

Supporting Information

Mn₃O₄@CoHCF nanocomposite for charge transfer regulated kinetic degradation of organophosphate pesticides

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Text S1. Information on chemicals, and solvents used in this study

All reagents and materials used in this study were of analytical grade and used as received without further purification. Ethion (ET 99.99%) and Chlorpyrifos (CP 99.99%) were supplied by Sigma-Aldrich (Merck, India). Sodium Hydroxide (98% purity), Manganese (II) Chloride Tetrahydrate 98% AR purchased from LOBA Chemie, India. Potassium Ferrocyanide [$K_4Fe(CN)_6$] 99% purity, Cobaltous Chloride Hexahydrate ($CoCl_2 \cdot 6H_2O$) 98% purity were purchased from RANKEM, India. For Scavenger analysis, tert-butyl hydroxide (t-BuOH, 98.9%), and chloroform ($CHCl_3$ 98.9%) and Ethylene diamine tetra acetic acid salt (EDTA) were brought from Rankem Chemical (Savita Scientific, Jaipur). Aloe vera leaves were collected locally from MNIT Campus, Jaipur.

Preparation details of plant extract:

Plant extract of Aloe vera was prepared by firstly removing the rind and outer layer from the leaf, and extract the gel from the leaves using various method like scraping. Deionized water was used throughout the experiment.

Role of Plant extract

In the present work, the use of *Aloe vera* extract in the synthesis of the $Mn_3O_4@CoHCF$ nanocomposite provides a dual functional role as both a reducing and stabilizing agent, owing to its rich phytochemical composition. The extract contains polyphenolic compounds, particularly anthraquinones (e.g., aloin and aloe-emodin) and flavonoids, which possess redox-active hydroxyl and carbonyl groups capable of donating electrons to facilitate the reduction of precursor metal ions during nanocomposite formation. Concurrently, naturally abundant polysaccharides (such as acemannan), along with proteins and saponins, contribute to effective surface capping and stabilization through hydrogen bonding and electrostatic interactions, thereby preventing nanoparticle agglomeration and controlling growth kinetics. Additionally, minor constituents such as ascorbic acid further enhance the reduction process. Collectively, these biomolecules enable a green, sustainable synthesis pathway, promoting the formation of a well-dispersed and structurally stable $Mn_3O_4@CoHCF$ heterostructure with improved

interfacial contact, which is advantageous for enhanced charge separation and photocatalytic performance.

Text S2: Instrumentation

The samples were subjected to PXRD analysis using CuK α radiation with $\lambda = 1.5406$ Å and a PAN analytical X-PRT PRO system from the USA. Samples that had been adequately ground and evenly distributed were placed in a coplanar form inside the sample container. A number of functional groups and bonds formed in the doped nanocomposite with full scan from 400-4000 cm^{-1} had been confirmed by FT-IR analyses conducted with an Agilent Cary 630 FTIR spectrometer (USA). This method doesn't require any special sample preparation or reference material, in contrast to normal FT-IR, which uses KBr as a reference. A FEI-Nova Nano SEM 450 microscope with a field emission gun was used to assess the surface morphology of the generated samples (Kensington UNSW Sydney, NSW 2052). It gives a resolution of 1.6 nm at 1 kv (TLD-SE) & <1 nm at 15 kv (TLD-SE). Electron dispersive spectroscopy was used to assess the element weight percentage in the sample. The EI ion source has been configured to operate at 350 degrees. An X-ray photoelectron spectroscopy (XPS) measurement was conducted using Oxford instrumentation. The Malvern Zetasizer (Version 7.11) was used to measure the zeta potential in order to evaluate the stability of the photocatalyst. The band gap has been determined by converting the Diffuse Reflectance Spectra data from a Shimadzu UV-Vis spectrometer to Tauc's plot. A UV Spectrophotometer (Agilent Pro) was used to measure the absorbance of the samples within the wavelength of 200–800 nm. The EIS study was conducted by CHI 7087E instrument with frequency range 0.01Hz to 105 Hz at the amplitude of 0.01 V. In addition, CPA-225D Sartius Analytical balance was used for weighing purpose and LABWAN –PH-61WW instruments was used for the preparation of pH solutions total organic carbon (TOC) analysis by the TOC -L analyser Shimazu.

Text S3: Band Gap Calculation

Band gap energy (E_g) was calculating using Tauc eq.

$$\alpha = \frac{C(h\nu - E_g)^n}{h\nu}$$

It is based on Tauc's relation which links the absorption coefficient to the photon energy that serves as the basis for it: $(\alpha h\nu)^2$ (eV cm^{-1})² vs Energy ($h\nu$).

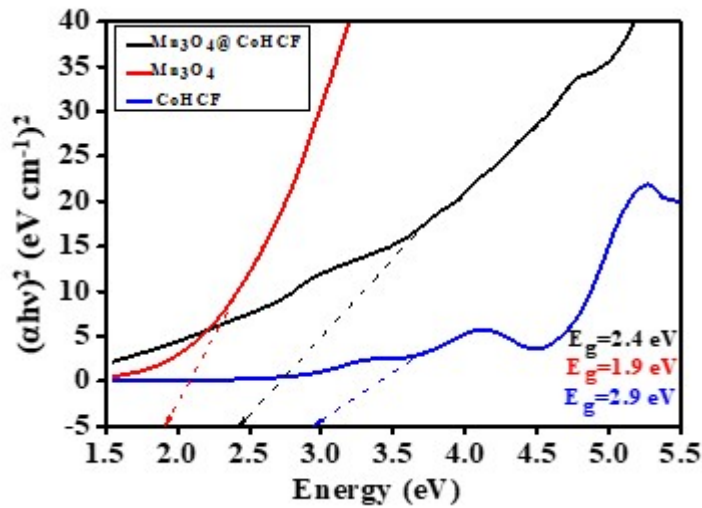


Fig.1S Band Gap analysis of $\text{Mn}_3\text{O}_4@\text{CoHCF}$, CoHCF and Mn_3O_4

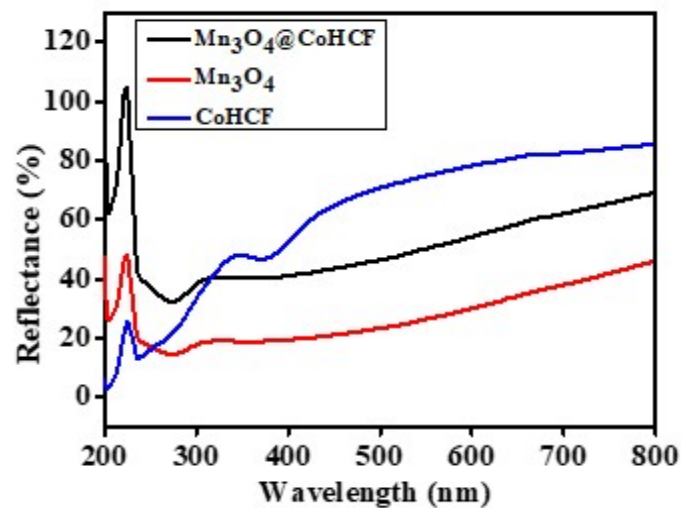


Fig.2S Reflectance spectra of $\text{Mn}_3\text{O}_4@\text{CoHCF}$, CoHCF and Mn_3O_4

Crystalline size can be calculated using the Scherrer equation i.e.

$$d = \frac{k\lambda}{\beta \cos\theta} \quad (\text{Eq1S})$$

where k is constant and its value is 0.9

λ is X-ray wavelength = 1.5 for Cu

β is the FWHM of the peak profile

θ is the diffraction angle of the reflection

The average crystallite size of $\text{Mn}_3\text{O}_4@\text{CoHCF}$, Mn_3O_4 and CoHCF is 16.89 nm, 8.39 nm and 11.18 nm, respectively.

Peak broadening and shifting indicated both the strain related to powder and the refinement of the crystal lattice. The following formula was used to determine the strain caused by crystal distortion and imperfection in powder:

$$\delta = 1/D^2 \quad (\text{Eq 2S})$$

$$\beta_{\text{hkl}} = \frac{K\lambda}{D\cos\theta} + 4\epsilon\tan\theta \quad (\text{Eq 3S})$$

Where, K = shape factor (0.9), ϵ = micro-strain, δ = dislocation density, and λ = wavelength of CuK radiation (0.154 nm). Plotting between $4 \sin\theta$ and $\beta_{\text{hkl}} \cos\theta$ provides crystalline size (intercept) and micro-strain (slope) from eq.2S. The dislocation density for the nanocomposite was calculated by eq. 3S.

Text S4: Statistical analysis

Sigma plot ver. 10.0 software (SPSS) was used to empirically analyze the reaction kinetics, rate constants, and best fit plots for eliminating Ethion and Chlorpyrifos from nanocatalyst surfaces. The standard deviation was computed using the Microsoft Excel program, which produces higher accuracy with triplicate samples. The results were then supported by Regression of coefficient (R^2) and p- value (≥ 0.05). For adsorption data analysis, several models were used (Langmuir isotherm, Freundlich, Dubinin-Radushkevich (DRK) and Temkin

isotherm). Linear curves ($Y=mX+c$) for adsorption and concentration-dependent study have been obtained while exponential graphs attaining maxima were observed for reduction.

The pollutant was removed by adsorption over a 60-minute period during the kinetic examination of negative time values. Therefore, the total removal of pesticides consisted of the contributions from both the adsorption and the photodegradation. The details of the adsorption experiment are given below-

The 20 mg of each adsorbent ($Mn_3O_4@CoHCF$, Mn_3O_4 , and $CoHCF$) was brought into contact with aq. Pesticides solution (50 mg L^{-1}) at 7 pH and reaction mixture was agitated at a speed of 300 rpm for 0-60 min under dark conditions. The percentage of pesticides in each of the aliquot samples was determined after 15, 30, 45, and 60 min by measuring the difference in absorbance using HPLC (Model: 1260WD).

Details of Langmuir, Freundlich, Temkin, Sip and D-R isotherms used in present study

The absorbance (X_e) was calculated through the equation (Figure 4S):

$$X_e = \frac{(C_i - C_e) \times \text{Vol of solution (mL)} \times \text{Molecular weight of adsorbent}}{\text{Amount of catalyst (mg)}}$$

Where C_i and C_e are the initial and the equilibrium concentrations of pesticides in mg L^{-1} ; X_e is the amount of EDCs adsorbed per gm weight of adsorbent (photocatalyst).

Adsorption data through Langmuir adsorption isotherms were calculated through a graph of C_e/X_e v/s C_e of the solute. These were calculated by fitting the adsorption data into the equation:

$$\frac{C_e}{X_e} = \frac{1}{k_L X_m} + \frac{C_e}{X_m}$$

or

$$\frac{1}{X_e} = \frac{1}{C_e} \left(\frac{1}{k_L X_m} \right) + \frac{1}{X_m}$$

Freundlich Isotherm.

A typical graph of X_e v/s C_e of the solute was a straight line.

$$X_e = K_f + C_e^{1/n} \text{ and the linear form is } \log X_e = \log K_f + \frac{1}{n} \log C_e$$

Where C_e is the equilibrium concentration of the VOCs solution; X_e is the number of VOCs adsorbed per gram weight of adsorbent; K_f is the Freundlich adsorption constant (mg g^{-1}); n is adsorption intensity.

Temkin Isotherm:

$$X_e = \frac{RT}{b} \ln A + \frac{RT}{b} \ln C_e$$

Where $B=RT/b$ constant related to heat of sorption (J/mol) obtained from the Temkin plot (q_e Vs $\ln C_e$); A (slope) = Temkin isotherm equilibrium binding constant (L g^{-1}); b (intercept) = Temkin isotherm constant; R = universal gas constant ($8.314 \text{ J mol}^{-1} \text{ K}^{-1}$) T = Temperature at 298, 308 and 318 K.

As C_e values were very low, the $\ln C_e$ values were coming out to be negative. Therefore, Temkin isotherm (X_e v/s $\ln C_e$) was not plotted for the present study.

Dubinin-Radushkevich (D-R) isotherm:

$$\ln X_e = \ln X_m - \beta \varepsilon^2$$

$$\varepsilon = RT \ln \left(1 + \frac{1}{C_e} \right)$$

Where X_m is the maximum adsorption capacity (mg g^{-1}) obtained from intercept; β ($\text{mol}^2 \text{J}^{-2}$) is an activity coefficient constant related to adsorption energy and obtained from slope; ε is Polanyi potential. D-R was plotted between $\ln X_e$ Vs ε^2 .

Sips Isotherm

It is plotted between $1/X_e \times 10^{-2}$ (g mg^{-1}) and $(1/C_e) \times 10^{-8}$ L mg^{-1} ; where $(1/\text{mg})$ and (mg/g) are the Sips equilibrium constant and maximum adsorption capacity values obtained from the slope and the intercept of the plot. The Sips isotherm equation is characterized by the dimensionless

heterogeneity factor 'n' which can also be employed to describe the system's heterogeneity when is between 0 and 1.

$$\frac{1}{X_e} = \frac{1}{X_m K_s} \left(\frac{1}{C_e} \right)^{1/n} + \frac{1}{X_m}$$

Therefore, different adsorption isotherms (Langmuir, Sips, Freundlich, Dubinin-Radushkevich, and Temkin) were investigated. Pesticides eradication was found to be initial adsorption which best fitted the Langmuir adsorption model. The photocatalytic process roughly followed first-order kinetics, which can be expressed as follows: $(C_t/C_0) = \exp^{-kt}$, where $|k|$ is the apparent rate constant and C_t and C_0 are the final and initial concentrations of VOCs, respectively. Both rate constant and rate suggest that pesticides removal was much faster in sunlight than dark i.e., photodegradation >>> adsorption. Additionally, nanocomposite shows much more photocatalytic activity compared to individual nanoparticles i.e., $Mn_3O_4@CoHCF > CoHCF > Mn_3O_4$.

Text S5: Turn over frequency (TOF)

As we know that TOF of catalysts determines their efficiency, i.e. higher the TOF more efficient is the catalyst. TOF value can be calculated with the help of following formula:

$$TOF = \frac{\text{No of moles of reactant/No of grams of photocatalysts}}{\text{Time (min)}} \times \text{yield}$$

No of gram of photocatalyst: 0.02 g. (2ppm)

Table 1S: TOF of $Mn_3O_4@CoHCF > CoHCF > Mn_3O_4$ for ET and CP pollutants

S.No.	Photocatalyst	TOF (mol min ⁻¹ g ⁻¹)	
		ET	CP
1.	Mn ₃ O ₄ @CoHCF	1.28×10 ⁻³	6.26×10 ⁻⁴
2.	CoHCF	1.16×10 ⁻³	5.60×10 ⁻⁴
3.	Mn ₃ O ₄	1.04×10 ⁻³	5.06×10 ⁻⁴

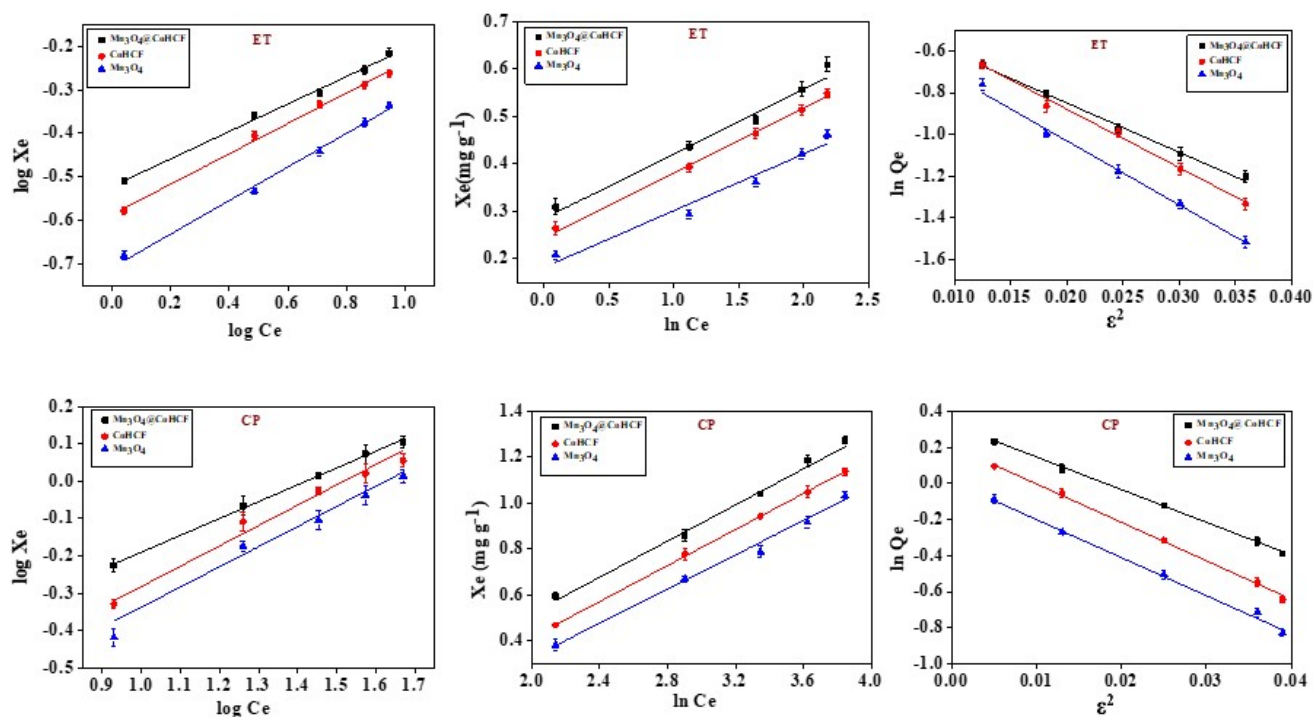


Figure 3S: (a) Freundlich (b) Temkin and (c) D-R isotherms for ET and CP adsorption. Note: Triplicate experiments (n=3) were evaluated for estimation of error bar.

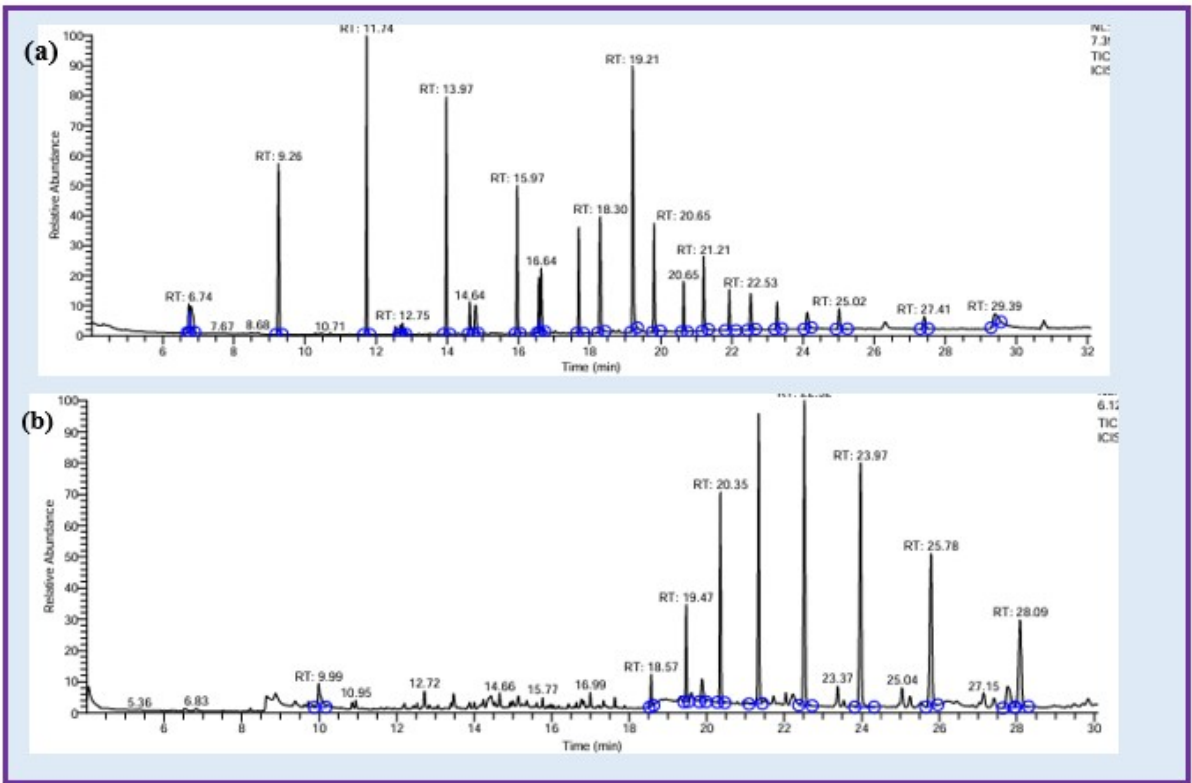
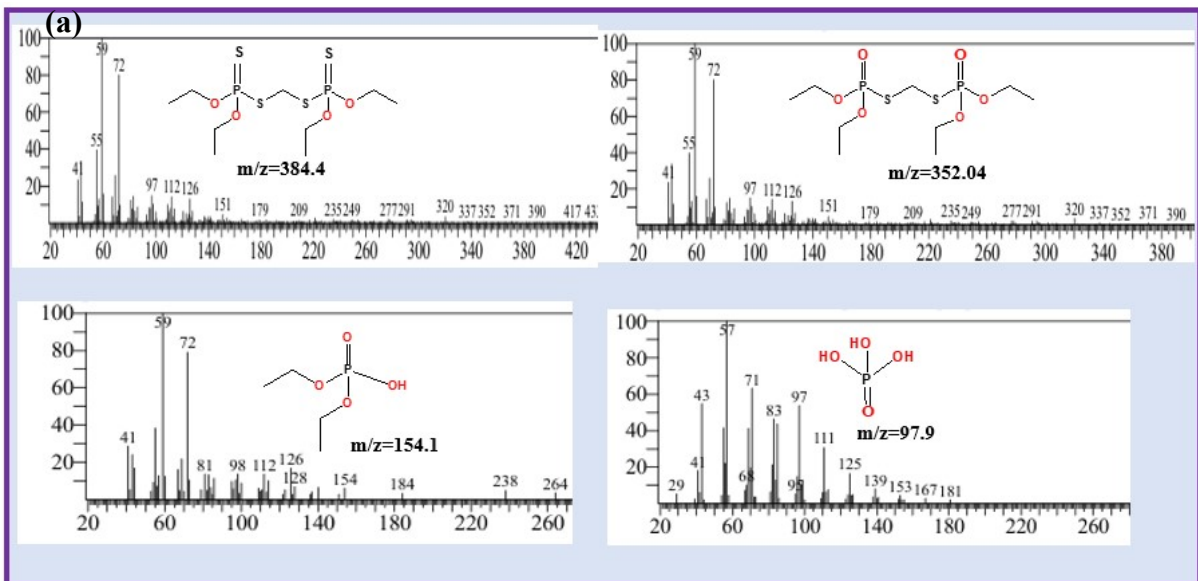


Figure 4S: Representative of Total ion chromatogram (TIC) of (a) Ethion and (b) Chlorpyrifos



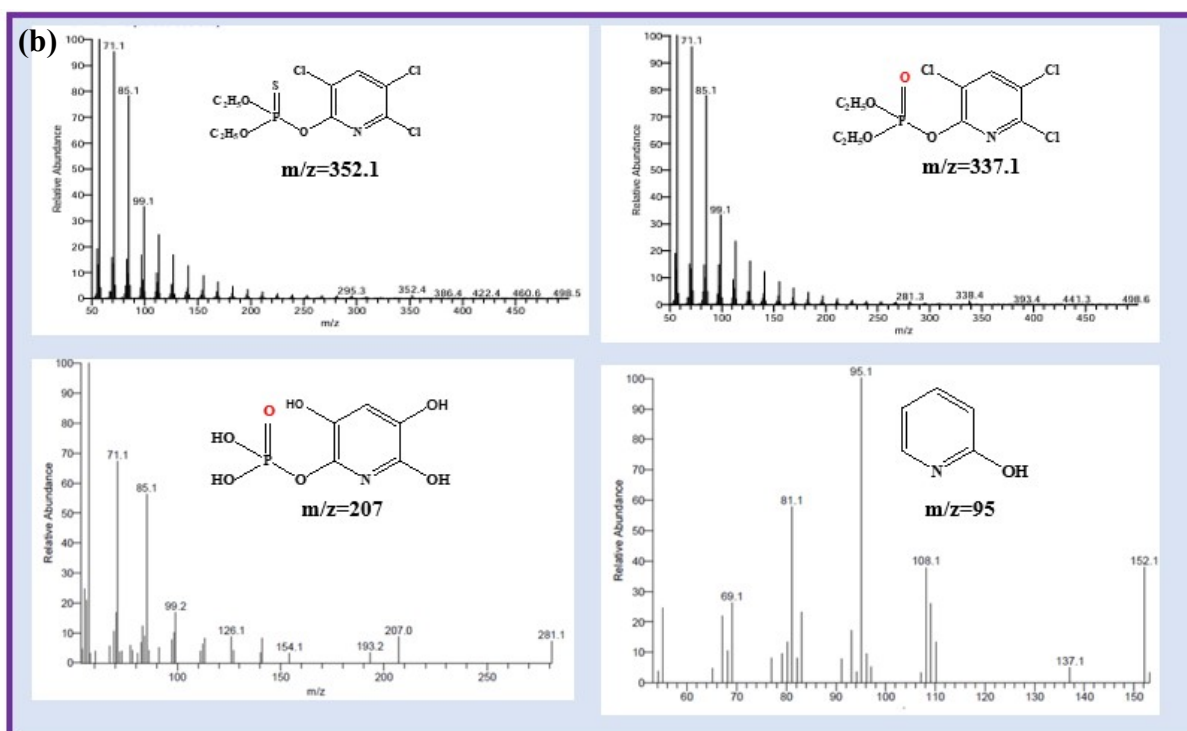


Figure 5S: mass spectrum of the degraded products formed of (a)Ethion and (b) Chlorpyrifos over $Mn_3O_4@CoHCF$ nanocatalysts

Table 2S. Details of angle, FWHM values, interplanar spacing and hkl values of respectively $Mn_3O_4@CoHCF$, Mn_3O_4 and $CoHCF$ respectively.

a) $Mn_3O_4@CoHCF$

Angle (2θ)	FWHM Value	d-spacing (\AA)	hkl values
19.24°	0.4684	4.59 \AA	(101)
29.21°	0.4015	3.06 \AA	(112)
32.53°	0.4015	2.74 \AA	(103)
36.75°	0.1004	2.47 \AA	(211)
38.16°	0.5353	2.35 \AA	(004)
51.18°	0.8029	1.78 \AA	(440)
60.14°	0.5348	1.53 \AA	(224)
64.70°	0.8029	1.43 \AA	(314)

Mn ₃ O ₄ @CoHCF	0.9989	0.003	0.9985	0.04	0.9736	0.001	0.9634	0.001
CoHCF	0.9939	0.02	0.9989	0.02	0.9972	0.001	0.9928	0.003
Mn ₃ O ₄	0.9945	0.01	0.9951	0.01	0.9663	0.002	0.9973	0.002

Catalyst	DRK				Freundlich			
	ET		CP		ET		CP	
	R ²	p	R ²	p	R ²	P	R ²	p
Mn ₃ O ₄ @CoHCF	0.9968	0.01	0.9986	0.16	0.9968	0.001	0.9981	0.001
CoHCF	0.9951	0.001	0.9971	0.23	0.9964	0.001	0.9885	0.002
Mn ₃ O ₄	0.9936	0.001	0.9957	0.61	0.9943	0.002	0.9687	0.002

Table 4S: Reduced chi square and Adjusted R square values of kinetics model for ET and CP

Catalyst	Mn ₃ O ₄ @CoHCF			CoHCF			Mn ₃ O ₄			Blank		
	Adj. R ²	Red. Chi ²	DO F	Adj. R ²	Red. Chi ²	DO F	Adj. R ²	Red. Chi ²	DO F	Adj. R ²	Red. Chi ²	DOF
ET	0.98016	0.00208	3	0.99643	0.00208	3	0.99533	0.0014	3	0.94004	5.8* 10 ⁻⁴	3
CP	0.97573	6.962	3	0.96736	2.291	3	0.99323	1.107	3	0.98324	0.075	3

Table 5S: Reduced chi square and Adj. R square values of kinetics model for ET and CP pollutant.

Catalyst	Langmuir						Temkin					
	ET			CP			ET			CP		
	Adj. R ²	Red. Chi ²	DO F	Adj. R ²	Red. Chi ²	DO F	Adj. R ²	Red. Chi ²	DO F	Adj. R ²	Red. Chi ²	DOF
Mn₃O₄@CoHCF	0.9989	0.07312	3	0.99801	0.25792	3	0.9649	1.92433	3	0.9649	1.924	3
CoHCF	0.9907	0.06589	3	0.99857	0.19169	3	0.99631	0.31188	3	0.9963	0.31188	3
Mn₃O₄	0.9926	0.06308	3	0.99353	1.12844	3	0.95517	5.16317	3	0.95517	5.16317	3

Catalyst	DRK						Freundlich					
	ET			CP			ET			CP		
	Adj. R ²	Red. Chi ²	DO F	Adj. R ²	Red. Chi ²	DO F	Adj. R ²	Red. Chi ²	DOF	Adj. R ²	Red. Chi ²	DOF
Mn₃O₄@CoHCF	0.99584	0.38996	3	0.99848	0.3238	3	0.9649	1.924	3	0.997948	0.15618	3

CoHCF	0.993	0.7715	3	0.997	1.193	3	0.99	0.311	3	0.98476	2.1225	3
	56	4		1	5		631	8				
Mn₃O₄	0.991	1.0549	3	0.995	1.029	3	0.95	5.162	3	0.95835	2.4333	3
	54	8		7	9		51					

Table 6S: Comparison table of different photocatalysts

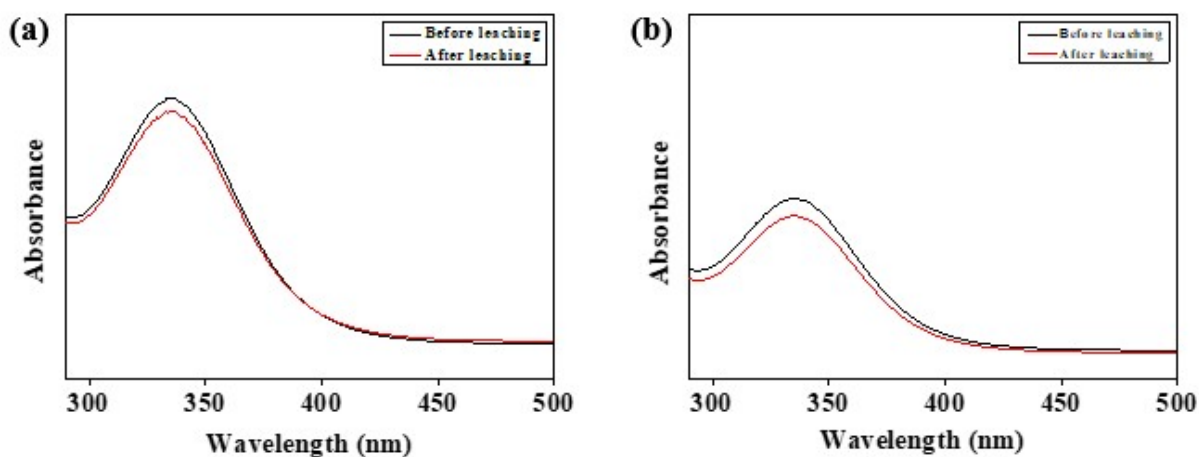


Fig.6S Leaching study in (a) 1 N HCl (b) 0.5 N HCl of Mn₃O₄@CoHCF nanocomposite

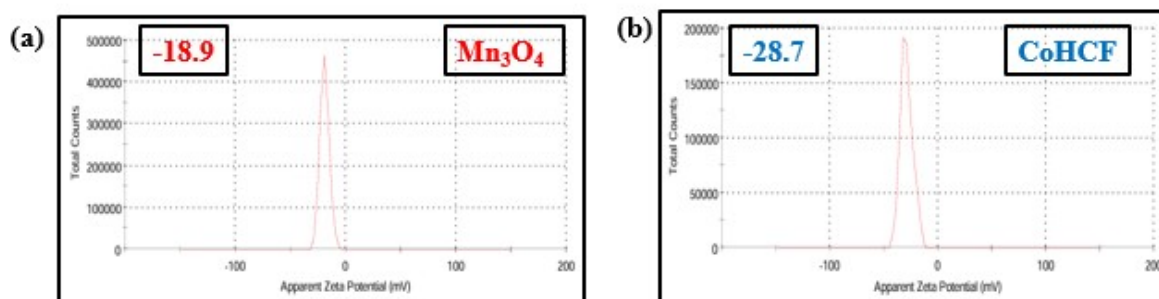


Fig.7S Zeta potential of (a) Mn₃O₄ and (b) CoHCF nanocomposite

References of Table 6S:

Catalyst	Pollutants	Characterization Technique	Mechanism and Degradation %	Reference
Zn-Mn ₃ O ₄ /TiO ₂	Methylene Blue	P-XRD, FTIR, TGA DSC, BET, HRTEM, GC-MS,	Sol-gel method, Photodegradation-96%	[1]
Prussian blue/Mn ₃ O ₄	Levofloxacin	P-XRD, FTIR, FE-SEM, EIS, PL	Z-scheme p-n heterojunction, Photodegradation-96%	[2]

Mn ₃ O ₄ @g-C ₃ N ₄	Erythrosine B (EB) Indigo Carmine (IC)	P-XRD, FTIR, Raman, XPS, FE-SEM	Follow pseudo-2nd-order kinetics Photodegradation- 92.7%	[3]
Fe ₃ S ₄ /Mn ₃ O ₄	Methylene Blue (MB) and Trichlorophenol (TCP)	P-XRD, FTIR, FE-SEM, XPS, BET, TGA	Hydrothermal process, Photodegradation-97.5%	[4]
CoFe ₂ O ₄ /Mn ₃ O ₄	Methylene Blue	P-XRD, FE-SEM, XPS, FTIR, Raman	Sol-gel-cum- hydrothermal method, LEDdriven photocatalysis -92%	[5]
Mn ₃ O ₄ /ZnO	Tetracycline	P-XRD, FTIR, FE-SEM, BET, XPS	Ultrasonification, Photodegradation- 97.5 %	[6]
ZrTiO ₄ /TiO ₂ / Mn ₃ O ₄ (ZTM)	Metoclopramide (MCP)	P-XRD, FTIR, UV-Vis DRS, EIS, TG/DTG, SEM, TEM	Z-scheme charge transfer, Photodegradation- 97.5 %	[7]
Al ₂ O ₃ /Mn ₃ O ₄ / Fe ₂ O ₃	Methylene Blue	P-XRD, FTIR, SEM-EDS	Co-precipitation method, UV-radiation 98.2%	[8]
NiO/Mn ₃ O ₄	Thiamethoxam, crystal violet, rhodamine B	P-XRD, FTIR, FE-SEM, EDS, XPS, HR-TEM	Hydrothermal method, Photocatalytic degradation -93% ,93.6%, 93.2%	[9]
Ag/Mn ₃ O ₄	Congo red	P-XRD, XRF, Raman, BET, SEM, TEM	Sol-gel and hydrothermal methods, Sonophotocatalytic degradation	[10]
Mn ₃ O ₄ @CoHCF	Ethion (ET) and Chlorpyrifos (CP)	P-XRD, FTIR, FE-SEM, XPS, BET, ESR	Photocatalytic degradation	Present Work

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