4$					
		I	II	III	IV	V	Total
		>100 K	5.40	36.22	5.28	13.50	2.51
Raw water	50~100 K	3.45	20.32	6.90	7.62	1.09	39.38
	30~50 K	8.50	26.23	6.01	11.97	1.26	53.96
	10~30 K	2.37	26.04	5.67	12.48	2.73	49.28
	3~10 K	5.22	44.82	7.30	23.62	7.73	88.69
	<3 K	53.41	258.02	88.23	129.63	87.01	616.31
	Total	78.36	411.65	119.38	198.81	102.34	910.53
CF	>100K	2.38	56.26	9.59	25.89	6.28	100.4
	50~100 K	0.56	11.04	2.04	12.47	1.70	27.82
	30~50 K	3.16	8.45	1.81	5.01	0.58	19.00
	10~30 K	0.76	11.80	2.56	-10.85	3.13	7.41
	3~10 K	3.51	29.09	4.79	14.94	-0.93	51.40
	<3 K	62.28	305.72	99.93	151.71	94.92	714.56
PS oxidation	Total	72.66	422.37	120.73	199.17	105.68	920.60
	>100 K	4.28	47.41	11.9	17.13	6.49	87.21
	50~100 K	3.90	20.81	4.40	7.68	5.25	42.04
	30~50 K	3.44	18.87	4.45	7.82	0.81	35.38
	10~30 K	1.17	12.32	0.90	2.57	1.03	17.99
	3~10 K	8.16	38.53	7.48	12.51	5.26	71.93
CF -PS oxidation	<3 K	37.15	163.45	52.02	86.02	46.47	385.11
	Total	58.10	301.39	81.14	133.72	65.30	639.66
	>100 K	2.30	38.75	1.56	9.72	3.79	56.14
	50~100 K	2.99	15.61	4.09	3.93	0.82	27.43
	30~50 K	2.99	15.61	4.09	3.93	0.82	27.43
	10~30 K	1.04	15.98	1.87	5.34	1.87	26.10
	3~10 K	1.86	13.21	0.59	6.94	1.65	24.25
	<3 K	33.78	159.71	61.94	81.21	54.12	390.75
	Total	43.52	249.69	70.88	115.75	61.78	541.62

Table. S4 Degradation of $\Phi_{i,n}$ (%) with different MWs

Processes	MWs	Removal of $\Phi_{i,n}/\%$
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		I	II	III	IV	V	Total
CF	>100 K	55.87	-55.32	-81.66	-91.83	-150.35	-59.60
	50-100 K	83.73	45.66	70.36	-63.78	-55.08	29.37
	30-50 K	62.85	67.80	69.90	58.12	53.85	64.78
	10-30 K	67.85	54.67	54.74	186.95	-14.49	84.97
	3-10 K	32.81	35.09	34.38	36.76	111.97	42.04
	<3 K	-16.60	-18.49	-13.26	-17.03	-9.08	-15.94
	Total	7.28	-2.60	-1.13	-0.18	-3.26	-1.11
PS oxidation	>100 K	20.88	-30.88	-125.34	-26.93	-158.68	-38.61
	50-100 K	-12.98	-2.37	36.17	-0.87	-380.53	-6.76
	30-50 K	59.55	28.05	25.95	34.65	36.08	34.43
	10-30 K	50.44	52.67	84.18	79.44	62.31	63.50
	3-10 K	-56.18	14.03	-2.45	47.04	31.99	18.89
	<3 K	30.44	36.65	41.04	33.65	46.60	37.51
	Total	25.86	26.78	32.03	32.74	36.19	29.75
CF -PS oxidation	>100 K	57.36	-6.98	70.42	27.96	-51.18	10.77
	50-100 K	13.42	23.22	40.69	48.45	25.13	30.35
	30-50 K	64.83	40.50	31.92	67.19	35.17	49.17
	10-30 K	56.05	38.63	67.03	57.22	31.46	47.04
	3-10 K	64.32	70.53	91.91	70.63	78.59	72.65
	<3 K	36.76	38.10	29.80	37.35	37.81	36.60
	Total	44.46	39.35	40.62	41.78	39.63	40.52

References

1. G. D. Fang , J. Gao , D. D. Dionysiou , C. Liu and D. M. Zhou , Activation of persulfate by quinones: free radical reactions and implication for the degradation of PCBs, *Environ. Sci. Technol.*, 2013, **47** , 4605 –4611.
2. W. T. Shang , Z. J. Dong , M. Li , X. L. Song , M. Zhang , C. C. Jiang and F. Y. Sun , 2019. Degradation of diatrizoate in water by Fe(II)-activated persulfate oxidation, *Chem. Eng. J.*, 2019, **361** , 1333 –1344. doi:10.1016/j.cej.2018.12.139.
3. A. Stefansson , 2007. Iron(III) hydrolysis and solubility at 25 °C, *Environ. Sci. Technol.*, 2007, **41** (17), 6117 –6123. doi:10.1021/es070174h.
4. Y. F. Rao , L. Qu , H. Yang and W. Chu , Degradation of carbamazepine by Fe(II)-activated persulfate process, *J. Hazard Mater.*, 2014, **268** , 23 –32.
5. T. Olmez-Hancı , I. Arslan-Alaton and B. Gene , Bisphenol a treatment by the hot persulfate process: oxidation products and acute toxicity, **J. Hazard Mater.**, 2013, **263** , 283 –290. doi:10.1016/j.jhazmat.2013.01.032.
6. O. S. Furman , A. L. Teel and R. J. Watts , Mechanism of base activation of persulfate, *Environ. Sci. Technol.*, 2010, **44** (16), 6423 –6428.
7. Y. Q. Gao , N. Y. Gao , Y. Deng , Y. Q. Yang and Y. Ma , Ultraviolet (UV) light-activated persulfate oxidation of sulfamethazine in water, *Chem. Eng. J.*, 2012, 195 –196, 248 –253. doi:10.1016/j.cej.2012.04.084.
8. W. Chu , Y. R. Wang and H. F. Leung , Synergy of sulfate and hydroxyl radicals in UV/ $\text{S}_2\text{O}_8^{2-}/\text{H}_2\text{O}_2$ oxidation of iodinated X-ray contrast medium iopromide, *Chem. Eng. J.*, 2011, **178**, 154 –160. doi:10.1016/j.cej.2011.10.033.