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## **Supplementary Information**

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**for**

3 **Aqueous-phase Photochemical Oxidation of Water-Soluble Brown Carbon**

4 **Aerosols Arising from Solid Biomass Fuel Burning**

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## 19 Section S1. Determination of Photon Flux

20 Photon flux indicates the absolute number of photons or light reaching the molecules  
21 present inside the aqueous solution in the photoreactor. The photon flux spectrum from 260 to  
22 400 nm was determined in this study. Chemical actinometry of 2-nitrobenzaldehyde (2NB) was  
23 employed to determine photon flux for different UV-lights, *viz.*, UVB in quartz and glass  
24 vessel, expressed as UVB(Q) and UVB(G), respectively and UVA in quartz vessel (UVA(Q))  
25 in the photoreactor. 2NB was chosen because it is thermally stable and its direct photolysis is  
26 well understood.<sup>1</sup> The direct photolysis rate of 2NB ( $J_{NB}$ ), expressed in  $\text{sec}^{-1}$ , can be simulated  
27 using following equation:<sup>1,2</sup>

$$J_{NB} (\text{sec}^{-1}) = \frac{2.303 \times 10^3}{N_A} \int_{\lambda_{min}}^{\lambda_{max}} I'_\lambda \times \phi(\lambda) \times \sigma(\lambda) \times d\lambda$$

28 Eq. S1

29 where  $N_A$  is Avogadro's number,  $I'_\lambda$  is the photon flux spectrum ( $\text{photons cm}^{-2}\text{s}^{-1}\text{nm}^{-1}$ ),  $d\lambda$  is  
30 wavelength interval between flux data points (nm),  $\phi(\lambda)$  and  $\sigma(\lambda)$  are quantum yield (molecule  
31 photon<sup>-1</sup>) and base-10 molar absorptivity ( $\text{M}^{-1}\text{cm}^{-1}$ ) of 2NB, respectively.  $\phi(\lambda)$  of 0.41 was  
32 used for all UV-lights discussed above as  $\phi(\lambda)$  has been reported to be spectrally independent  
33 in the wavelength spectrum ranging from 260 to 400 nm.<sup>1</sup>  $\sigma(\lambda)$  of 2NB changes rapidly with  
34 wavelength and adopted from a previous study.<sup>1</sup>  $I'_\lambda$  for all UV-lights in the photoreactor was  
35 determined using a combination of direct emission spectra measurement and 2NB based  
36 chemical actinometry. Firstly, the direct measurement of emission spectra of all UV-lights was  
37 carried out using a spectroradiometer (Ocean Optics, USB2000+ER). Secondly, the direct  
38 photolysis rates of 2NB corresponding to all the aforementioned UV-lights in the photoreactor  
39 were monitored using high-performance liquid chromatography (HPLC).<sup>1</sup> Further, the direct  
40 emission spectra recorded by the spectroradiometer were scaled until it matched the photon  
41 flux needed to achieve observed  $J_{NB}$ . Figure S2 shows photon flux spectra ( $\text{cm}^{-2}\text{s}^{-1}\text{nm}^{-1}$ ) for

42 UVB(Q), UVB(G) and UVA(Q) lights used during aqueous-phase direct photolysis of water-  
43 soluble brown carbon (WS-BrC) aerosols in the photoreactor and their comparison with actinic  
44 solar photon flux.

45 Further, we simulated wavelength-dependent direct photolysis rates, referred to as action  
46 spectra, of pine wood WS-BrC ( $dJ_{WS-BrC}/d\lambda$ ) as a case study to assess and compare atmospheric  
47 relevance of all three UV-lights with respect to clear-sky actinic solar photon flux of Earth  
48 retrieved from Simple Model of Atmospheric Radiative Transfer of Sunshine (SMARTS)  
49 program.<sup>3</sup> The  $dJ_{WS-BrC}/d\lambda$ , expressed in  $\text{sec}^{-1}\text{nm}^{-1}$ , was simulated from 260 to 400 nm using  
50 following equation:<sup>1,2</sup>

$$51 \frac{dJ_{WS-BrC}}{d\lambda} (\text{sec}^{-1}\text{nm}^{-1}) = I_{\lambda}' \times \phi(\lambda) \times \sigma(\lambda) \quad \text{Eq. S2}$$

52 where  $I_{\lambda}'$  is the photon flux spectrum expressed in photons  $\text{cm}^{-2}\text{s}^{-1}\text{nm}^{-1}$  (Figure S2),  $\phi(\lambda)$  is  
53 quantum yield (molecule photon<sup>-1</sup>) and  $\sigma(\lambda)$  is absorption cross-section ( $\text{cm}^2\text{molecule}^{-1}$ ) of WS-  
54 BrC.  $\phi(\lambda)$  of 1 was assumed for all photon flux conditions discussed above.  $\sigma(\lambda)$  values were  
55 based spectrophotometric measurements carried out in this study. Figure S3 shows comparison  
56 of action spectra of WS-BrC corresponding to UVB(Q), UVB(G), UVA(Q) and clear-sky  
57 actinic solar photon fluxes.

## 58 Section S2. Calculation of [OH<sub>ss</sub>] During OH oxidation

59 Pimelic acid is highly sensitive to ESI<sup>-</sup> ionization technique. For quantification of steady-  
60 state OH concentration ( $[OH_{ss}]$ ), the aqueous solution of each solid biomass fuel burning  
61 sample was spiked with 10  $\mu$ M pimelic acid and its decay during OH oxidation was monitored  
62 using an LC-MS (LTQ XL LC-MS, Thermo Scientific). The decay profiles of pimelic acid  
63 during OH oxidation of aqueous solution of all three solid biomass fuel burning samples are  
64 shown in Figure S4. The  $[OH_{ss}]$  was calculated using following equation:

$$65 \quad [OH_{ss}] = \frac{k_{decay, PA}^I}{k_{OH, PA}^{II}} \quad \text{Eq. S3}$$

66 where  $k_{decay, PA}^I$ , expressed in  $s^{-1}$ , is first-order decay constant for pimelic acid degradation  
67 during OH oxidation of WS-BrC arising from dung cakes and pine wood burning.  $k_{OH, PA}^{II}$  is  
68 OH radical second-order rate constant of pimelic acid in aqueous medium. In this study,  
69  $k_{decay, PA}^I$  was calculated using methodology above and shown in Figure S4. Moreover,  
70  $k_{OH, PA}^{II}$  value of  $2.8 \times 10^9 \text{ M}^{-1}\text{s}^{-1}$  was taken from a recent study.<sup>4</sup>

71

### 72 Section S3. Total Organic Carbon (TOC) and Total Nitrogen (TN) Measurements

73 TOC and TN concentration in aqueous extract of particulate laden filters were measured  
74 using a TOC analyzer with a TN measurement unit (Shimadzu TOC-L CPH Analyzer). The  
75 TOC concentration, referred to as water-soluble organic carbon (WSOC) in this study, was  
76 quantified using non-purgeable organic carbon (NPOC) method. In NPOC method, an aliquot  
77 of aqueous extracts was acidified with 1M HCl and subsequently sparged to remove purgeable  
78 organic (volatile fraction) and inorganic carbon. The remaining carbon in aqueous extract is  
79 then converted to carbon dioxide (CO<sub>2</sub>) gas and detected by a non-dispersive infrared detector.  
80 The NPOC method is more accurate for TOC measurements compared to Total  
81 Carbon/Inorganic Carbon method.<sup>5,6</sup> However, loss of volatile organic carbon during the  
82 purging may lead to an underestimation of TOC concentrations. For TN measurements, the  
83 extract was combusted to NO<sub>x</sub> and subsequently reacted with O<sub>3</sub> to form NO<sub>2</sub> in an excited  
84 state. Photons emitted by NO<sub>2</sub> during this process were measured by a chemiluminescence  
85 detector. The quantitation of WSOC and TN was achieved by calibrating the TOC instrument  
86 with known ACS grade standards, *i.e.*, potassium hydrogen phthalate for NPOC, and nitrate  
87 salts for TN.

88

## 89 Section S4. Calculation of Solar-weighted Total Absorbance

90 The comparison of absorption properties of WS-BrC at a single wavelength (*e.g.*, 365  
91 nm) might not be sufficient to understand their photochemical evolution over time.<sup>7,8</sup>  
92 Therefore, the solar-weighted total absorbance ( $A_{Tot-(WS-BrC)-SW}$ ) of WS-BrC aerosols emitted  
93 from different solid biomass fuel burning types was determined in this study using following  
94 equations:

$$A_{(WS-BrC)-SW-\lambda} = A_{WS-BrC-\lambda} \times \frac{S_{\lambda}}{\int_{330}^{500} S_{\lambda} d\lambda}$$

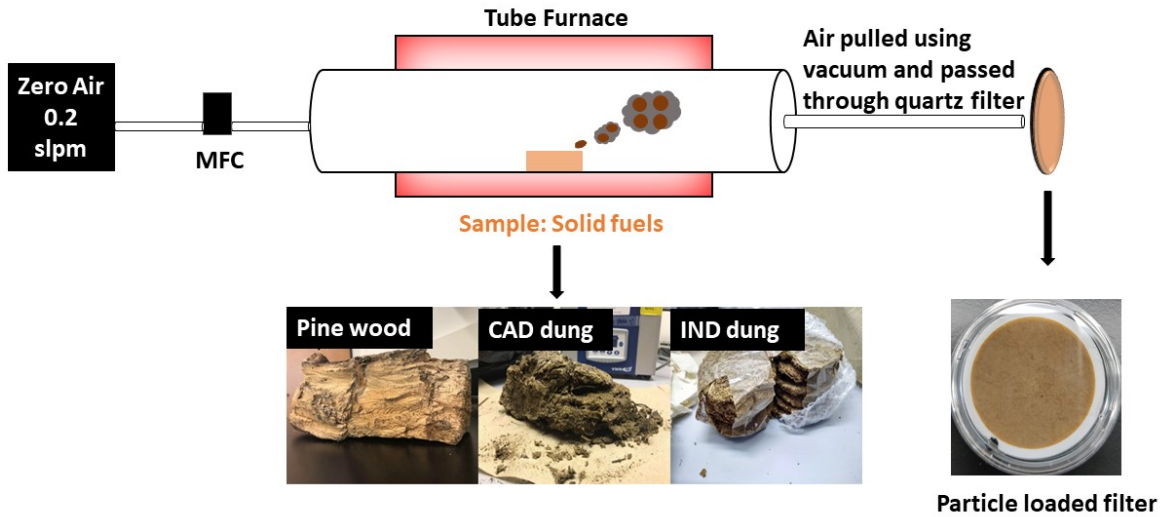
95 Eq. S4

$$A_{Tot-(WS-BrC)-SW} = \int_{330}^{500} A_{(WS-BrC)-SW-\lambda} d\lambda$$

96 Eq. S5

97 where  $A_{WS-BrC-\lambda}$  is WS-BrC absorbance at a given wavelength ( $\lambda$ ),  $A_{(WS-BrC)-SW-\lambda}$  is solar-weighted  
98 absorbance at  $\lambda$  and  $S_{\lambda}$  is clear sky spectral solar irradiance (W/m<sup>2</sup>/nm) on tilted receptor plane  
99 at  $\lambda$  and was adopted from Simple Model of Atmospheric Radiative Transfer of Sunshine  
100 (SMARTS) program.<sup>3</sup> The above integrations were carried out from 330 to 500 nm using  
101 trapezoidal rule [data interval = 1 nm].

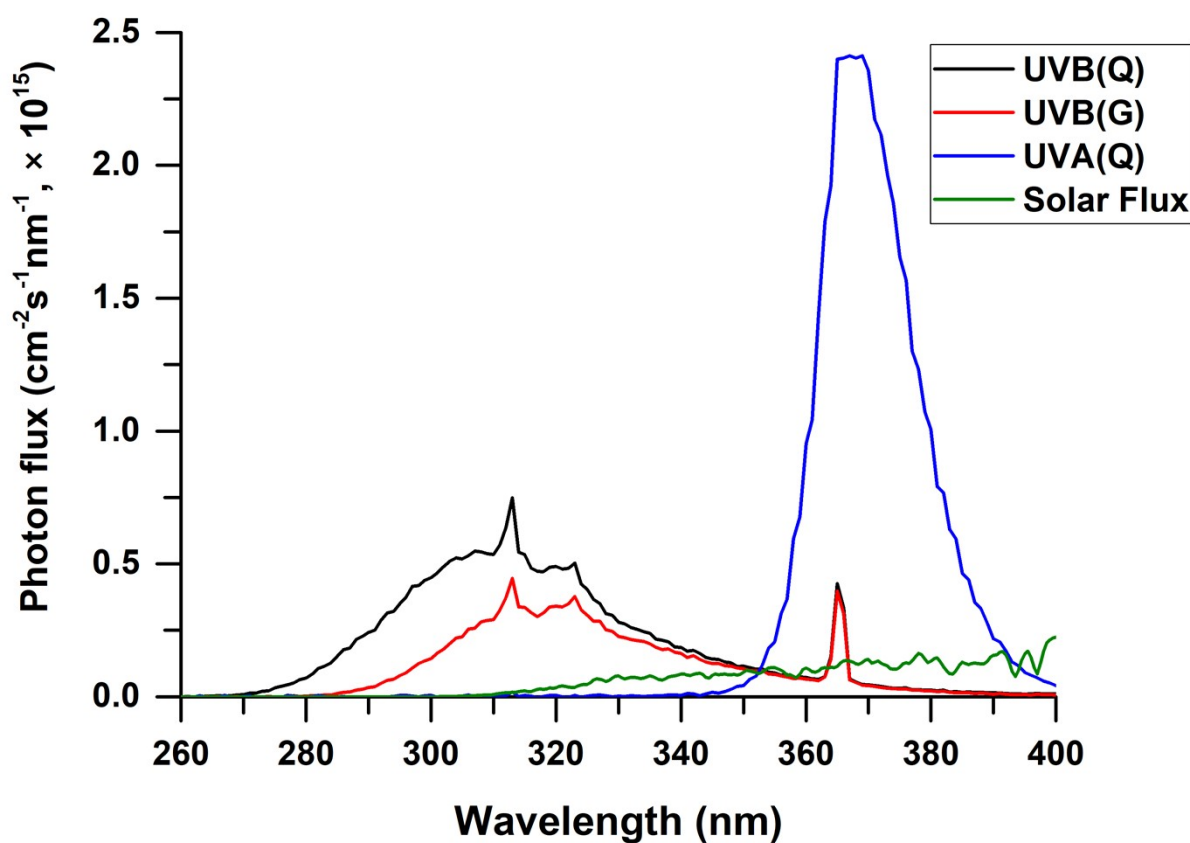
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104 **Figure S1.** Schematic of the tube furnace set-up along with pictures of solid biomass fuels  
 105 explored in this study. These solid biomass fuels were pine wood, Indian (IND) dung cakes  
 106 and Canadian (CAD) dung cakes. Moreover, MFC refers to mass flow controller.

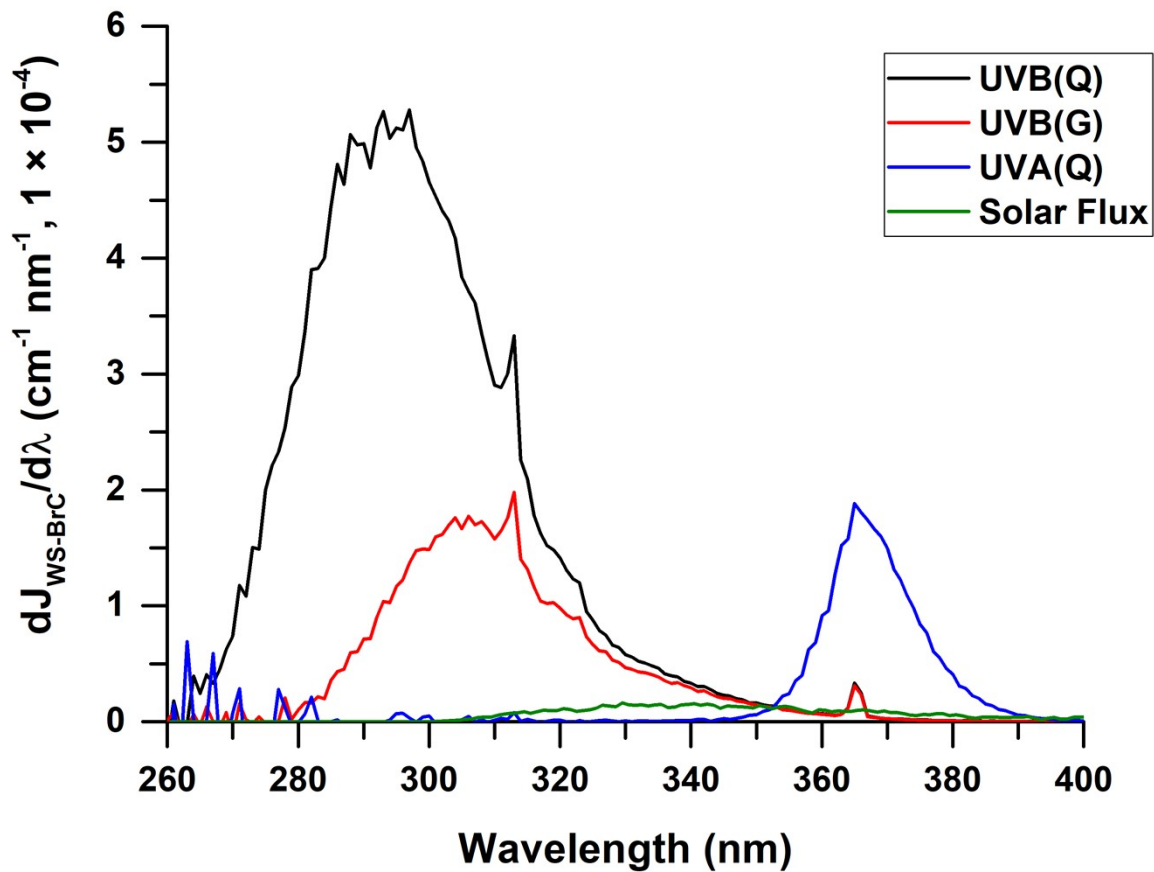
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 109 **Figure S2.** Photon flux spectra ranging from 260 to 400 nm, expressed in cm<sup>-2</sup> s<sup>-1</sup> nm<sup>-1</sup>, of  
 110 distinct ultraviolet (UV) lamps utilized during aqueous-phase direct photolysis of WS-BrC  
 111 aerosols in the photoreactor and their comparison with actinic solar photon flux. Here, UVB(Q)  
 112 and UVB(G) indicate UVB spectrum in quartz and glass vessel, respectively. UVA(Q)  
 113 indicates UVA spectrum in quartz vessel. The actinic solar photon flux spectrum is based on  
 114 Simple Model of Atmospheric Radiative Transfer of Sunshine (SMARTS) program.<sup>3</sup>

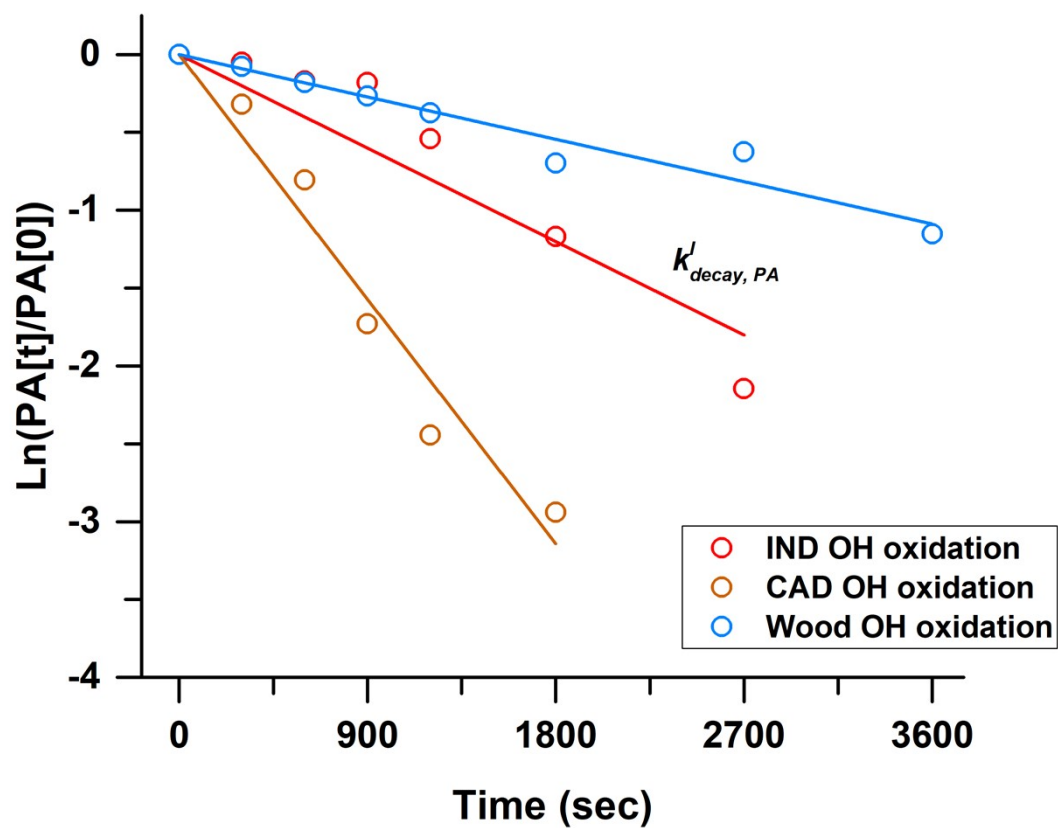
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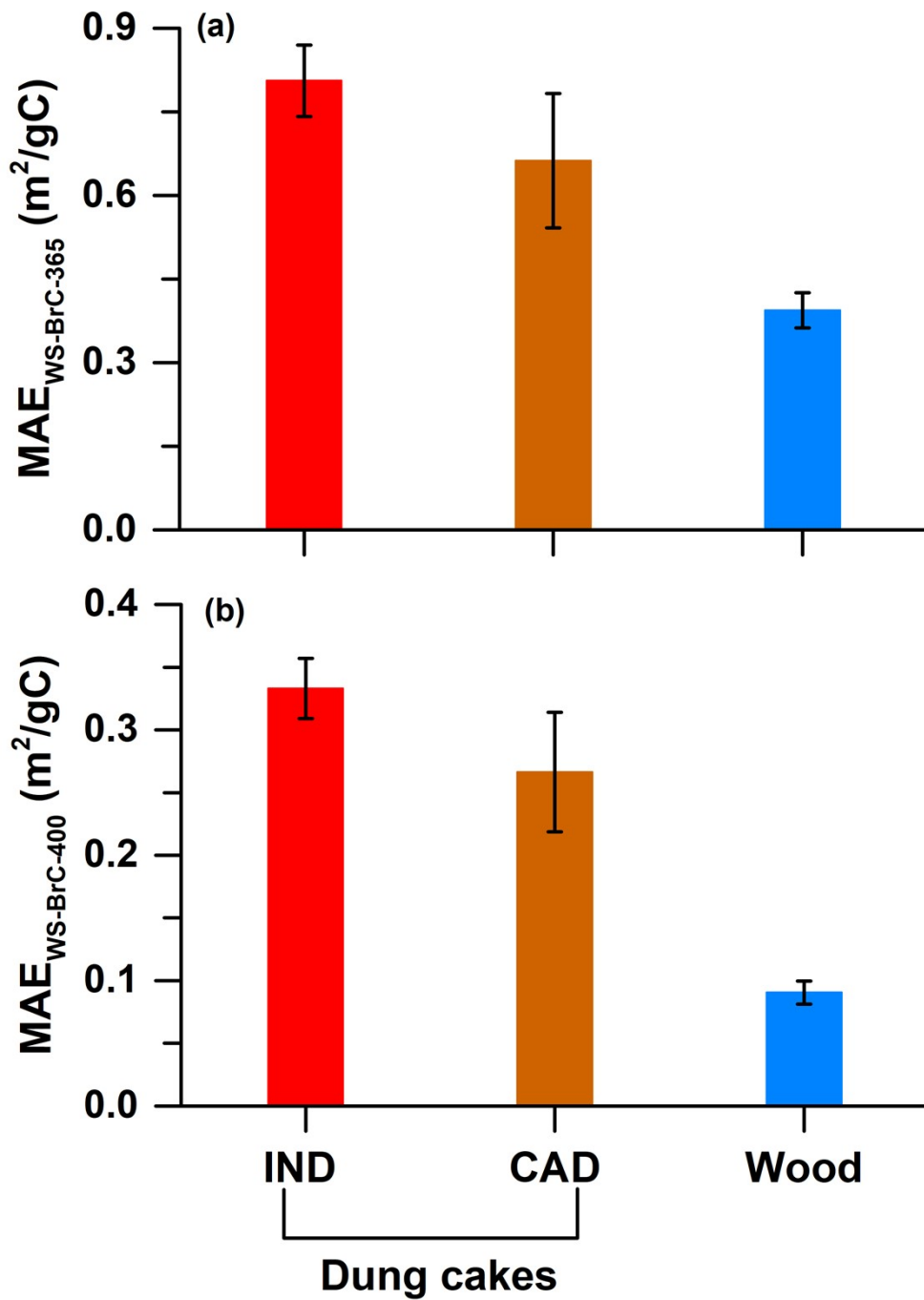
117 **Figure S3.** Comparison of action spectra, simulated for UVB(Q), UVB(G), UVA(Q) and clear-  
 118 sky actinic solar photon flux ranging from 260 to 400 nm, of WS-BrC ( $dJ_{\text{WS-BrC}}/d\lambda$ ) arising  
 119 from pine wood burning.



120

121 **Figure S4.** Photochemical decay profile of pimelic acid (PA), expressed by temporal change  
 122 in pimelic acid intensity normalized to its initial value on logarithm scale ( $\text{Ln}(\text{PA}[t]/\text{PA}[0])$ ),  
 123 during OH oxidation of WS-BrC arising from IND dung, CAD dung and pine wood burning.

124 Where,  $k_{decay, PA}^I$  indicate first-order decay constant obtained by linear regression fit of the  
 125 decay profile.

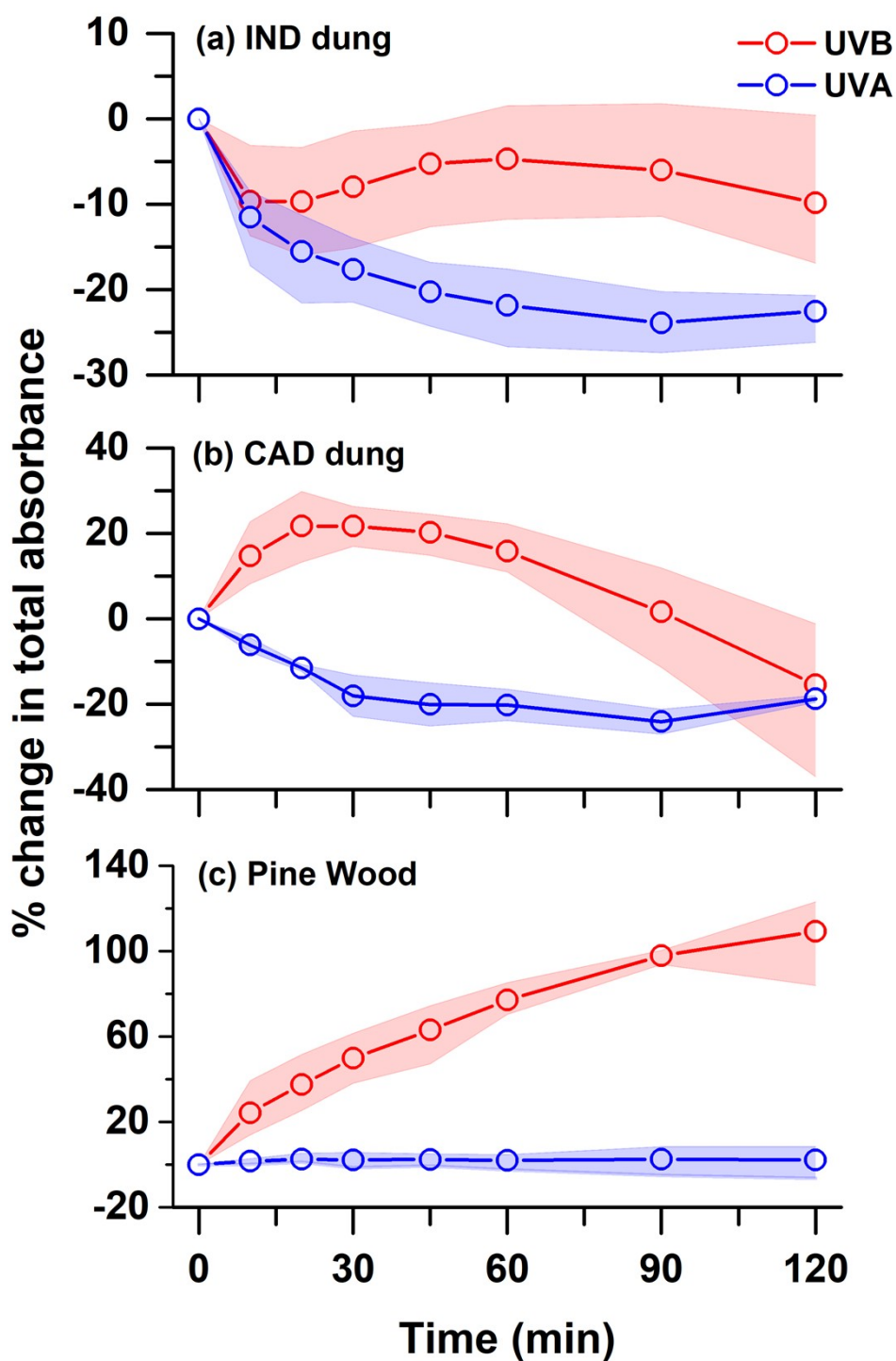


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127 **Figure S5.** Mass absorption efficiency of WS-BrC ( $MAE_{WS-BrC-365}$ ) arising from dung cakes

128 (from India (IND) and Canada (CAD)) and pine wood burning at: (a) 365 nm and (b) 400 nm.

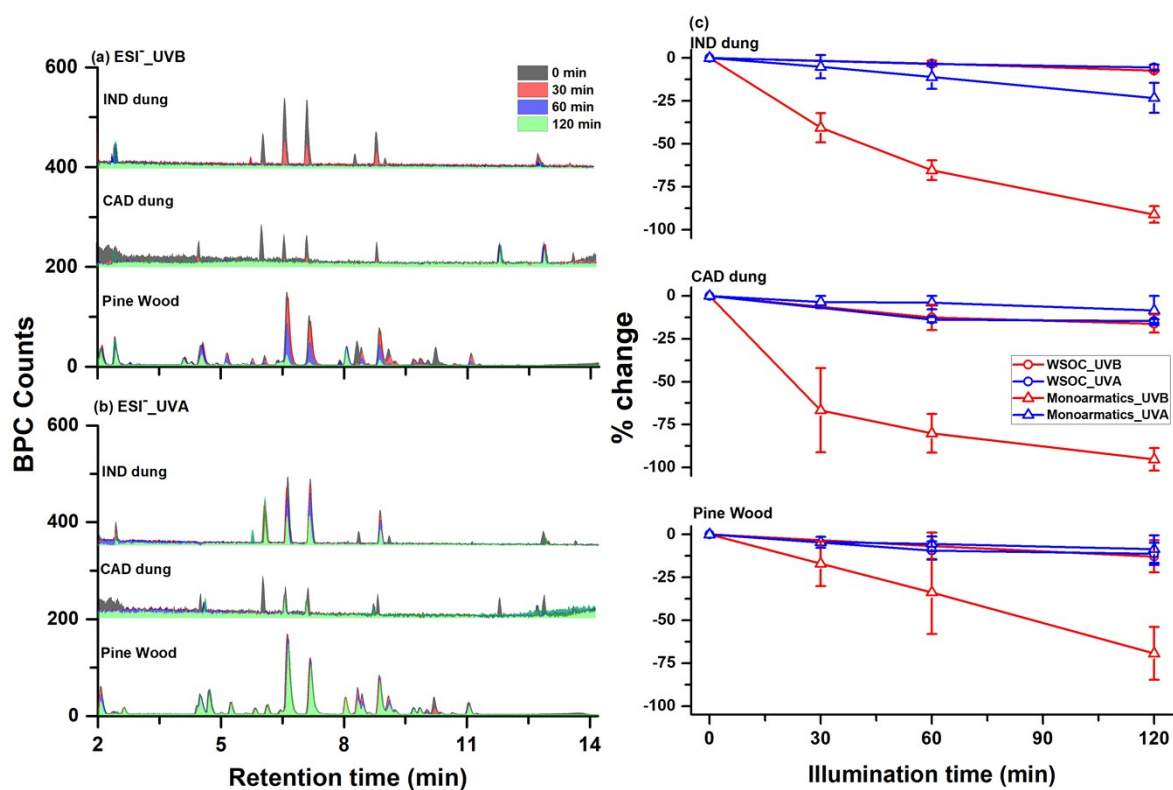
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131 **Figure S6.** Temporal variability of % change in solar-weighted total absorbance of WS-BrC  
 132 aerosols emitted from (a) IND dung, (b) CAD dung and (c) Pine wood burning during their  
 133 direct photolysis using UVB(Q) and UVA(Q) light. Where, (Q) denotes that the experiments  
 134 were performed in a quartz vessel

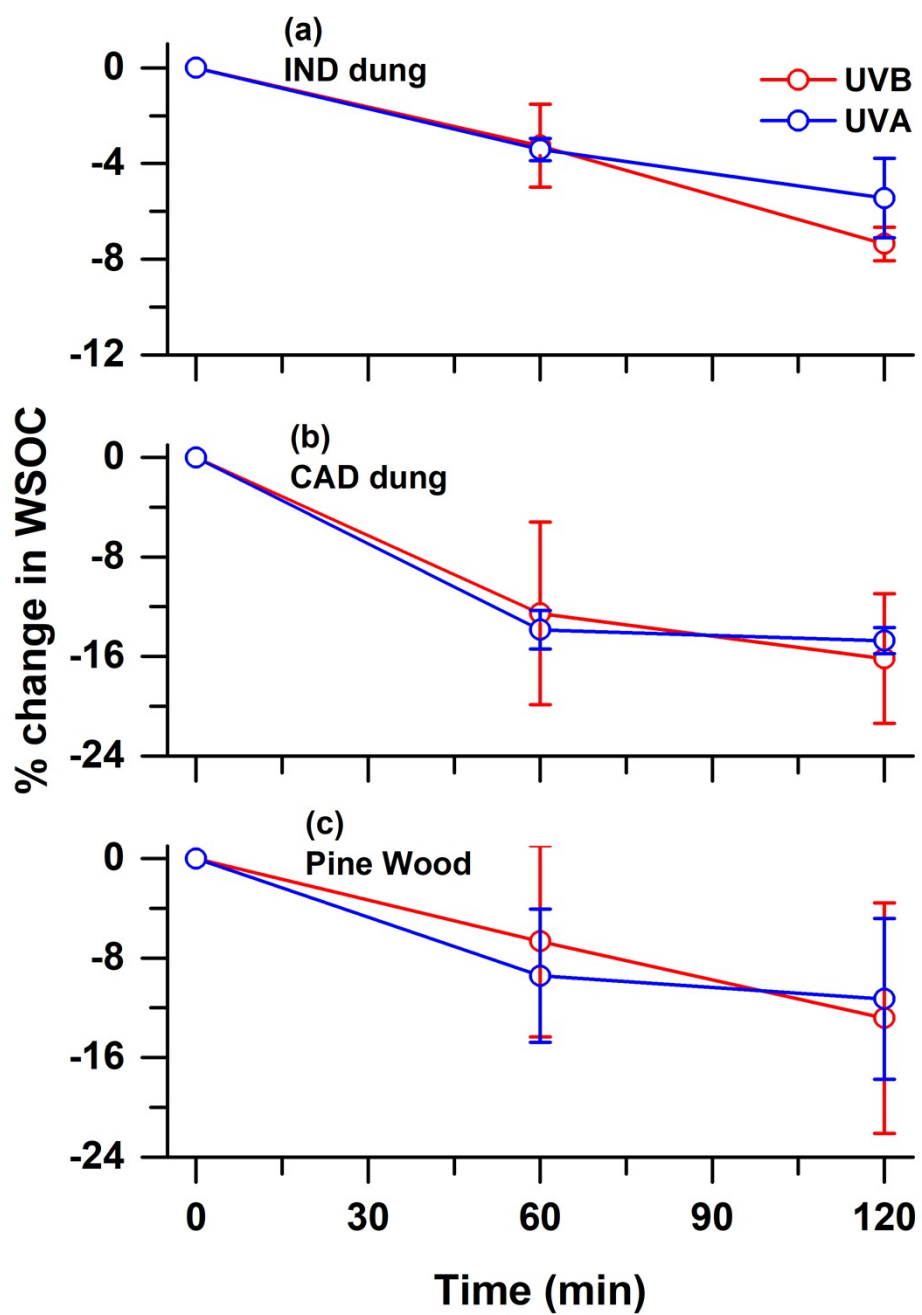
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138 **Figure S7.** Temporal variation in ESI<sup>-</sup> LC-MS BPC chromatogram of water-soluble organics  
 139 upon exposure to UVB(Q) and UVA(Q)-light during aqueous-phase direct photolysis (a, b)  
 140 and % change in net intensity of monoaromatic compounds along with WSOC concentrations  
 141 (c) for all three solid biomass fuel burning types.

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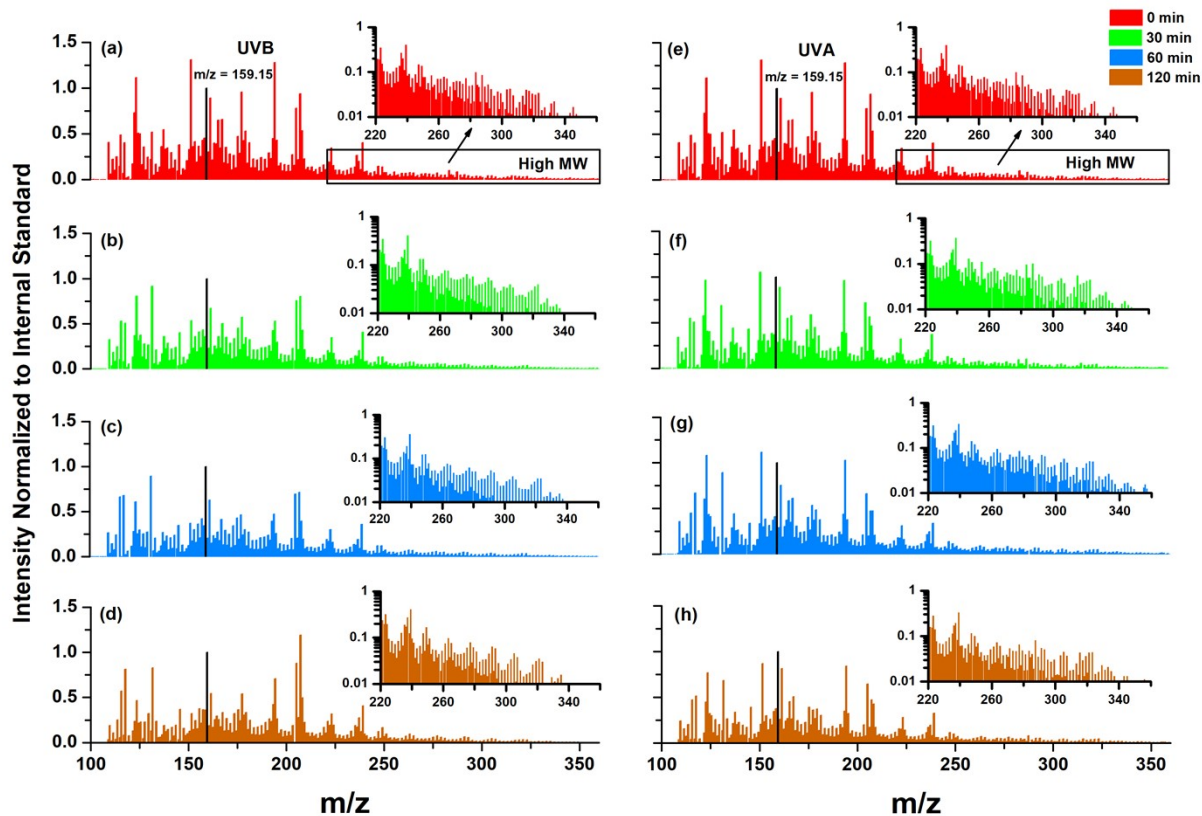


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144 **Figure S8.** Temporal variation of % change in WSOC concentration during the direct  
 145 photolysis using UVB(Q) and UVA(Q) light for (a) IND dung, (b) CAD dung and (c) Pine  
 146 wood burning.

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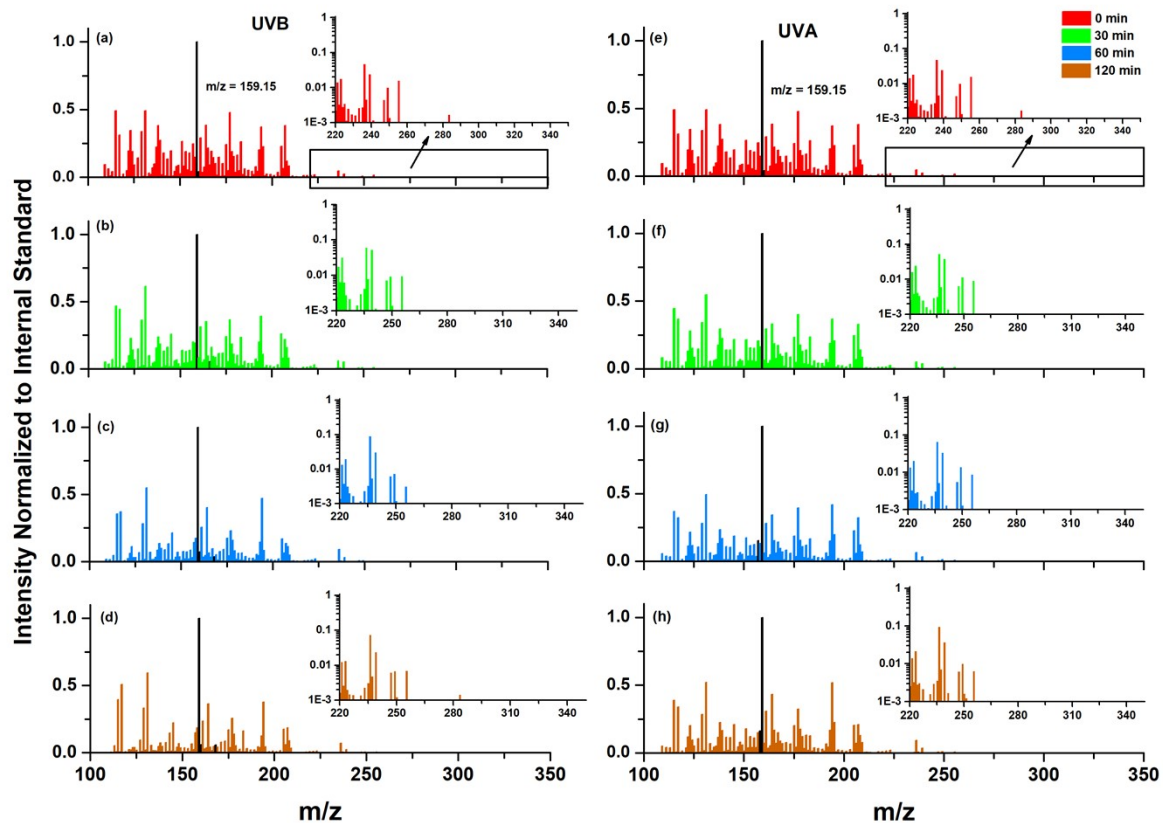
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150 **Figure S9.** Internal standard normalized direct flow-injection MS injection spectra at 0, 30, 60  
 151 and 120 min for pine wood burning during UVB(Q) and UVA(Q) direct photolysis  
 152 experiments. Here, black bar is normalized signal for internal standard (*i.e.*, pimelic acid,  $m/z$   
 153 = 159.15).

154



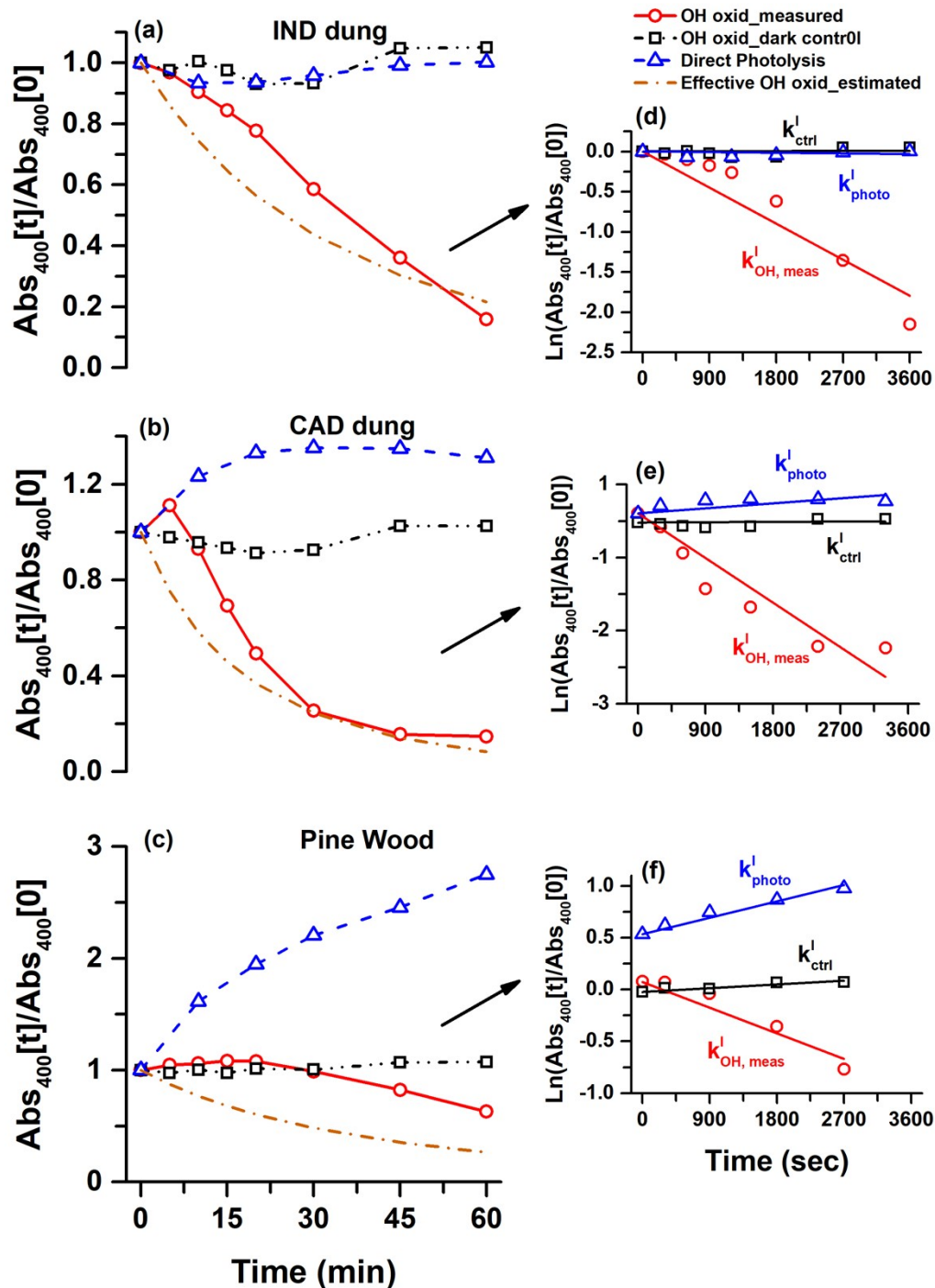
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156 **Figure S10.** Internal standard normalized direct flow-injection MS injection spectra at 0, 30,  
 157 60 and 120 min for IND dung burning samples during UVB(Q) and UVA(Q) aqueous-phase  
 158 direct photolysis experiments. Here, black bar is normalized signal for internal standard (*i.e.*,  
 159 pimelic acid,  $m/z = 159.15$ ).

160

161





162

163 **Figure S11.** Temporal profiles of absorbance at 400 nm normalized to initial value  
 164 ( $Abs_{400}[t]/Abs_{400}[0]$ ) during aqueous-phase OH oxidation (red), dark control (black) and direct  
 165 photolysis (blue) experiments for WS-BrC emitted from IND dung (a, d), CAD dung (b, e) and  
 166 pine wood burning (c, f). The plots (a), (b) and (c) indicate photochemical decay/enhancement  
 167 profile of  $Abs_{400}[t]/Abs_{400}[0]$ , whereas (d), (e), and (f) indicate corresponding first-order

168 decay/enhancement plots for WS-BrC. The maximum absorbance datapoint was considered as  
169 origin for determination of first order-decay constant. Moreover, the brown color dashed lines  
170 represent effective OH oxidation

171

172 **Table S1.** High-resolution ESI- LC-MS analysis of mixture of five known standard  
173 compounds.

<b>Compound Name</b>	<b>Chemical formula</b>	<b>Calculated Mass to Charge (m/z)*</b>	<b>Intensity</b>
Vanillin	C <sub>8</sub> H <sub>8</sub> O <sub>3</sub>	151.04	1.66×10 <sup>4</sup>
Sinapaldehyde	C <sub>11</sub> H <sub>12</sub> O <sub>4</sub>	207.07	4.24×10 <sup>4</sup>
Coniferyl aldehyde	C <sub>10</sub> H <sub>10</sub> O <sub>3</sub>	177.06	6.88×10 <sup>4</sup>
4-Nitroguaiacol	C <sub>7</sub> H <sub>7</sub> NO <sub>4</sub>	168.03	3.20×10 <sup>5</sup>
4-Nitrocatechol	C <sub>6</sub> H <sub>5</sub> NO <sub>4</sub>	154.01	5.16×10 <sup>5</sup>

174 All these standard compounds are commonly observed in biomass burning smoke.<sup>9</sup>

175 \*Assigned by MassHunter software

176

177

178 **Table S2.** WSOC and TN concentrations for different types of solid biomass fuel burning.

Fuel type	Parameters*			
	WSOC (mg-C/L)	TN (mg-N/L)	TIN <sup>§</sup> (µg-N/L)	WSON/WSOC <sup>&amp;</sup>
Blanks <sup>a, c</sup>	3.8 ± 0.1	0.37 ± 0.07	< LOD	-
IND dung <sup>#, b, d</sup>	139.7 ± 30.2	10.65 ± 2.05	< LOD	0.077 ± 0.004
CAD dung <sup>#, a, e</sup>	34.2 ± 4.5	3.90 ± 1.55	-	0.115 ± 0.047
Pine Wood <sup>#, b, e</sup>	403.6 ± 300.4	5.85 ± 4.44	<LOD	0.015 ± 0.006

179 \* Mean ± standard deviation; # all values are blank corrected; WSOC = water-soluble organic  
 180 carbon; TN = total nitrogen (water-soluble); TIN = total inorganic nitrogen (water-soluble);  
 181 WSON = water-soluble organic nitrogen

182 <sup>§</sup> TIN =  $(NO_2^- - N + NO_3^- - N)$ . Measured using colorimetric autoanalyzer (ThermoFisher  
 183 Gallery Beermaster Plus). Limit of detection (LOD) = 4.3 µg/L

184 <sup>&</sup> Since, inorganic nitrogen constituents (e.g.,  $NO_2^-$ ,  $NO_3^-$ ) are negligible compared to TN.  
 185 Therefore, TN/WSOC ratio (i.e., organic + inorganic fraction) would likely to reflect more  
 186 closely with water-soluble organic nitrogen fraction (i.e., WSON/WSOC) than inorganic  
 187 fraction.

188 <sup>a</sup> extracted in 25 mL; <sup>b</sup> extracted in 30 mL (diluted by 8 times during WSOC, TN and TIN  
 189 measurements)

190 <sup>c</sup> n = 2; <sup>d</sup> n = 3; <sup>e</sup> n = 4

191

192 **Table S3.** Molecular composition details of monoaromatic compounds found in water-extracts  
 193 of dung cakes and pine wood burning samples using high-resolution ESI- LC-MS analysis.

Calculated Mass to Charge (m/z)	Molecular Formula	Possible Identity	IND	CAD	Wood
109.0296	C <sub>6</sub> H <sub>6</sub> O <sub>2</sub>	Catechol	Y	Y	Y
121.0298	C <sub>7</sub> H <sub>6</sub> O <sub>2</sub>	Benzoic acid	Y	N	Y
123.0453	C <sub>7</sub> H <sub>8</sub> O <sub>2</sub>	Guaiacol	Y	Y	Y
125.0246	C <sub>6</sub> H <sub>6</sub> O <sub>3</sub>	Isomaltol	Y	N	Y
135.0454	C <sub>8</sub> H <sub>8</sub> O <sub>2</sub>	-	Y	Y	Y
137.0609	C <sub>8</sub> H <sub>10</sub> O <sub>2</sub>	1, 2-Dimethoxybenzene	Y	N	Y
149.0610	C <sub>9</sub> H <sub>10</sub> O <sub>2</sub>	4-Ethylbenzoic acid	Y	N	Y
151.0401	C <sub>8</sub> H <sub>8</sub> O <sub>3</sub>	Vanillin	Y	Y	Y
151.0767	C <sub>9</sub> H <sub>12</sub> O <sub>2</sub>	-	Y	N	Y
155.0351	C <sub>7</sub> H <sub>8</sub> O <sub>4</sub>	-	Y	N	Y
166.0514	C <sub>8</sub> H <sub>9</sub> NO <sub>3</sub>	4-Nitrophenethyl alcohol	Y	N	N
177.0559	C <sub>10</sub> H <sub>10</sub> O <sub>3</sub>	Coniferyl aldehyde	Y	Y	Y
181.0515	C <sub>9</sub> H <sub>10</sub> O <sub>4</sub>	Syringaldehyde	Y	N	Y
195.0667	C <sub>10</sub> H <sub>12</sub> O <sub>4</sub>	Trimethoxy benzaldehyde	Y	N	N

194 Where, Y indicates that specie was identified in ESI- LC-MS spectra; N indicates that specie  
 195 couldn't identified.

196

197 **Table S4.** Temporal variation in ratio of internal standard normalized total intensity of high  
198 molecular-weight compounds to low molecular-weight compounds during aqueous-phase  
199 UVB(Q) and UVA(Q) direct photolysis of pine wood burning water-soluble organics.

Exposure time (min)	Pine wood	
	UVB*	UVA*
0	0.141 ± 0.078	0.189 ± 0.092
30	0.151 ± 0.091	0.156 ± 0.081
60	0.155 ± 0.090	0.169 ± 0.079
120	0.172 ± 0.103	0.166 ± 0.069

200 Where, m/z = 215 to 350 were considered high molecular-weight compounds (dimer range);  
201 m/z = 100 to 215 were taken as low molecular-weight compounds (monomer range); 20µM  
202 pimelic acid was used as internal standard.

203 \* Results are average of direct flow-injection MS analysis of two pine wood burning filter  
204 water-extracts

205

206

207 **Table S5.** Comparison of average intensity of a few monoaromatics compounds in water-  
208 extracts of IND dung and pine wood burning samples.

<b>Compound Name</b>	<b>Intensity*</b>	
	<b>IND dung</b>	<b>Pine wood</b>
Guaiacol	127.4	201.2
Vanillin	19.3	28.0
Coniferyl aldehyde	16.0	54.3

209 \* Normalized to average WSOC concentration of respective solid biomass fuel burning

210

211 **Table S6.** Aqueous-Phase second-order rate constants and estimated cloud-water half-life of  
 212 WS-BrC emitted from pyrolysis of different solid biomass fuels, calculated from absorbance  
 213 at 400 nm.

Origin/Type	$k_{OH, meas}^{II}$ ( $M^{-1} s^{-1}; \times 10^8$ )	$k_{OH, eff}^{II}$ § ( $M^{-1} s^{-1}; \times 10^8$ )	$\tau_{1/2, cloud-water}^{\#}$ (min)
IND dung	29.1 ± 3.7	28.4 ± 1.4	59 ± 3
CAD dung	19.0 ± 14.8	20.2 ± 15.2	138 ± 83
Pine Wood	15.9 ± 4.3	40.4 ± 12.5	45 ± 14

214 & Corrected for dark ( $H_2O_2$ ) control.

215 § Corrected for contribution due to dark control and UVB direct photolysis.

216 # Estimated corresponding  $k_{OH, eff}^{II}$  and by assuming an ambient cloud water with [OH]<sub>ss</sub> of 1  
 217  $\times 10^{-13}M$ , representing upper band of OH in remote cloud-waters.<sup>10</sup>  
 218



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