Supplementary Information (SI) for New Journal of Chemistry. This journal is © The Royal Society of Chemistry and the Centre National de la Recherche Scientifique 2025

1 Supplemental file for:

2 Adsorption of trace perfluorooctanoic acid on corn stover-based lignin amine by

- 3 synergy of electrostatic and hvdrophobic interactions
- 4 Baoyuan Shi^{a, b, c}, Jun Dong^{a, b, c, *}, Yunhao Li^{a, b, c}, Weihong Zhang^{a, b, c}, Yongxin Li^{a, b, c}
- 5 a Key Lab of Groundwater Resources and Environment Ministry of Education, Jilin University,
- 6 China
- 7 b Jilin Provincial Key Laboratory of Water Resources and Water Environment, Jilin University,
- 8 China
- 9 c National and Local Joint Engineering Laboratory for Petrochemical Contaminated Site Control
- 10 and Remediation Technology, Jilin University, China
- 11 Corresponding Author: Jun Dong. Email: dongjun@jlu.edu.cn
- 12

13 SUPPLEMENTARY NOTES

14 Supplementary Note 1 | The formulas of removal rate and adsorption capacity of15 PFOA by CSLA:

The removal rate of CSLA towards PFOA (R, %) and the mass of PFOA adsorbed by unit mass of CSLA at desired time (q_t, mg/g) and equilibration time (q_e, mg/g) were calculated according to the following equations (1–3):

19
$$P(\%) = \frac{C_0 - C_e}{C_0} \times 100\%$$
 (1)

20
$$q_t = \frac{(C_0 - C_t)V}{m}$$
(2)

21
$$q_e = \frac{(C_0 - C_e)V}{m}$$
(3)

Where C₀, C_t and C_e (mg/L) denote the initial, immediate and final PFOA concentrations, respectively; V (L) stands for the volume of the PFOA solution used, and m (mg) presents the weight of CSLA.

25

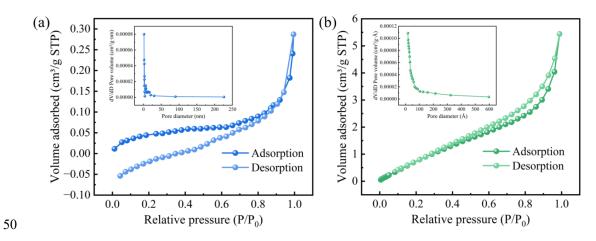
26 Supplementary Note 2 | Measurement and analysis:

This study used scanning electron microscopy (SEM, Hitachi Regulus 8100, 27 Japan) and energy dispersive X-Ray spectrometer (EDS, Ultim Max 65, UK) to 28 observe the surface morphology and element weight ratio of CSLA; the element 29 content analysis and valence state calibration of CSLA were performed by X-ray 30 photoelectron spectroscopy (XPS, Thermo fisher Nexsa G2, USA); fourier transform 31 infrared spectroscopy (FTIR, Thermo Fisher Scientific Nicolet iS5, USA) was used to 32 evaluate types of surface functional groups and chemical bonds of CSLA in KBr flake; 33 the surface potentials were determined by a Nano particle potential analyzer 34 (Zetasizer Nano S90, Malvern, UK); thermal stability of CSLA was measured using 35 thermal gravimetric analyzer (TGA, TA TGA 550, USA); the surface area and 36 porosity analyzer (BET, Micromeritics 3FLEX, USA) was used to gauge the specific 37 surface area of CSLA. 38

The concentration of PFOA in the solution was measured by High-Performance Liquid Chromatography Mass Spectrometry (U3000/TSQ quantum). The instrument test conditions of Liquid Chromatography are as follows: the mobile phase is 5 mmol/L ammonium acetate aqueous solution and acetonitrile (30:70, V/V); the injection volume is 1 μ L; the column temperature is 40°C; the flow rate is 0.15

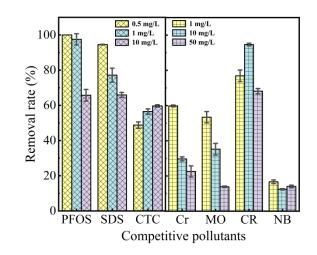
44 mL/min. The mass spectrometer used the anion mode HESI ion source. The scanning
45 conditions were: the ion source temperature was 350°C; the ion source voltage was
46 3000 V, the ion transport tube temperature was 350°C; the sheath gas pressure was 35
47 bar; the auxiliary gas pressure was 10 bar.

49 SUPPLEMENTARY FIGURES



51 Fig. S1. The N_2 adsorption-desorption isotherms of (a) lignin and (b) CSLA.

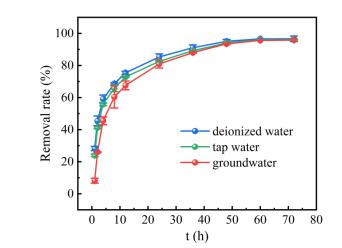
52



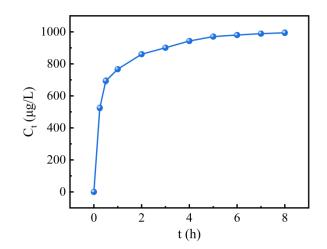
53

54 Fig. S2. Removal rate of competitive pollutants by CSLA

55

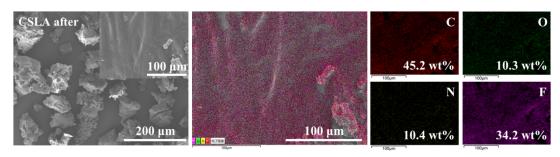


57 Fig. S3. The adsorption of PFOA on CSLA in actual water.

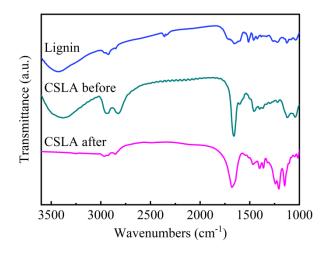


59 Fig. S4. Desorption efficiency experiment.

The react conditions of the adsorption test are: the initial concentration of PFOA 60 was 1000 µg/L, the volume of PFOA solution was 50 mL, and the addition amount of 61 CSLA was 5 mg. In order to obtain the desorption efficiency of PFOA by the mixed 62 solution of 1% NaCl and 70% methanol, the adsorbed and dried CSLA was added to 63 50 mL mixed solution for desorption. The concentration of PFOA in the mixed 64 solution was monitored during the desorption process, and it was found that the 65 desorption equilibrium was reached after 6 h. The concentration of PFOA in the 66 mixed solution at desorption equilibrium was 994.05 µg/L, and the desorption rate 67 reached 99.41%, indicating that the mixed solution of 1% NaCl and 70% methanol 68 desorption effect PFOA. 69 has а good on



- 70
- 71 Fig. S5. The SEM-EDS images of PFOA adsorbed CSLA.



74 Fig. S6. Complete FTIR spectra of lignin and CSLA before and after adsorption of

75 PFOA.

76 SUPPLEMENTARY TABLES

| Model Nonlinear equation | | Parameters | | | |
|--------------------------|--|---|--|--|--|
| Pseudo-first-order | $q_{t} = q_{e}(1 - e^{-k_{1}t})$ | q_e and $q_t (mg/g)$ are the adsorbed amount | | | |
| | | at an equilibrium concentration (Ce, | | | |
| | | mg/g) and a predetermined time (t, h) $k_{\rm l}$ | | | |
| | | (h ⁻¹) is the rate constants. | | | |
| Pseudo-second-order | $q_t = \frac{k_2 q_e^2 t}{1 + k_2 q_e^2 t}$ | $k_2 \; (g/(mg {\cdot} h))$ is the rate constants for the | | | |
| | | pseudo-second order. | | | |
| Elovich | $q_t = \frac{1}{\beta} \ln(1 + \alpha \beta t)$ | α (mg/(g·h)) is the initial rate constant, β | | | |
| | | (mg/g) is the desorption constant. | | | |
| Intra-particle diffusion | $\boldsymbol{q}_t = \boldsymbol{K}_t t^{1/2} + \boldsymbol{C}_t$ | $K_i~(mg/(g{\cdot}h^{0.5}))$ is the adsorption rate | | | |
| | | constants of intra-particle diffusion | | | |
| | | model and C_i is the constant for film | | | |
| | | thickness of the intra-particle diffusion | | | |
| | | model. | | | |

77 Table S1. Adsorption kinetic models.

| Model | Nonlinear equation | Parameters | | |
|------------|---|--|--|--|
| Langmuir | $q_{\rm max} K_L C_e$ | q _{max} (mg/g) is the maximum adsorption | | |
| | $q_e = rac{q_{\max} \mathcal{K}_L C_e}{1 + \mathcal{K}_L C_e}$ | capacity of the adsorbent, $q_e (mg/g)$ is the adsorption capacity at equilibrium, C | | |
| | | | | |
| | | (mg/L) is the adsorbate concentration | | |
| | | equilibrium, K_L (L/mg) is the constant | | |
| | | for the affinity between the adsorbate an | | |
| | | the adsorbent. | | |
| Freundlich | $q_e = K_F C_e^{1/n}$ | $K_F~(mg/g){\cdot}(L/mg)^{1/n}$ is the Freundlic | | |
| | 'e r e | constant, n is a dimensionless Freundlic | | |
| | | intensity parameter. | | |
| | | R is the universal gas constant (8.31 | | |
| | $\Delta G^{\circ} = -RT \ln \frac{q_e}{C_e}$ | J/mol·K), T is the absolute temperature | | |
| | $\Delta G^{0} = \Delta H^{0} - T \Delta S^{0}$ | (K), q _e (mg/g) is the adsorption capaci | | |
| | | at equilibrium, and C_e (mg/L) is the | | |
| | | adsorbate concentration at equilibriur | | |
| | | The values of ΔS^0 and ΔH^0 were obtained | | |
| | | from the intercept and slope of the lin | | |
| | | that was plotted by ΔG^0 versus T. | | |

79 Table S2. Adsorption isotherm models and adsorption thermodynamics equation.

| Parameters | PFOA solution (µg/L) | | | | |
|--|----------------------|--------|--------|--|--|
| | 250 | 500 | 1000 | | |
| Pseudo first-order | | | | | |
| $q_e (mg/g)$ | 4.74 | 9.49 | 19.33 | | |
| k_1 (h ⁻¹) | 0.87 | 0.63 | 0.16 | | |
| R ² | 0.9175 | 0.9648 | 0.9387 | | |
| Pseudo second-order | | | | | |
| q _e (mg/g) | 4.91 | 9.91 | 20.16 | | |
| $k_1 (h^{-1})$ | 0.32 | 0.10 | 0.01 | | |
| R ² | 0.9902 | 0.9764 | 0.9891 | | |
| Elovich | | | | | |
| $\alpha (mg/(g \cdot h))$ | 10191.80 | 873.96 | 56.68 | | |
| β | 2.93 | 1.11 | 0.39 | | |
| R ² | 0.8154 | 0.7870 | 0.9627 | | |
| Intra-particle diffusio | n | | | | |
| $K_1\left(mg/(g{\cdot}h^{1/2})\right)$ | 1.04 | 3.55 | 6.29 | | |
| $C_1 (mg/g)$ | 2.26 | 1.42 | -0.56 | | |
| R ² | 0.8976 | 0.9751 | 0.9862 | | |
| $K_2 (mg/(g \cdot h^{1/2}))$ | 0.23 | 0.43 | 1.34 | | |
| $C_2 (mg/g)$ | 3.81 | 7.67 | 10.04 | | |
| R ² | 0.9490 | 0.9905 | 0.9794 | | |
| $K_3 (mg/(g \cdot h^{1/2}))$ | 0.03 | 0.09 | 0.17 | | |
| C ₃ (mg/g) | 4.60 | 8.98 | 17.87 | | |
| R ² | 0.9652 | 0.9830 | 0.9945 | | |

81 Table S3. The parameters of adsorption kinetics models and intra-particle diffusion model.

| | Т | 283 K | 293 K | 303 K |
|----------------|--|--------|--------|---------|
| Langmuir | q _{max} (mg/g) | 426.87 | 654.11 | 1018.10 |
| | K _L (L/mg) | 0.07 | 0.10 | 0.22 |
| | R ² | 0.9970 | 0.9942 | 0.9911 |
| Freundlich | $K_F(mg/g)(L/mg)^{1/n}$ | 51.10 | 95.78 | 219.86 |
| | n | 1.93 | 1.99 | 2.02 |
| | R ² | 0.9941 | 0.9940 | 0.9881 |
| Thermodynamics | ΔG^0 (kJ/mol) | -10.56 | -15.08 | -18.12 |
| | ΔH^0 (kJ/mol) | | 96.22 | |
| | $\Delta S^0 \left(kJ/(mol \cdot K) \right)$ | | 0.38 | |

83 Table S4. The parameters of adsorption isotherm models and adsorption thermodynamics.

85 Table S5. Concentrations of contaminants in CSLA leaching solution.

| Samples | Mass concentration / $mg \cdot L^{-1}$ | Standard / mg·L ⁻¹ | | |
|--------------|--|-------------------------------|--|--|
| Cl- | 0.05224 | 50 ^a | | |
| Formaldehyde | 0.07903 | 0.9 ^b | | |

86 a Category I standard of *Standard for groundwater quality* (GB/T 14848-2017)

87 b Standard limits for specific projects of centralized drinking water surface water sources of

88 Environmental quality standards for surface water (GB 3838-2002)

| Material | C_0 | Т | pН | R | q_e^a | q_{max}^{b} | Reference |
|---|-------------|-----|-----|-------|---------|---------------|------------|
| | $(\mu g/L)$ | (K) | | (%) | (mg/g) | (mg/g) | |
| calcium-based MOF | 10 | - | 7.0 | 72.00 | 0.0038 | 0.061 | 1 |
| MWNTs | 100 | 297 | 6.5 | 65.00 | 0.00094 | 0.0035 | 2 |
| Alumina | 100 | 297 | 4.3 | - | 0.0049 | 0.014 | 3 |
| Carbon nanotube sponge | 100 | 297 | - | 75.00 | 1.52 | - | 4 |
| Boehmite | 200 | 297 | 7.0 | - | 0.027 | 0.19 | 5 |
| AE-APTMS@ gFe ₂ O ₃ | 1000 | 297 | - | 65.00 | 3.30 | 12.06 | 6 |
| MOF (In ₂ O ₃ -400) | 20000 | 297 | 4.6 | - | 25.45 | - | 7 |
| CS_MBC | 50000 | 297 | 3.5 | 63.42 | 158.55 | 517.00 | 8 |
| PEI-PVC NF | 100000 | 297 | 7.0 | 51.20 | 175.01 | 223.36 | 9 |
| | 250 | | | 99.03 | 4.95 | | |
| CSLA | 500 | 293 | 6.0 | 99.31 | 9.93 | 654.11 | This study |
| | 1000 | | | 98.45 | 19.69 | | |

90 Table S6. Comparison of PFOA adsorption on different adsorbents.

91 a The q_e is the saturated adsorption capacity obtained from the adsorption kinetics

92 experiment.

93 b The q_{max} is the maximum adsorption capacity of the material obtained by Langmuir

94 model fitting in the adsorption isotherm experiment.

95 References:

- 96 1 F. F. Sukatis, M. R. Razak, L. J. Looi, H. N. Lim, M. B. Abdul Rahman and A. Z.
- 97 Aris, *Microporous Mesoporous Mat.*, 2024, **380**, 113316.
- 98 2 X. N. Li, S. Chen, X. Quan and Y. B. Zhang, *Environ. Sci. Technol.*, 2011, 45,
 99 8498-8505.
- 100 3 F. Wang and K. Shih, *Water Res.*, 2011, **45**, 2925-2930.
- 101 4 A. Xue, Z. W. Yuan, Y. Sun, A. Y. Cao and H. Z. Zhao, Chemosphere, 2015, 141,
- 102 120-126.
- 103 5 F. Wang, C. Liu and K. Shih, *Chemosphere*, 2012, **89**, 1009-1014.
- 104 6 D. Y. Xing, Y. Chen, J. Zhu and T. Liu, Chemosphere, 2020, 251, 126384.
- 105 7 X. Gao, J. Chen, H. A. Che, Y. H. Ao and P. F. Wang, ACS ES&T Wat., 2022, 2,
- 106 1344-1352.
- 107 8 B. Saawarn, B. Mahanty and S. Hait, *Environ. Pollut.*, 2025, 368, 125734.
- 108 9 S. B. Kang, Z. Wang, W. Zhang, K.-Y. Kim and S. W. Won, Sep. Purif. Technol.,
- 109 2023, **326**, 124853.
- 110