

1 **Supporting Information for**

2

3 **Health Implications of Wintertime Fine Particulate Matter from**

4 **Southwestern China**

5

6

7 Jinyitao Wang^{a,1}, Fang Zhou^{a,1}, Wei Zhang^a, Xinquan Zhao^b, Steven J. Campbell^c, Li Zhou^d,
8 Jialiang Feng^a, Qingyan Fu^e, Arthur W.H. Chan^f, Fumo Yang^d, Mi Tian^{b,*}, Shunyao Wang^{a,*}

9

10

11 ^aSchool of Environmental and Chemical Engineering, Shanghai University, Shanghai, 200444,
12 China

13

14 ^b Key Laboratory of Three Gorges Reservoir Region's Eco-Environment, Ministry of
15 Education, Chongqing University, Chongqing, 400044, China

16

17 ^cMRC Centre for Environment and Health, Environmental Research Group, Imperial College
18 London, 86 Wood Lane, London W12 0BZ, United Kingdom

19

20 ^dNational Engineering Research Center for Flue Gas Desulfurization, Department of
21 Environmental Science and Engineering, Sichuan University, Chengdu, 610065, China

22

23 ^eShanghai Academy of Environmental Sciences, Shanghai 200233, China

24

25 ^fDepartment of Chemical Engineering and Applied Chemistry, University of Toronto, Toronto,
26 Ontario M5S 3E5, Canada

27

28 ¹These authors contributed equally to this work.

29

30

31 **Correspondence to**

32 Mi Tian

33 tianmi628@cqu.edu.cn

34

35 Shunyao Wang

36 syw@shu.edu.cn

37

38 **This file includes**

39 *Number of pages:* 28

40 *Number of Table:* 5

41 *Number of Figures:* 10

42 **Supplemental Text**

43

44 **S1. Analysis by ion chromatography**

45 For the analysis of water-soluble inorganic ions, a quarter of each quartz filter was first
46 extracted using ultrapure water in an ultrasonic bath for 30 min, and then filtered
47 through a syringe filter (0.45 µm pore size). Both anions (SO_4^{2-} , NO_3^- and Cl^-) and
48 cations (Na^+ , NH_4^+ , K^+ , Mg^{2+} and Ca^{2+}) were determined using ion chromatograph
49 (Dionex, Dionex 600, USA). Anions were separated using an AS11-HC column with
50 30 mM KOH as an eluent at a flow rate of 1.0 mL min⁻¹. Cations were determined using
51 a CS12A column with 20 mM MSA (methanesulfonic acid) at a flow rate of 1.0 mL
52 min⁻¹. Individual standard solutions of all investigated ions (1000 mg L⁻¹) were diluted
53 in series to construct the calibration curves. The correlation coefficients of the linear
54 regression of the standard curves were all above 0.999. Field blanks were prepared and
55 analyzed together with the samples, and then subtracted from the samples. The
56 concentrations of water-soluble inorganic ions in field blanks were in the range of
57 0.008–0.13 µg m⁻³. The relative standard deviation of each ion was measured to be
58 within 8% for all replicates.

59

60 **S2. OC/EC measurement**

61 Organic carbon (OC) and elemental carbon (EC) were analyzed using a multi-
62 wavelength Carbon Analyzer (Model 2015, DRI, USA) in accordance with the
63 Interagency Monitoring of Protected Visual Environment (IMPROVE-A) protocol ¹⁻³.
64 Four OC fractions (i.e. OC₁, OC₂, OC₃, and OC₄) were measured at the temperature of
65 140 °C, 280 °C, 480 °C, and 580 °C, respectively, in a helium buffer gas. During each
66 run, three EC fractions (i.e. EC₁, EC₂, and EC₃) were measured at the temperature of
67 580 °C, 740 °C, and 840 °C, respectively.

68

69 **S3. Health risk of metallic elements**

70 Hazard quotient (HQ) is used to measure the non-carcinogenic health risks of heavy
71 metals in ambient PM (Eqn.S1). ADD represents the average daily exposure dose of
72 heavy metals through different exposure pathways (mg kg⁻¹day⁻¹). Detailed formulas
73 and parameterization of ADD calculation regarding the three different exposure

74 pathways (ingestion, inhalation, and dermal contact) can be found in Text S3. Based on
 75 the calculated ADD, HQ can be calculated by dividing the ADD by the specific
 76 reference dose (RfD). The reference dose values (Table S3) used in this study are based
 77 on previous literature^{34,35}.

$$78 \quad HQ = \frac{ADD}{RfD} \quad (S1)$$

79 Note that the reference dose is an approximation of maximum risk through daily
 80 exposure taken into account during the lifetime³⁴. If $HQ < 1$ (the average daily dose is
 81 less than the reference dose), it means that the adverse health effects are negligible.
 82 Otherwise, if $HQ > 1$ (the average daily dose is greater than the reference dose), the
 83 specific exposure pathway may have significant adverse effects on human health³⁷.

84

85 Excess cancer risk (ECR) can be estimated as the incremental probability that an
 86 individual develops cancer over a lifetime due to total exposure to potential
 87 carcinogens, which can be calculated using Eqn.S2³³.

$$88 \quad ECR = \frac{C \times ET \times EF \times ED \times IUR}{AT} \quad (S2)$$

89 Where C is the pollutant concentration ($\mu\text{g m}^{-3}$); ET is the exposure time (8 hours day⁻¹); EF is the exposure frequency (180 days year⁻¹); ED is the duration of exposure,
 90 which in this study was 24 years (adults) and 6 years (children) in this study; IUR is the
 91 inhalation unit risk ($\mu\text{g m}^{-3}$)⁻¹; AT is the average exposure time of the specific
 92 carcinogen (70 years \times 365 days year⁻¹ \times 24 hours day⁻¹). It should be noted that only
 93 the carcinogenic risk of metals via inhalation was considered in this study.

95

96 Average daily exposure dose of heavy metals through ingestion (ADD_{ing} , $\text{mg kg}^{-1}\text{day}^{-1}$)

$$97 \quad ADD_{ing} = \frac{C \times IngR \times EF \times ED \times CF}{BW \times AT} \quad (S3)$$

98 Average daily exposure dose of heavy metals through inhalation (ADD_{inh} , mg
 99 $\text{kg}^{-1}\text{day}^{-1}$)

$$100 \quad ADD_{inh} = \frac{C \times IngR \times EF \times ED}{BW \times AT \times PEF} \quad (S4)$$

101 Average daily exposure dose of heavy metals through dermal contact (ADD_{derm}, mg
102 kg⁻¹day⁻¹)

$$103 \quad ADD_{derm} = \frac{C \times SA \times AF \times EV \times ABS \times EF \times ED \times CF}{BW \times AT} \quad (S5)$$

104 Where C is the concentration of metal in PM (mg kg⁻¹); IngR is the intake rate (mg day⁻¹), which was 30 mg day⁻¹ for adults and 60 mg day⁻¹ for children in this study. InhR is
105 the inhalation rate (m³ day⁻¹), which in this study was 7.63 (adults) and 20 (children);
106 EF is the exposure frequency, 180 days year⁻¹ in this study; ED is the duration of
107 exposure, which in this study was 24 years (adults) and 6 years (children); BW is body
108 weight. In this study, adults were 70 kg and children were 15 kg. SA is the skin surface
109 area parameter, in this study, adults 5700 cm² and children 2800 cm²; AF is the adhesion
110 factor of soil to skin, which in this study was 0.07 mg cm⁻² event⁻¹ (adults) and 0.2 mg
111 cm⁻² event⁻¹ (children). EV is the event frequency, which is 1 event day⁻¹; ABS is skin
112 absorption fraction 0.001 in this study; PEF is the particle emission factor of 1.36 × 10⁹
113 m³ kg⁻¹. The average time of non-carcinogens AT was ED × 365 days year⁻¹ in this
114 study; CF is a conversion factor of about 10⁻⁶ kg mg⁻¹. All the parameters used in the
115 calculation can be found in EPA report ⁴. Reference concentration values (RfC, mg m⁻³) and
116 inhalation unit risks (IUR, µg⁻¹ m³) were used to calculate the hazard quotient
117 (HQ) and excess carcinogenic risk ^{5, 6}.

119

120 As illustrated in Figure S4a, the HQ values of Chongqing wintertime PM_{2.5} for adults
121 were found to be higher than those estimated for children. Most of the HQ values are
122 less than 1, while HQ_{As} and HQ_{Pb} are greater than 1. Previous work has pointed out that
123 the non-carcinogenic risk of heavy metals should not be considered as negligible when
124 HQ > 1 ⁷. Therefore, the levels of As and Pb in Chongqing PM_{2.5} may cause significant
125 non-carcinogenic adverse effects. Among different exposure pathways, dermal (HQ_{der})
126 caused non-cancer risk was found to be the lowest (less than 1) while ingestion (HQ_{ing})
127 and inhalation (HQ_{inh}) caused non-cancer risk exceeded 1 for Cr, As, and Pb. Attentions
128 should also be given that the HQ_{inh} of PM_{2.5} in wintertime Chongqing (5.90 for adults
129 and 11.58 for children) was higher than those reported in other cities on the East coast

130 of China, such as Guangzhou (1.04-1.16 for adults, 1.93-2.15 for children), Hangzhou
131 (2.41 for adults, 1.32-1.43 for children), Shanghai (1.05-1.21 for adults, 1.95-2.25 for
132 children), and Beijing (1.02-1.19 for adults, 1.9-2.2 for children). The HQ_{inh} value
133 estimated for Chongqing in this study was more similar to the less developed cities
134 located on the central west part of the country, such as Xi'an (7.24-7.84 for adults, 13.4-
135 14.5 for children) and Kunming (6.91-7.11 for adults, 12.8-13.2 for children)⁷.

136

137 Cr, As, Pb, Cd are known carcinogenic metals in PM⁷. According to Figure S4b, the
138 ECR of carcinogenic metal elements follows As > Cr (VI) > Cd > Pb. Cr and As are
139 mainly derived from coal combustion⁸. It should be noted that nearly 80% coal
140 resources of the whole country is distributed in the west region⁷, explaining that Cr
141 (VI) and As emerged as the two heavy metals with the highest excessive carcinogenic
142 risk in Chongqing. Overall, the observed ECR values for metallic composition in
143 Chongqing are comparable with those observed in other Asian countries. For example,
144 the ECR values of Cd, Cr (VI), and Cd in PM_{2.5} collected Indian were 1.90 × 10⁻⁶, 1.01
145 × 10⁻⁴, and 0.40 × 10⁻⁶⁹. In Ulsan from South Korea, the ECR values for Cd and Cr (VI)
146 were reported to be 8.7 × 10⁻⁶ and 1.74 × 10⁻⁵, respectively¹⁰. More efforts should be
147 paid in the future for regulating metallic airborne pollution in the southwestern area of
148 China.

149

150 **S4. Multilinear regression results**

151 The unstandardized regressions with different independent variables were shown as
152 following. The first equation in each group in italics is the regression included in the
153 main text of the paper (S4, S15, S39).

154

155 For OP_v^{DTT},

156 $OP_v^{DTT} = 0.10OC + 7.14Cu + 1.69Zn + 1.79PAHs + 8.67OPAHs + 0.52$

157 $r^2 = 0.84$ (S4)

158 $OP_v^{DTT} = 0.53PAHs + 14.58Cu + 0.097$ $r^2 = 0.62$ (S5)

159	$OP_v^{DTT} = 0.47OPAHs + 16.68Cu + 0.084$	$r^2 = 0.61$ (S6)
160	$OP_v^{DTT} = 9.84NPAHs + 1.57Cu + 0.006$	$r^2 = 0.62$ (S7)
161	$OP_v^{DTT} = 0.25EC + 60.41Cu + 0.89$	$r^2 = 0.73$ (S8)
162	$OP_v^{DTT} = 0.23EC + 5.66Cu + 5.56Zn + 0.11PAHs + 5.77OPAHs$	$r^2 = 0.76$ (S9)
163	$OP_v^{DTT} = 0.096OC + 18.57Mn + 0.56$	$r^2 = 0.86$ (S10)
164	$OP_v^{DTT} = 0.99OC + 21.58Fe + 4.74$	$r^2 = 0.86$ (S11)
165	$OP_v^{DTT} = 0.25EC + 2.40Fe + 0.79$	$r^2 = 0.78$ (S12)
166	$OP_v^{DTT} = 0.108OC + 14.99OPAHs + 0.482$	$r^2 = 0.84$ (S13)
167	$OP_v^{DTT} = 0.25EC + 6.72Zn + 0.79$	$r^2 = 0.76$ (S14)
168		
169	For OP_v^{EPR} ,	
170	$OP_v^{EPR} = 0.001EC + 0.007Fe + 0.088Cu - 0.63PAHs + 14.63NPAHs - 0.001$	$r^2 = 0.73$ (S15)
171		
172	$OP_v^{EPR} = 10.2NPAHs + 0.44Cu + 0.01$	$r^2 = 0.59$ (S16)
173	$OP_v^{EPR} = 0.17OPAHs - 0.4PAHs + 3.88Cu + 0.41Mn + 0.02$	$r^2 = 0.62$ (S17)
174	$OP_v^{EPR} = 0.0003OC + 13.301NPAHs$	$r^2 = 0.70$ (S18)
175	$OP_v^{EPR} = 0.0003OC + 0.317Cu$	$r^2 = 0.60$ (S19)
176	$OP_v^{EPR} = 0.0004OC + 0.511Cu - 0.036Zn$	$r^2 = 0.65$ (S20)
177		
178	For <i>in vitro ROS_V</i> ,	
179	<i>in vitro ROS_V</i> = $1.19Fe + 99.52Cu + 807.84NPAH - 47.17Cd - 74.57Ti + 0.36$	
180		$r^2 = 0.68$ (S21)

181 *in vitro* $ROS_V = 0.15EC + 169.33NPAH + 3.38Zn - 14.61Pb + 0.19$
 182 $r^2 = 0.54$ (S22)

183 *in vitro* $ROS_V = 86.91NPAH + 11.09Cu - 4.29Cd - 5.22Ti + 0.028$
 184 $r^2 = 0.67$ (S23)

185 *in vitro* $ROS_V = 0.018EC + 174.45NPAH + 0.34Zn + 32.04As - 49.94Cd - 22.12Pb + 0.023$
 186 $r^2 = 0.67$ (S24)

187 *in vitro* $ROS_V = 0.19EC + 185.76NPAH + 3.54Zn - 53.01Cd - 18.12Pb + 0.21$
 188 $r^2 = 0.61$ (S25)

189 *in vitro* $ROS_V = 8.68OPAH + 0.11EC - 46.09Cd + 0.45$ $r^2 = 0.50$ (S26)

190 *in vitro* $ROS_V = 0.98EC + 79.21OPAH + 4.64$ $r^2 = 0.44$ (S27)

191

192 The standardized regressions with different independent variables were shown as
 193 following. The first equation in each group in italics is the regression included in the
 194 main text (S21, S32).

195

196 For OP_v^{DTT} ,

197 $OP_v^{DTT} = 0.697OC + 0.033Cu + 0.081Zn + 0.63PAHs + 0.141OPAHs$
 198 $r^2 = 0.84$ (S28)

199 $OP_v^{DTT} = 0.18PAHs + 0.63Cu - 1.4988E-15$ $r^2 = 0.61$ (S29)

200 $OP_v^{DTT} = 0.071OPAHs + 0.72Cu - 1.507E-15$ $r^2 = 0.61$ (S30)

201 $OP_v^{DTT} = 0.15NPAHs + 0.68Cu - 1.3862E-15$ $r^2 = 0.62$ (S31)

202 $OP_v^{DTT} = 0.129PAHs + 0.127NPAHs + 0.59Cu - 1.3738E-15$ $r^2 = 0.62$ (S32)

203 $OP_v^{DTT} = 0.271OC + 0.065PAHs + 0.189NPAHs + 0.093Cu + 0.166Mn + 0.231Zn$
 204 $r^2 = 0.71$ (S33)

205 $OP_v^{DTT} = 0.683OC + 0.332Mn$ $r^2 = 0.86$ (S34)

206 $OP_v^{DTT} = 0.701OC + 0.317Fe$ $r^2 = 0.86$ (S35)

- 207 $OP_v^{DTT}=0.62EC + 0.352Fe$ $r^2= 0.78$ (S36)
- 208 $OP_v^{DTT}=0.765OC + 0.243OPAHs$ $r^2= 0.84$ (S37)
- 209 $OP_v^{DTT}=0.609EC + 0.325Zn$ $r^2= 0.76$ (S38)
- 210
- 211 For OP_v^{EPR} ,
- 212 $OP_v^{EPR}=0.585EC + 0.242Fe + 0.091Cu -0.492PAHs +0.488NPAHs$ $r^2=0.73$ (S39)
- 213
- 214 $OP_v^{EPR}=0.524NPAHs + 0.475EC + 0.07Cu +0.026$ $r^2=0.68$ (S40)
- 215 $OP_v^{EPR}=0.069PAHs + 0.523Fe + -8.9743E-17$ $r^2= 0.34$ (S41)
- 216 $OP_v^{EPR}=0.288EC + 0.015Mn + 0.478Cu +0.016$ $r^2=0.54$ (S42)
- 217 $OP_v^{EPR}= 0.525OC + 0.443NPAHs$ $r^2=0.70$ (S43)
- 218 $OP_v^{EPR}= 0.487OC + 0.327Cu$ $r^2=0.60$ (S44)
- 219 $OP_v^{EPR}= 0.631OC + 0.528Cu - 0.389Zn$ $r^2=0.65$ (S45)
- 220
- 221 For *in vitro ROS_v*,
- 222 *in vitro ROS_V*= $0.344Fe+ 0.903Cu + 0.235NPAH -0.250Cd- 0.753Ti$ $r^2=0.68$ (S46)
- 223
- 224 *in vitro ROS_V*= $0.750EC+ 0.493NPAH + 0.323Zn -0.701Pb$ $r^2=0.54$ (S47)
- 225
- 226 *in vitro ROS_V*= $0.253NPAH + 1.007Cu -0.227Cd- 0.528Ti$ $r^2=0.67$ (S48)
- 227
- 228 *in vitro ROS_V*= $0.881EC+0.507NPAH + 0.331Zn+0.312As -0.264Cd- 1.062Pb$ $r^2=0.67$ (S49)
- 229
- 230 *in vitro ROS_V*= $0.924EC+0.540NPAH + 0.338Zn -0.280Cd-0.869Pb$

231 $r^2=0.61$ (S50)
 232 *in vitro* ROS_V= 0.275OPAH + 0.523EC -0.227Cd $r^2=0.50$ (S51)
 233 *in vitro* ROS_V= 0.523EC +0.251OPAH $r^2=0.44$ (S52)
 234
 235

236 **Supplemental Tables**

237

238 **Table S1.** Comparison of OP^{DTT} values measured in this study and other studies.

Location	Year	OP_V^{DTT} (nmol min ⁻¹ m ⁻³)	Reference
Chongqing	Winter, 2015–2016	2.83±1.28	This study
Wuhan	2012	1.8–8.2	11
Hangzhou	2017	0.62 ± 0.24	12
Shanghai	2016	0.19 ± 0.04	13
Xiamen (Siming)	2015–2018	0.632 ± 0.299	14
Xiamen (Xiang'an)		0.562 ± 0.247	
Guangzhou	January, 2018	0.93 ± 0.21	15
	April, 2018	0.89 ± 0.20	
Jinzhou	2015–2016	4.4 ± 2.6	
Tianjin	2015–2016	6.8 ± 3.4	16
Yantai	2015–2016	4.2 ± 2.7	
Nanjing (Xianlin)	2020–2021	2.42	17
Nanjing (Gulou)		1.34	
Nanjing	2019–2020	2.1 ± 1.1	18
Qingdao	2021–2022	2.15 ± 1.24	19
	Spring, 2017	0.53 (0.19–0.85)	
Xi'an	Summer, 2017	0.50 (0.24–0.80)	20
	Autumn, 2017	0.40 (0.24–0.74)	
	Winter, 2017	0.64 (0.43–1.10)	
Lecce	2013–2016	0.29 ± 0.12	21
Amsterdam		0.20 ± 0.01	
Frankfurt	2012	0.28 ± 0.02	22
London		0.20 ± 0.01	
Milan		0.51 ± 0.07	

Stockholm		0.22 ± 0.02	
Thessaloniki		0.41 ± 0.02	
Southern Alps		0.11 ± 0.01	
Toronto	2016–2017	0.0306 ± 0.0107	23
Atlanta	2012–2013	0.31	24

239

240

241

242

243

244

245

246

247

248

249

250

251

252

253

254

255

256

257

258

259

260

261

262

263

264

265

266

267

268

269

270

271

272

273

274

275

276 **Table S2.** Comparison of OP^{EPR} values measured in this study and other studies.

Location	Year	OP_V^{EPR} (spins m ⁻³)	Reference
Chongqing	winter, 2015–2016	$3.3 \pm 1.55 \times 10^{13}$ ($1.24\text{--}6.59 \times 10^{13}$) $1.2 \times 10^{-2} \pm 5 \times 10^{-3}$ (nmol m ⁻³)	This study
Wanzhou, Chongqing	2017	$7.0 \pm 1.7 \times 10^{13}$	25
	Jan.–Feb., 2017	$2.58\text{--}23.47 \times 10^{14}$	26
	spring, 2017	1.7×10^{14}	
	summer, 2017	9×10^{13}	27
Xi'an	autumn, 2017	7×10^{13}	
	winter, 2017	2.1×10^{14}	
	2017	$2.58\text{--}5.41 \times 10^{14}$	28
Linfen	Before lockdown, 2019	1.3×10^{13} ($0.65\text{--}2.7 \times 10^{13}$)	29
	During lockdown, 2020	1.4×10^{13} ($0.62\text{--}5.2 \times 10^{13}$)	
Beijing	winter, 2020	non-heating and heating: 16.2×10^{12} and 14.2×10^{12} ($3.4\text{--}39.5 \times 10^{12}$)	30
Yuncheng		non-heating and heating: 12.7×10^{12} and 28.2×10^{12} ($4.8\text{--}59.2 \times 10^{12}$)	
Erenhot		1.6×10^{13}	
Zhangbei	2016	5.71×10^{13}	31
Jinan		4.57×10^{13}	
Nanjing	2019	7.61×10^{12} ($2.78 \times 10^{12}\text{--}1.72 \times 10^{13}$)	32
Mainz	2019	$3.6 \pm 3.1 \times 10^{12}$	33
Lahore, Pakistan	summer, 2019	1.78×10^{13} ($2.9 \times 10^{12}\text{--}4.6 \times 10^{13}$)	34
	winter, 2019	1.20×10^{14} ($2.9 \times 10^{13}\text{--}2.9 \times 10^{14}$)	

277
278
279
280
281
282
283
284

285 **Table S3.** Recommended values of Reference Doses (RfD) ($\text{mg kg}^{-1} \text{ day}^{-1}$).

Heavy metal	As	Cd	Cr	Cu	Ni	Pb	V	Zn
RFD _{ing}	3.00×10^{-4}	1.00×10^{-3}	3.00×10^{-3}	4.00×10^{-2}	2.00×10^{-2}	3.50×10^{-3}	7.00×10^{-3}	3.00×10^{-1}
RFD _{inh}	3.00×10^{-4}	1.00×10^{-3}	2.86×10^{-5}	4.02×10^{-2}	3.52×10^{-3}	3.52×10^{-3}	7.00×10^{-3}	3.01×10^{-1}
RFD _{der}	1.23×10^{-4}	1.00×10^{-5}	6.00×10^{-5}	1.20×10^{-2}	5.40×10^{-3}	5.25×10^{-4}	7.00×10^{-5}	6.00×10^{-1}

286 RFD_{ing}: Ingestion reference dose.

287 RFD_{inh}: Inhalation reference dose.

288 RFD_{der}: Dermal contact reference dose.

Table S4. Correlation between OP_V (OP_V^{DTT}, OP_V^{EPR} and *in vitro* ROS_V) and non-metallic PM_{2.5} components measured in Chongqing.

	OP _V ^{DTT}	OP _V ^{EPR}	ROS _V	SO ₄ ²⁻	NO ₃ ⁻	NH ₄ ⁺	Cl ⁻	K ⁺	OC	EC	SO ₂	NO ₂	O ₃	CO	PAHs	NPAHs	OPAHs	CHO	CHNO	CHOS	CHNOS
OP _V ^{DTT}	1	.716**	.582**	.800**	.782**	.803**	.710**	.820**	.890**	.838**	.400*	.784**	-.390*	.699**	.595**	.638**	0.168	.440*	0.189	0.118	1
OP _V ^{EPR}	.716**	1	.606**	.721**	.762**	.740**	.581**	.742**	.743**	.736**	.372*	.778**	-.369*	.513**	.681**	.485**	0.039	0.084	-0.101	-0.172	.716**
ROS _V	.582**	.606**	1	.585**	.595**	.611**	.655**	.650**	.564**	.634**	0.356	.639**	-.495**	.584**	.605**	.557**	0.011	0.179	-0.050	-0.046	.582**
SO ₄ ²⁻	.800**	.721**	.585**	1	.952**	.990**	.680**	.957**	.942**	.890**	.408*	.881**	-0.279	.423*	.407*	0.346	-0.092	.379*	-0.059	-0.130	.800**
NO ₃ ⁻	.782**	.762**	.595**	.952**	1	.978**	.721**	.945**	.932**	.888**	.454*	.911**	-0.326	.450*	.494**	.421*	-0.051	0.261	-0.143	-0.196	.782**
NH ₄ ⁺	.803**	.740**	.611**	.990**	.978**	1	.727**	.970**	.944**	.903**	.432*	.898**	-0.306	.455*	.446*	.401*	-0.070	0.344	-0.096	-0.149	.803**
Cl ⁻	.710**	.581**	.655**	.680**	.721**	.727**	1	.805**	.767**	.826**	.414*	.740**	-.540**	.753**	.485**	.693**	0.090	0.269	-0.032	-0.056	.710**
K ⁺	.820**	.742**	.650**	.957**	.945**	.970**	.805**	1	.955**	.945**	.466*	.910**	-.435*	.560**	.493**	.515**	-0.002	0.341	-0.056	-0.102	.820**
OC	.890**	.743**	.564**	.942**	.932**	.944**	.767**	.955**	1	.937**	.482**	.901**	-.411*	.579**	.491**	.513**	0.059	.414*	0.042	-0.078	.890**
EC	.838**	.736**	.634**	.890**	.888**	.903**	.826**	.945**	.937**	1	.471**	.869**	-.562**	.673**	.528**	.599**	0.027	0.328	-0.069	-0.095	.838**
SO ₂	.400*	.372*	0.356	.408*	.454*	.432*	.414*	.466*	.482**	.471**	1	.483**	-0.124	.561**	0.313	.579**	.413*	0.272	.405*	0.325	.400*
NO ₂	.784**	.778**	.639**	.881**	.911**	.898**	.740**	.910**	.901**	.869**	.483**	1	-.463*	.596**	.596**	.520**	0.162	0.327	-0.003	-0.070	.784**
O ₃	-.390*	-.369*	-.495**	-0.279	-0.326	-0.306	-.540**	-.435*	-.411*	-.562**	-0.124	-.463*	1	-.481**	-.518**	-.511**	-0.122	-0.089	0.035	-0.040	-.390*
CO	.699**	.513**	.584**	.423*	.450*	.455*	.753**	.560**	.579**	.673**	.561**	.596**	-.481**	1	.660**	.921**	.482**	.403*	0.345	0.349	.699**
PAHs	.595**	.681**	.605**	.407*	.494**	.446*	.485**	.493**	.491**	.528**	0.313	.596**	-.518**	.660**	1	.671**	0.251	0.233	0.138	0.189	.595**
NPAHs	.638**	.485**	.557**	0.346	.421*	.401*	.693**	.515**	.513**	.599**	.579**	.520**	-.511**	.921**	.671**	1	.566**	0.355	.398*	.428*	.638**
OPAHs	0.168	0.039	0.011	-0.092	-0.051	-0.070	0.090	-0.002	0.059	0.027	.413*	0.162	-0.122	.482**	0.251	.566**	1	0.275	.709**	.727**	0.168
CHO	.440*	0.084	0.179	.379*	0.261	0.344	0.269	0.341	.414*	0.328	0.272	0.327	-0.089	.403*	0.233	0.355	0.275	1	.571**	0.308	.440*
CHNO	0.189	-0.101	-0.050	-0.059	-0.143	-0.096	-0.032	-0.056	0.042	-0.069	.405*	-0.003	0.035	0.345	0.138	.398*	.709**	.571**	1	.801**	0.189
CHOS	0.118	-0.172	-0.046	-0.130	-0.196	-0.149	-0.056	-0.102	-0.078	-0.095	0.325	-0.070	-0.040	0.349	0.189	.428*	.727**	0.308	.801**	1	0.118
CHNOS	1	.716**	.582**	.800**	.782**	.803**	.710**	.820**	.890**	.838**	.400*	.784**	-.390*	.699**	.595**	.638**	0.168	.440*	0.189	0.118	1

**. At level 0.01, the correlation was significant. *. At level 0.05, the correlation was significant.

Table S5. Correlation between OP_V (OP_V^{DTT}, OP_V^{EPR} and *in vitro* ROS_V) and metallic PM_{2.5} compositions measured in Chongqing.

OP _V ^{DTT}	OP _V ^{EPR}	ROS _V	Ca	Sc	Ti	V	Cr	Mn	Fe	Co	Ni	Cu	Zn	Ga	As	Se	Br	Sr	Ba	Pb	Cd	
OP _V ^{DTT}	1	.716**	0.177	0.148	-.550**	.619**	0.318	0.282	.741**	.699**	-.467**	0.282	.776**	.750**	.841**	.633**	.802**	.721**	.482**	.461*	.863**	0.259
OP _V ^{EPR}	.716**	1	.442*	0.231	-.479**	.569**	0.327	0.263	.606**	.581**	-0.299	0.345	.708**	.495**	.689**	.667**	.648**	.433*	0.349	0.357	.685**	0.154
ROS _V	0.177	.442*	1	0.182	-0.023	0.314	-0.013	0.227	0.218	0.231	0.016	0.294	0.201	0.133	0.174	0.249	0.223	0.013	0.034	-0.023	0.143	-0.200
Ca	0.148	0.231	0.182	1	-.537**	.693**	0.019	0.346	0.312	.434*	-0.050	0.265	0.276	0.190	0.093	0.039	0.130	0.231	.639**	0.090	0.113	0.013
Sc	-.550**	-.479**	-0.023	-.537**	1	-.639**	0.049	-0.253	-.471**	-.521**	0.114	-0.303	-.522**	-.505**	-.493**	-.414*	-.613**	-.517**	-.696**	-0.281	-.515**	0.107
Ti	.619**	.569**	0.314	.693**	-.639**	1	0.304	.649**	.860**	.923**	-0.343	.616**	.781**	.607**	.658**	.492**	.619**	.675**	.617**	0.153	.566**	0.089
V	0.318	0.327	-0.013	0.019	0.049	0.304	1	0.175	.402*	.393*	-0.139	0.166	.394*	0.154	.375*	0.299	.414*	0.222	0.015	0.079	0.325	.642**
Cr	0.282	0.263	0.227	0.346	-0.253	.649**	0.175	1	.702**	.754**	-.420*	.811**	.473**	0.351	0.344	.369*	0.197	.550**	0.252	0.078	0.254	0.091
Mn	.741**	.606**	0.218	0.312	-.471**	.860**	.402*	.702**	1	.976**	-.508**	.667**	.865**	.717**	.804**	.648**	.643**	.790**	.396*	0.258	.676**	0.184
Fe	.699**	.581**	0.231	.434*	-.521**	.923**	.393*	.754**	.976**	1	-.471**	.685**	.857**	.708**	.760**	.620**	.647**	.784**	.495**	0.254	.633**	0.178
Co	-.467**	-0.299	0.016	-0.050	0.114	-0.343	-0.139	-.420*	-.508**	-.471**	1	-.372*	-.437*	-.407*	-.414*	-.396*	-.253	-.447*	-0.021	-0.153	-.438*	-0.060
Ni	0.282	0.345	0.294	0.265	-0.303	.616**	0.166	.811**	.667**	.685**	-.372*	1	.573**	.382*	.470**	.469**	0.349	.545**	0.268	0.027	0.345	0.031
Cu	.776**	.708**	0.201	0.276	-.522**	.781**	.394*	.473**	.865**	.857**	-.437*	.573**	1	.781**	.843**	.718**	.755**	.722**	.498**	0.348	.779**	0.214
Zn	.750**	.495**	0.133	0.190	-.505**	.607**	0.154	0.351	.717**	.708**	-.407*	.382*	.781**	1	.857**	.673**	.675**	.685**	.636**	.527**	.778**	0.146
Ga	.841**	.689**	0.174	0.093	-.493**	.658**	.375*	0.344	.804**	.760**	-.414*	.470**	.843**	.857**	1	.816**	.816**	.743**	.478**	0.330	.917**	0.251
As	.633**	.667**	0.249	0.039	-.414*	.492**	0.299	.369*	.648**	.620**	-.396*	.469**	.718**	.673**	.816**	1	.703**	.648**	0.311	0.249	.782**	0.097
Se	.802**	.648**	0.223	0.130	-.613**	.619**	.414*	0.197	.643**	.647**	-0.253	0.349	.755**	.675**	.816**	.703**	1	.632**	.479**	0.244	.760**	0.213
Br	.721**	.433*	0.013	0.231	-.517**	.675**	0.222	.550**	.790**	.784**	-.447*	.545**	.722**	.685**	.743**	.648**	.632**	1	.418*	0.195	.648**	0.280
Sr	.482**	0.349	0.034	.639**	-.696**	.617**	0.015	0.252	.396*	.495**	-0.021	0.268	.498**	.636**	.478**	0.311	.479**	.418*	1	.583**	.526**	0.145
Ba	.461*	0.357	-0.023	0.090	-0.281	0.153	0.079	0.078	0.258	0.254	-0.153	0.027	0.348	.527**	0.330	0.249	0.244	0.195	.583**	1	.445*	0.139
Pb	.863**	.685**	0.143	0.113	-.515**	.566**	0.325	0.254	.676**	.633**	-.438*	0.345	.779**	.778**	.917**	.782**	.760**	.648**	.526**	.445*	1	.232
Cd	0.259	0.154	-0.200	0.013	0.107	0.089	.642**	0.091	0.184	0.178	-0.060	0.031	0.214	0.146	0.251	0.097	0.213	0.280	0.145	0.139	0.232	1

**. At level 0.01, the correlation was significant. *. At level 0.05, the correlation was significant.

Supplemental Figures

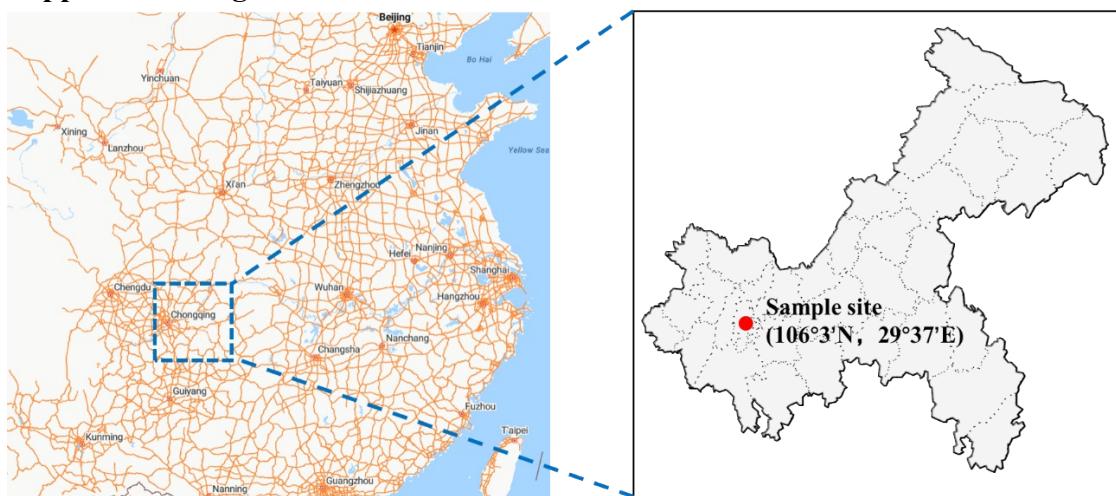


Figure S1. Geographical schematic of the sampling site ($106^{\circ}3'N$, $29^{\circ}37'E$) in Chongqing.

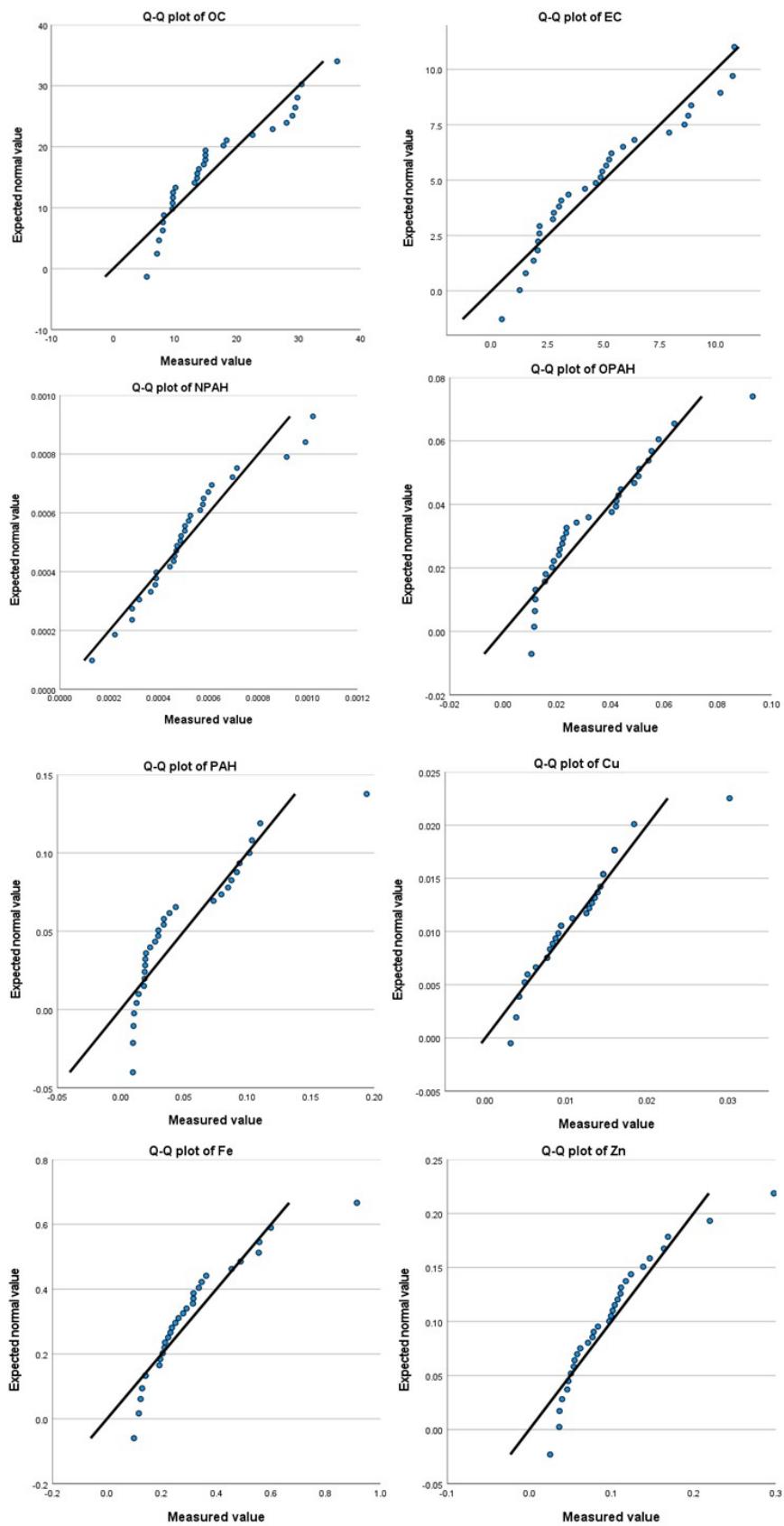


Figure S2. The residual plots (Q-Q plots) of the partial components of PM_{2.5} in Chongqing.

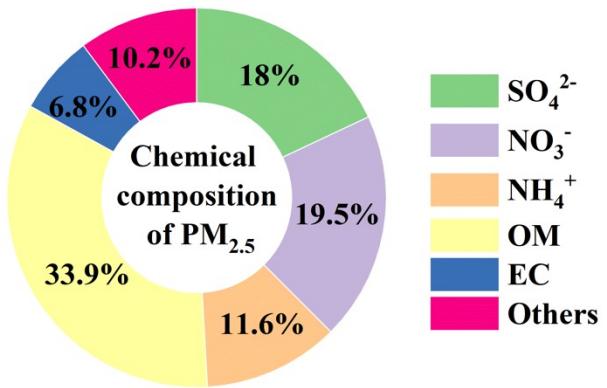


Figure S3. Chemical composition of PM_{2.5} in wintertime Chongqing.

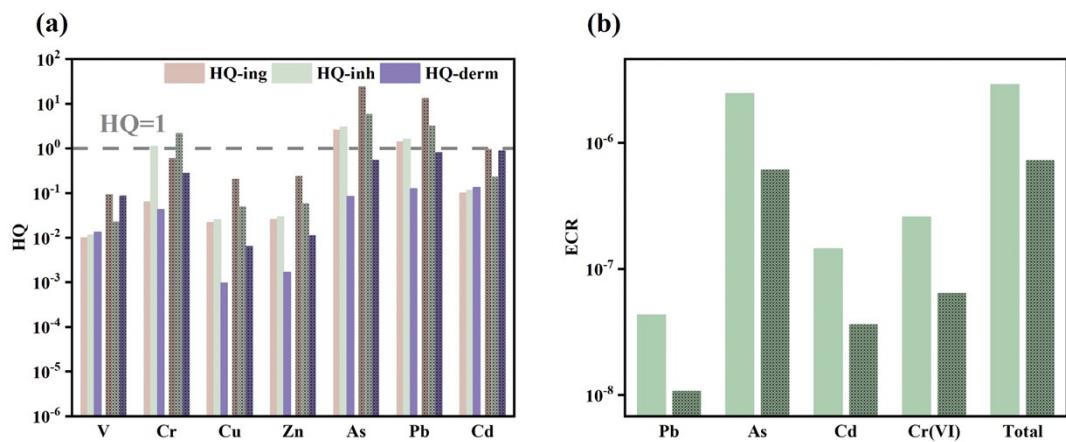


Figure S4. (a) HQ for adults (light color) and children (dark color) upon metal exposure (V, Cr, Cu, Zn, As, Pb, Cd) via ingestion (HQ-ing), inhalation (HQ-inh), and dermal (HQ-derm) exposure pathways. (b) ECRs of metals (Cr, As, Pb, and Cd) for adults (light color) and children (dark color) in PM_{2.5} collected during wintertime Chongqing.

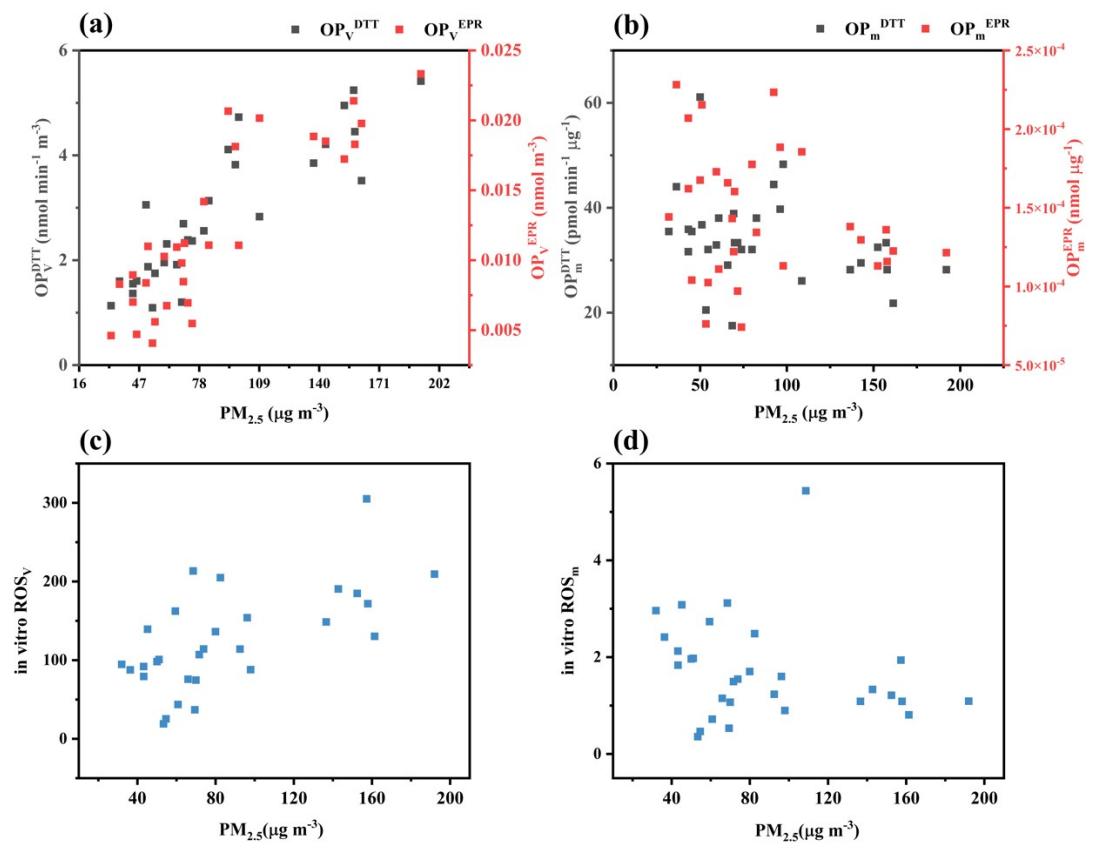


Figure S5. Correlation between (a) OP_V , (b) in vitro ROS_V , (c) OP_m and (d) in vitro ROS_m with $\text{PM}_{2.5}$ mass concentration, respectively.

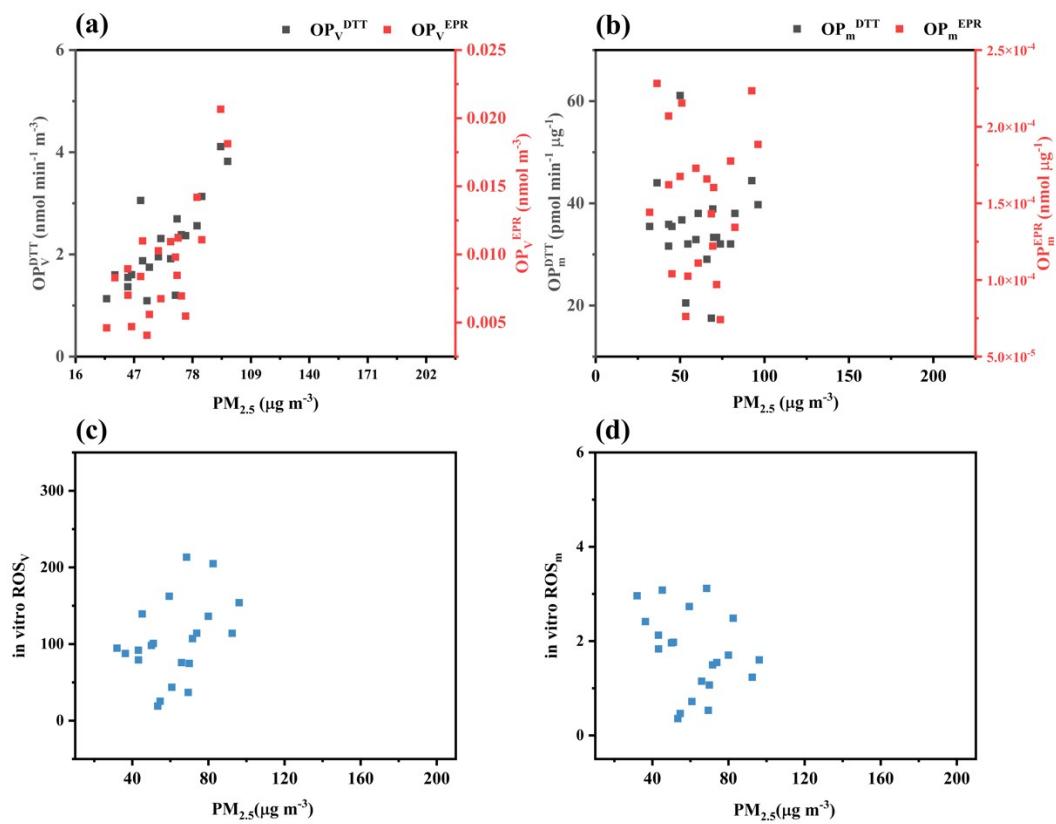


Figure S6. Correlation between (a) OP_V , (b) in vitro ROS_V , (c) OP_m and (d) in vitro ROS_m with $\text{PM}_{2.5}$ mass concentration in non-haze episodes, respectively.

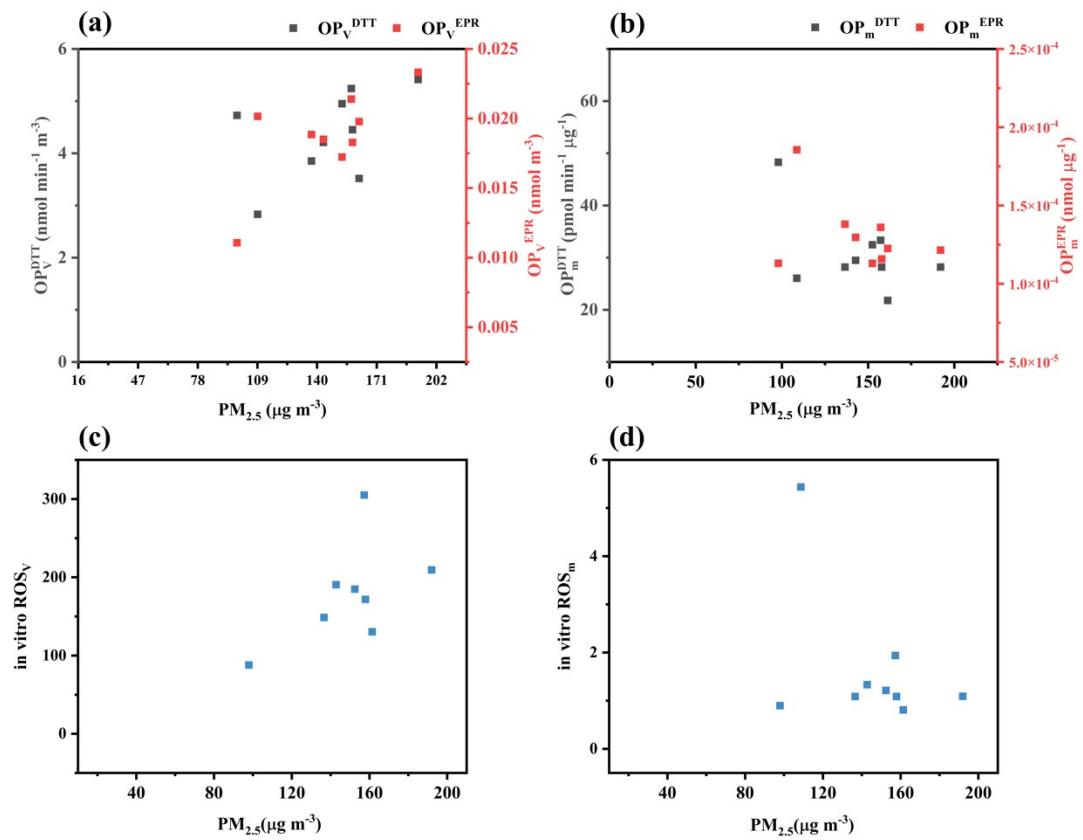


Figure S7. Correlation between (a) OP_v , (b) *in vitro* ROS_v , (c) OP_m and (d) *in vitro* ROS_m with $PM_{2.5}$ mass concentration in non-haze episodes, respectively.

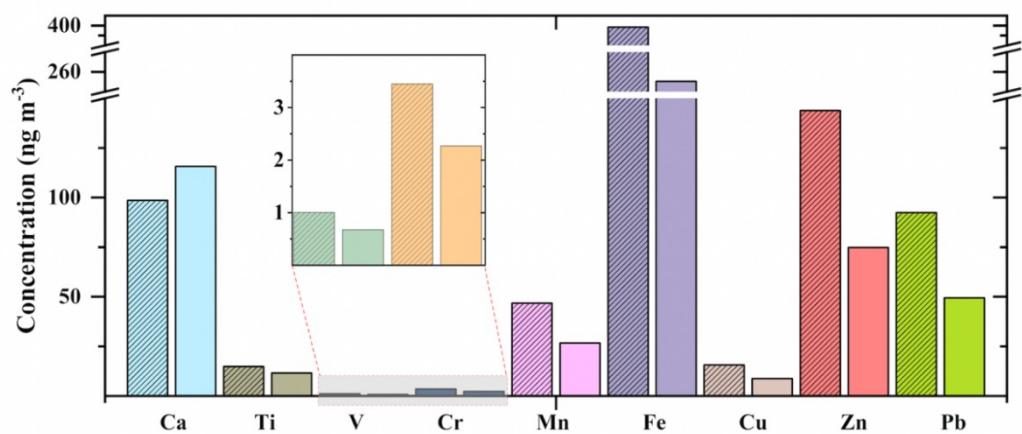


Figure S8. Comparison of metallic elements in $\text{PM}_{2.5}$ during haze (shadow) and non-haze period (unshadow) in Chongqing.

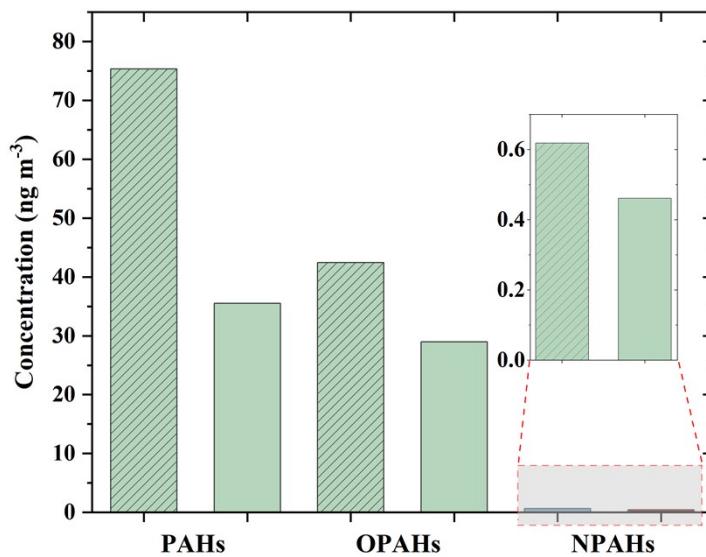


Figure S9. Comparison of PAHs, OPAHs, and NPAHs in $\text{PM}_{2.5}$ during haze (shadow) and non-haze period (unshadow) in Chongqing.

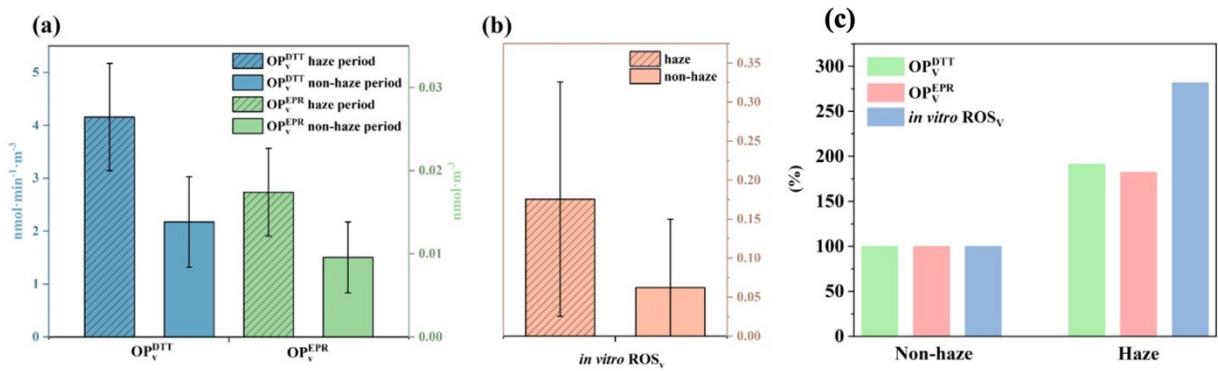


Figure S10. Comparison of (a) OP_V^{DTT} , OP_V^{EPR} and (b) $in vitro ROS_V$ of $\text{PM}_{2.5}$ during haze (shadow) and non-haze period (unshadow) in Chongqing. (c) OP_V^{DTT} , OP_V^{EPR} and $in vitro ROS_V$ during haze periods normalized to those measured during non-haze episodes.

289 **References**

- 290 1. J. C. Chow, J. G. Watson, L. W. A. Chen, W. P. Arnott, H. Moosmüller and K. Fung,
291 Equivalence of elemental carbon by thermal/optical reflectance and transmittance with
292 different temperature protocols, *Environ. Sci. Technol.*, 2004, **38**, 4414-4422.
- 293 2. J. C. Chow, J. G. Watson, J. Robles, X. Wang, L. W. A. Chen, D. L. Trimble, S. D. Kohl, R.
294 J. Tropp and K. K. Fung, Quality assurance and quality control for thermal/optical analysis
295 of aerosol samples for organic and elemental carbon, *Anal. Bioanal. Chem.*, 2011, **401**,
296 3141-3152.
- 297 3. J. C. Chow, J. G. Watson, L. W. A. Chen, M. C. O. Chang, N. F. Robinson, D. Trimble and
298 S. Kohl, The IMPROVE_A Temperature Protocol for Thermal/Optical Carbon Analysis:
299 Maintaining Consistency with a Long-Term Database, *J. Air Waste Manag. Assoc.*, 2007,
300 **57**, 1014-1023.
- 301 4. EPA, Risk Assessment Guidance for Superfund (RAGS), Volume I Human Health
302 Evaluation Manual (Part F, Supplemental Guidance for Inhalation Risk
303 Assessment).*Journal*, 2009.
- 304 5. N. Galindo, E. Yubero, J. F. Nicolás, M. Varea and J. Crespo, Characterization of metals in
305 PM₁ and PM₁₀ and health risk evaluation at an urban site in the western Mediterranean,
306 *Chemosphere*, 2018, **201**, 243-250.
- 307 6. H. Li, H. Wu, Q. g. Wang, M. Yang, F. Li, Y. Sun, X. Qian, J. Wang and C. Wang,
308 Chemical partitioning of fine particle-bound metals on haze-fog and non-haze-fog days in
309 Nanjing, China and its contribution to human health risks, *Atmos. Res.*, 2017, **183**, 142-
310 150.
- 311 7. F. Li, J. Yan, Y. Wei, J. Zeng, X. Wang, X. Chen, C. Zhang, W. Li, M. Chen and G. Lü,
312 PM_{2.5}-bound heavy metals from the major cities in China: Spatiotemporal distribution,
313 fuzzy exposure assessment and health risk management, *J. Cleaner Prod.*, 2021, **286**,
314 124967.
- 315 8. J. Lv, M. Li, J. Xie, Z. Di, L. Zhao and R. Liu, Seasonal variation and chemical speciation
316 analysis of PM_{2.5} heavy Metals in Taiyuan City, *Environ. Sci. Technol.*, 2016, **39**, 126-131.
- 317 9. P. Pandey, D. K. Patel, A. H. Khan, S. C. Barman, R. C. Murthy and G. C. Kisku, Temporal
318 distribution of fine particulates (PM_{2.5}, PM₁₀), potentially toxic metals, PAHs and Metal-
319 bound carcinogenic risk in the population of Lucknow City, India, *J. Environ. Sci. Health*
320 *A*, 2013, **48**, 730-745.
- 321 10. N. T. Hieu and B.-K. Lee, Characteristics of particulate matter and metals in the ambient air
322 from a residential area in the largest industrial city in Korea, *Atmos. Res.*, 2010, **98**, 526-
323 537.
- 324 11. Q. Liu, Z. Lu, Y. Xiong, F. Huang, J. Zhou and J. J. Schauer, Oxidative potential of ambient
325 PM2.5 in Wuhan and its comparisons with eight areas of China, *Science of The Total*

- 326 *Environment*, 2020, **701**, 134844.
- 327 12. J. Wang, X. Lin, L. Lu, Y. Wu, H. Zhang, Q. Lv, W. Liu, Y. Zhang and S. Zhuang,
328 Temporal variation of oxidative potential of water soluble components of ambient PM_{2.5}
329 measured by dithiothreitol (DTT) assay, *Sci. Total Environ.*, 2019, **649**, 969-978.
- 330 13. Y. Lyu, H. Guo, T. Cheng and X. Li, Particle Size Distributions of Oxidative Potential of
331 Lung-Deposited Particles: Assessing Contributions from Quinones and Water-Soluble
332 Metals, *Environ. Sci. Technol.*, 2018, **52**, 6592-6600.
- 333 14. J.-M. Li, S.-M. Zhao, S.-H. Xiao, X. Li, S.-P. Wu, J. Zhang and J. J. Schwab, Source
334 apportionment of water-soluble oxidative potential of PM_{2.5} in a port city of Xiamen,
335 Southeast China, *Atmos. Environ.*, 2023, **314**, 120122.
- 336 15. Y. Yu, P. Cheng, Y. Li, J. Gu, Y. Gong, B. Han, W. Yang, J. Sun, C. Wu, W. Song and M.
337 Li, The association of chemical composition particularly the heavy metals with the
338 oxidative potential of ambient PM_{2.5} in a megacity (Guangzhou) of southern China,
339 *Environ. Res.*, 2022, **213**, 113489.
- 340 16. W. Liu, Y. Xu, W. Liu, Q. Liu, S. Yu, Y. Liu, X. Wang and S. Tao, Oxidative potential of
341 ambient PM_{2.5} in the coastal cities of the Bohai Sea, northern China: Seasonal variation and
342 source apportionment, *Environ. Pollut.*, 2018, **236**, 514-528.
- 343 17. L. Zhang, X. Hu, S. Chen, Y. Chen and H.-Z. Lian, Characterization and source
344 apportionment of oxidative potential of ambient PM_{2.5} in Nanjing, a megacity of Eastern
345 China, *Environ. Pollut. Bioavailability*, 2023, **35**, 2175728.
- 346 18. F. Yang, C. Liu and H. Qian, Comparison of indoor and outdoor oxidative potential of
347 PM_{2.5}: pollution levels, temporal patterns, and key constituents, *Environ. Int.*, 2021, **155**,
348 106684.
- 349 19. R. Li, C. Yan, Q. Meng, Y. Yue, W. Jiang, L. Yang, Y. Zhu, L. Xue, S. Gao, W. Liu, T.
350 Chen and J. Meng, Key toxic components and sources affecting oxidative potential of
351 atmospheric particulate matter using interpretable machine learning: Insights from fog
352 episodes, *J. Hazard. Mater.*, 2024, **465**, 133175.
- 353 20. Q. Chen, M. Wang, Y. Wang, L. Zhang, Y. Li and Y. Han, Oxidative Potential of Water-
354 Soluble Matter Associated with Chromophoric Substances in PM_{2.5} over Xi'an, China,
355 *Environ. Sci. Technol.*, 2019, **53**, 8574-8584.
- 356 21. D. Chirizzi, D. Cesari, M. R. Guascito, A. Dinoi, L. Giotta, A. Donateo and D. Contini,
357 Influence of Saharan dust outbreaks and carbon content on oxidative potential of water-
358 soluble fractions of PM_{2.5} and PM₁₀, *Atmos. Environ.*, 2017, **163**, 1-8.
- 359 22. M. M. Shafer, J. D. C. Hemming, D. S. Antkiewicz and J. J. Schauer, Oxidative potential of
360 size-fractionated atmospheric aerosol in urban and rural sites across Europe, *Faraday
361 Discuss.*, 2016, **189**, 381-405.

- 362 23. S. Weichenthal, M. Shekarrizfard, A. Traub, R. Kulka, K. Al-Rijleh, S. Anowar, G. Evans
363 and M. Hatzopoulou, Within-City Spatial Variations in Multiple Measures of PM_{2.5}
364 Oxidative Potential in Toronto, Canada, *Environmental Science & Technology*, 2019, **53**,
365 2799-2810.
- 366 24. T. Fang, V. Verma, J. T. Bates, J. Abrams, M. Klein, M. J. Strickland, S. E. Sarnat, H. H.
367 Chang, J. A. Mulholland, P. E. Tolbert, A. G. Russell and R. J. Weber, Oxidative potential
368 of ambient water-soluble PM_{2.5} in the southeastern United States: contrasts in sources and
369 health associations between ascorbic acid (AA) and dithiothreitol (DTT) assays, *Atmos.*
370 *Chem. Phys.*, 2016, **16**, 3865-3879.
- 371 25. R. Qian, S. Zhang, C. Peng, L. Zhang, F. Yang, M. Tian, R. Huang, Q. Wang, Q. Chen, X.
372 Yao and Y. Chen, Characteristics and potential exposure risks of environmentally persistent
373 free radicals in PM_{2.5} in the three gorges reservoir area, Southwestern China, *Chemosphere*,
374 2020, **252**, 126425.
- 375 26. Q. Chen, M. Wang, Y. Wang, L. Zhang, J. Xue, H. Sun and Z. Mu, Rapid determination of
376 environmentally persistent free radicals (EPFRs) in atmospheric particles with a quartz
377 sheet-based approach using electron paramagnetic resonance (EPR) spectroscopy, *Atmos.*
378 *Environ.*, 2018, **184**, 140-145.
- 379 27. Q. Chen, H. Sun, Z. Mu, Y. Wang, Y. Li, L. Zhang, M. Wang and Z. Zhang, Characteristics
380 of environmentally persistent free radicals in PM_{2.5}: Concentrations, species and sources in
381 Xi'an, Northwestern China, *Environ. Pollut.*, 2019, **247**, 18-26.
- 382 28. Q. Chen, H. Sun, M. Wang, Z. Mu, Y. Wang, Y. Li, Y. Wang, L. Zhang and Z. Zhang,
383 Dominant Fraction of EPFRs from Nonsolvent-Extractable Organic Matter in Fine
384 Particulates over Xi'an, China, *Environ. Sci. Technol.*, 2018, **52**, 9646-9655.
- 385 29. L. Wang, W. Zhao, P. Luo, Q. He, W. Zhang, C. Dong and Y. Zhang, Environmentally
386 persistent free radicals in PM_{2.5} from a typical Chinese industrial city during COVID-19
387 lockdown: The unexpected contamination level variation, *J. Environ. Sci.*, 2024, **135**, 424-
388 432.
- 389 30. J. Ai, W. Qin, J. Chen, Y. Sun, Q. Yu, K. Xin, H. Huang, L. Zhang, M. Ahmad and X. Liu,
390 Pollution characteristics and light-driven evolution of environmentally persistent free
391 radicals in PM_{2.5} in two typical northern cities of China, *J. Hazard. Mater.*, 2023, **454**,
392 131466.
- 393 31. Q. Chen, M. Wang, H. Sun, X. Wang, Y. Wang, Y. Li, L. Zhang and Z. Mu, Enhanced
394 health risks from exposure to environmentally persistent free radicals and the oxidative
395 stress of PM_{2.5} from Asian dust storms in Erenhot, Zhangbei and Jinan, China, *Environ.*
396 *Int.*, 2018, **121**, 260-268.
- 397 32. X. Guo, N. Zhang, X. Hu, Y. Huang, Z. Ding, Y. Chen and H.-z. Lian, Characteristics and
398 potential inhalation exposure risks of PM_{2.5}-bound environmental persistent free radicals

- 399 in Nanjing, a mega-city in China, *Atmos. Environ.*, 2020, **224**, 117355.
- 400 33. A. Filippi, R. Sheu, T. Berkemeier, U. Pöschl, H. Tong and D. R. Gentner, Environmentally
401 persistent free radicals in indoor particulate matter, dust, and on surfaces, *Environ. Sci.:
402 Atmos.*, 2022, **2**, 128-136.
- 403 34. M. Ahmad, J. Chen, Q. Yu, M. Tariq Khan, S. Weqas Ali, A. Nawab, W. Phairuang and S.
404 Panyametheekul, Characteristics and Risk Assessment of Environmentally Persistent Free
405 Radicals (EPFRs) of PM_{2.5} in Lahore, Pakistan, *Int. J. Environ. Res. Public Health*, 2023,
406 **20**, 2384.
- 407